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UNIVERSITY OF ALBERTA

THE WETLANDS OF ELK ISLAND NATIONAL PARK: VEGETATION,
DEVELOPMENT, AND CHEMISTRY

BY



BARBARA NICHOLSON

A thesis submitted to the Faculty of Graduate Studies and
Research in partial fulfillment of the requirements for the
degree of DOCTOR OF PHILOSOPHY

IN

PEATLAND ECOLOGY

DEPARTMENT OF BOTANY

EDMONTON, ALBERTA

Spring 1993



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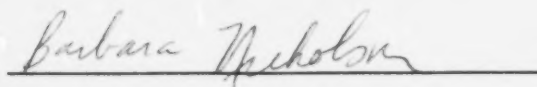
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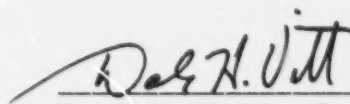
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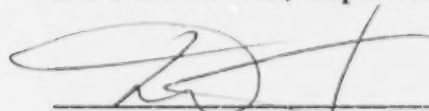
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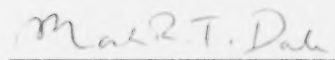
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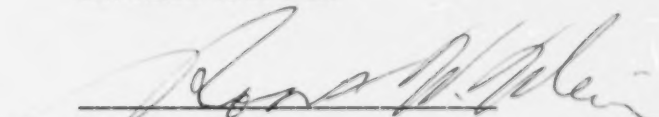
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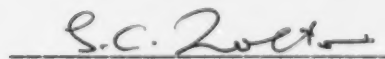
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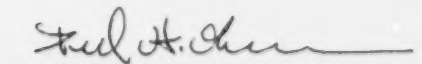

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To all who strive and aspire.

To my son Trevor, in hopes that someday he will reach beyond himself to achieve a higher goal and accomplish his life's greatest ambition.

To Dale, the man whom has taught, guided, listen, shared and cared. Thanks for showing me how, and generally devoting nine years of your life to see the successful completion of this thesis.

To Scott, for giving me the incentive and ambition to begin and finish this project.

Abstract

A vegetation survey of the wetlands of Elk Island National Park (EINP) resulted in the recognition of thirteen TWINSPAN vegetation groups ranging from saline marshes to forested bogs. Vegetation is predominantly controlled by water flow and cation gradients. Nutrient (nitrogen, phosphorus) content of the surface waters is not as important in determining species composition. The surface water of marshes, swamps, and moderate-rich fens are circumneutral in pH and high in calcium, magnesium, and sodium. Poor fens and bogs are acidic, with lower mineral contents. Elevation, number of inlet and outlet streams, water level, and hummock height all contribute to the diversity and composition of species in these wetlands. All wetlands at EINP have formed in small shallow basins. Two successional pathways of peatland evolution were identified from peat profiles; Aquatic - Rich fen - Poor fen - Bog and Aquatic - Marsh - Swamp. Autogenic factors drive the successional pathway. Autogenic changes evident in the developmental profiles are: 1) elevation of the peat surface from the local water table 2) acidification of the environment, 3) oligotrophism. Edaphic factors, not climate, control the rate and direction of successional changes. Severe water table fluctuations prevent the establishment of *Sphagnum* and the development of bog vegetation. Climate has affected peatland development of the park by 1) delaying peat initiation until after about 6700 yBP, 2) expanding the peat basins during a wetter period following 6720 yBP, 3) limiting peat formation to small terrestrialized basins and 4) delaying peatland succession compared to more northerly peatlands. Physical and chemical parameters measured in the peat profiles indicate that minerals were abundant during early aquatic stages. Oligotrophism and elevation of the peat surface is a process which is occurring in all wetland types. Results of the discriminant analysis on macrofossil peat types indicates that it is possible to identify peat types based upon chemical and physical characteristics alone. Sediments formed under aquatic conditions have high ash and bulk density values ($> 20\%$, $> 0.15 \text{ g/cm}^3$), and high elemental contents. Marsh and swamp peats contain less than 20% ash and less than 0.15

g/cm³ bulk density values, but greater than 1×10^4 mg/Kg calcium. Fen peats contain less than 12% ash and 1×10^4 mg/Kg of calcium, while bog peats contain 6% ash and less than 5×10^3 mg/Kg calcium.

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TABLE OF CONTENTS

	Page
INTRODUCTION.....	1
Wetland Terminology and Classification.....	3
A. OBJECTIVES.....	8
B. STUDY AREA.....	9
Climate.....	10
Vegetation.....	10
C. METHODS.....	12
Vegetation Inventory.....	12
Water Chemistry.....	14
Wetland Development Analysis.....	15
Peat Chemistry.....	18
RESULTS.....	19
Vegetation Classification.....	19
Correlation of Vegetation to Water Chemistry and Environmental Factors.....	28
Evolutionary Development.....	32
Peat Chemistry.....	55
Peat Profiles.....	66
Physical Profiles.....	67
Chemical Profiles.....	74
DISCUSSION.....	81
Classification of Wetland Sites.....	81
Correlation of Vegetation to Water Chemistry and Environmental Factors.....	87
Peatland Development.....	93
Paleoenvironmental Reconstructions.....	99
Comparisons to Published Literature.....	105
Peat Chemistry Comparison to Published Literature.....	109
Peat Profiles.....	112
Physical Profiles.....	112
Chemical Profiles.....	116
CONCLUSIONS.....	119
LITERATURE CITED.....	126
APPENDIX I.....	144

APPENDIX II.....	159
APPENDIX III.....	160

LIST OF TABLES

Table	Page
1. TWINSpan Vegetational Groups.....	20-23
2. Relationship to Wetland Type to Environmental Factors.....	29
3. Radiocarbon Dates.....	33
4. Canonical Discriminant Analysis Multivariate Statistics and F Approximations.....	58
5. Canonical Discriminant Analysis, Total-Sample Standardized Canonical Coefficients.....	60
6. Canonical Discriminant Analysis, Peat Classification.....	61
7. Peat Chemistry Variables, Means and Standard Deviations.....	63
8. Wetland Surface Water Chemistry Data, Published Reports.....	88
9. Published Peat Stratigraphic Profiles.....	106
10. Peat Chemistry, Published Reports.....	110-111

LIST OF FIGURES

Figure	Page
1. Location of 63 Wetlands used in the Vegetation Stand Inventory.....	13
2. Location of Six Peatlands used for Evolutionary Analysis.....	16
3. Macrofossil Analysis Fen Site 2-1.....	34
4. Macrofossil Analysis Fen Site 2-2.....	36
5. Macrofossil Analysis Fen Site 2-3.....	38
6. Macrofossil Analysis Fen Site 3-1.....	39
7. Macrofossil Analysis Swamp Site 3-3.....	41
8. Macrofossil Analysis Bog Site 22-1.....	43
9. Macrofossil Analysis Swamp Site 22-2.....	44
10. Macrofossil Analysis Marsh Site 22-3.....	45
11. Macrofossil Analysis Bog Site 27-1.....	47
12. Macrofossil Analysis Bog Site 27-4.....	49
13. Macrofossil Analysis Marsh Site 35-1.....	50
14. Macrofossil Analysis Marsh Site 35-2.....	52
15. Macrofossil Analysis Marsh Site 35-3.....	53
16. Macrofossil Analysis Swamp Site 48-1.....	54
17. Macrofossil Analysis Swamp Site 48-3.....	56
18. Canonical Graph of Peat Type Class Means.....	59
19. Percentage Ash Profiles-Fens.....	68
20. Percentage Ash Profiles-Bogs.....	69
21. Percentage Ash Profiles-Marsh/Swamps.....	70
22. Bulk Density Profiles-Fens.....	71
23. Bulk Density Profiles-Bogs.....	72
24. Bulk Density Profiles-Marsh/Swamps.....	73
25. Calcium Profiles-Fens.....	75
26. Calcium Profiles-Bogs.....	76
27. Calcium Profiles-Marsh/Swamps.....	78
28. Copper Profiles-Fens.....	79
29. Copper Profiles-Bogs.....	80
30. Copper Profiles-Marsh/Swamps.....	82
31. Peatland Development Profiles, Fens and Bogs.....	94

32.	Peatland Development Profiles, Marshes.....	98
33.	Paleoenvironmental Reconstructions: Height Above the Water Table - Fen Profiles.....	100
34.	Paleoenvironmental Reconstructions: Height Above the Water Table - Bog Profiles.....	102
35.	Paleoenvironmental Reconstructions: pH - Fen Profiles.....	103
36.	Paleoenvironmental Reconstructions: pH - Bog Profiles.....	104

INTRODUCTION

Climate change that has occurred since the retreat of the Laurentide Ice Sheet has had a significant impact on the vegetation and lakes of Alberta. Paleolimnological records indicate that a warm dry period reduced lake levels in central Alberta from between 7000 to 3000 years before present (yBP) (Forbes and Hickman 1981, Hickman and Klarer 1981, Hickman 1987). This prolonged and severe drought lowered water levels from eight to eighteen meters in the deeper lakes and dried shallower ponds to salty pans (Anderson *et al.* 1987). Pollen records from central Alberta indicate that this period of increased warmth and decreased precipitation reached its maximum at 5500 to 6000 yBP (Lichti-Federovich 1970). Vegetation in the area surrounding Elk Island National Park (EINP) prior to 4500 yBP was an open grassland with high fire frequency (Vance 1979, Vance *et al.* 1983). With increased precipitation and lower temperatures, the grassland vegetation became covered by a poplar, spruce and birch canopy by 3000 yBP. Pollen records from Clearwater bog located in the southern boreal forest of Manitoba (53°59'N, 101°12'W) indicate that a maximum expansion of prairie grassland occurred in this area between 6,700-4,200 yBP. Spruce forest returned after 4,200 yBP (Nichols 1969).

Climate controls the accumulation of peat material. Basal dates of organic sediments in 52 fens located across west-central Canada are older than 6000 yBP in a zone north of the 54°30'N latitude and younger south of this zone (Zoltai and Vitt 1990). This is attributed to the warm, dry, early Holocene climate causing seasonal droughts and preventing the establishment of peat-forming fen vegetation. As the

climate cooled, fens expanded southwards. Vance (1979) and Vance *et al.* (1983) found that peat development in and surrounding Elk Island National Park did not begin until after 4180 yBP. Climate controls not only the time at which peat begins to accumulate, but the process by which peatlands are initiated.

Peatlands are formed through two processes. The first is termed terrestrialization, an infilling process where small open basins of water become progressively infilled by sediments that accumulate on the bottom, and rafts of vegetation that extend from the lake edge (Kratz and DeWitt 1986). Paludification is a process by which mesic sites become increasingly hydric and are either gradually encroached upon by neighboring peatlands (Futyma and Miller 1986, Kubiw *et al.* 1989, Miller and Futyma 1987, Zoltai and Johnson 1985) or develops conditions suitable for the accumulation of sedge and/or moss peats (Sjörs 1961). Expansion of peatlands via paludification has been documented in Alberta by Kubiw (1987), Kubiw *et al.* (1989), Nicholson (1989), Nicholson and Vitt (1990); in Minnesota by Glaser *et al.* (1990), Griffin (1977), Heinselman (1963), (1970), and Janssens *et al.* 1992; in Labrador by Foster and King (1984); and in Alaska by Klinger *et al.* (1990). Peat development via paludification is more dependent upon regional climatic conditions to maintain hydric conditions than terrestrializing lake basins, which have their own water source. As climatic conditions become more arid, peatland development becomes limited to terrestrialized basins.

At present, Elk Island National Park is situated in a southerly extension of the mixedwood section of the Boreal Forest region (Rowe 1972). A rise in the local bedrock topography of the Cooking Lake

moraine has created a corresponding change in climate (cooler, moister). This area is characterized by a mixture of *Populus tremuloides* Michx., *Populus balsamifera* L., *Betula papyrifera* Marsh., *Picea mariana* (Mill.)BSP., *Picea glauca* (Moench) Voss, *Abies balsamea* (L.) Mill. and *Pinus banksiana* Lamb. Areas immediately adjacent to the moraine are located in the aspen grove or parkland zone (Bird 1961, Rowe 1972), which is considered to be a wide transition zone or ecotone (Rowe 1972, Vance 1979, Vance *et al.* 1983). In this transition, vegetation of the great plains grassland and the boreal forest mix in a complex mosaic. *Populus tremuloides* is the dominant forest cover with sporadic occurrence of *Populus balsamifera* and *Betula papyrifera*. Published maps that identify the southern limit of coniferous trees on the Canadian prairies (Zoltai 1975), place Elk Island just inside the southern limit for several coniferous species, namely, white spruce (*Picea glauca*), black spruce (*Picea mariana*), tamarack (*Larix laricina* (Du Roi) K. Koch), and jackpine (*Pinus banksiana*). The position of these vegetation boundaries and transition zones is predominantly controlled by climate.

Wetland terminology and classification

The term wetland can be defined in many ways (Millar 1976, Mitsch and Gosselink 1986, Zoltai 1988, Zoltai *et al.* 1975). A common concept that occurs in all is the presence of a water table at or near the surface for a sufficient length of time to promote the growth of hydrophilic vegetation. Peatlands are a subgroup of wetlands that are characterized by the accumulation of peat or organic soil. The

Canadian System of Soil Classification (Canada Soil Survey Committee 1978) requires 60 cm of peat accumulation if the peat is fibric, 40 cm if it is mesic or humic, and 10 cm if underlain by bedrock. Many of the terms used to describe different wetland types originated in Europe and are now being used in Canada and the United States. Some terms (swamp, marsh, fen, bog) are used to describe very specific wetland types (Crum 1988, Mitsch and Gosselink 1986, Zoltai 1988). Others are less definitive in nature (mire, muskeg, moor, bottomland, peatland) and are used to describe any peat accumulating wetland. Terms such as ombrotrophic and minerotrophic are used to identify plant communities within the peatland. Ombrotrophic defines plant communities that are "rain fed", or receive water and nutrients from precipitation only (Damman 1986, Malmer 1962 and 1985, Zoltai 1988), and minerotrophic defines plant communities that receive water which has been in contact with the mineral soil. Another set of peatland terminology describes the water source in combination with the basin configuration. Limnogenous peatlands are riparian peatlands (located adjacent to a river or lake), and the water source of the peatland is dependent upon the lake or stream (Damman 1986, Malmer 1985, Zoltai 1988). Topogenous peatlands are isolated depressional basins that have perched water tables. Water source for these peatlands is rain and surface runoff. Ombrogenous bogs, are peatlands that receive nutrient poor water from precipitation only. They can exist in depressional basins or along slopes.

Peat is generally defined as organic residues of incomplete decomposition under water-saturated conditions (National Wetlands

Working Group 1988). The Canadian System of Soil Classification has defined peat physically as soils containing 17% or more organic C or 30% or more organic matter (Canada Soil Survey Committee, Subcommittee on Soil Classification 1978). Peats can be subdivided into three types based upon where it is formed in relation to the water table. Peats which formed below the water table are termed limnic peats; those formed between high water table levels and low water table levels are telmatic peats; those formed above the high water table level are terrestrial peats (Faegri and Iversen 1989).

In the past, wetland classification systems have been designed to inventory waterfowl habitats. Particularly important, have been classification systems that are suitable for large scale aerial surveys and mapping. More recently, wetland classification systems have placed greater emphasis on vegetation, hydrology, and water chemistry parameters. In western Canada, there have been two notable wetland classification systems, J. B. Millar's (1976) system for western Canada, and the Canadian Wetland Classification System (Zoltai *et al.* 1975). In Millar's system, wetlands are restricted to marshes and shallow open waters; he excludes bogs and fens from his system. Millar segregates eight wetland types on vegetation characteristics, and qualifies his groupings by the length of time each group has been flooded. Millar further attempts to relate his classification to water chemistry parameters, namely salinity. The following is a summary of Millar's (1976) wetland classification.

- 1) **Wet Meadow-** an area flooded for 3-4 weeks at a time, and occupied by low stature grasses and sedges with a variety of forbes. In the aspen parkland, willows characteristically dominate.
- 2) **Shallow Marsh-** an area where flooding lasts until July or early August. Vegetation dominated by intermediate height grasses, sedges, and forbes.
- 3) **Deep Emergent Marsh-** an area that is inundated from spring to late summer or fall. Only five species predominate and these are grass like plants, which are taller than the vegetation of the shallow marsh zone.
- 4) **Transitional Open Water Wetland-** an unstable wetland that develops following flooding of emergent vegetation. Vegetation variable, consisting of *Lemna*, aquatic mosses, and liverworts.
- 5) **Open Water Marsh-** a wetland, which has less than 75% of it's diameter as shallow open water, with rooted or floating aquatic plants.
- 6) **Shallow Open Water Wetland-** a wetland which has greater than 75% of it's diameter as shallow open water with aquatic plants.
- 7) **Open Alkali Wetland-** a wetland which is devoid of all vegetation in it's central vegetation zone, except for *Ruppia maritima*.
- 8) **Disturbed Wetlands-** areas either cultivated, grazed or drawn down. Vegetation is primarily forbes mixed with low stature grasses.

The Canadian Wetland Classification System (Zoltai *et al.* 1975) developed by the National Wetlands Working Group is broader in its application and hierarchical, starting with wetland classes that are based on vegetation, hydrology, and water quality. These classes are bog, fen, swamp, marsh, and shallow open water. According to the Wetlands Working Group, bogs are wetlands that have accumulated over 40 cm of peat. The living vegetation is not enriched by groundwater and is hence ombrotrophic. Correspondingly the water is nutrient poor and acidic (pH < 4.6, specific conductivity < 80 μ S). Vegetation is dominated by *Sphagnum* mosses, and ericaceous shrubs. *Picea* trees may be present. Fens have high water tables that are situated on gentle slopes. The water source is a mixture of precipitation and groundwater, thus it has a higher nutrient status and pH than bogs. In situations where groundwater sources are extremely nutrient poor, fens can have low pH and nutrient levels, similar to those of bogs. These nutrient poor fens can also be dominated by *Sphagnum* mosses. However, other fens are dominated by brown mosses mostly of the family Amblystegiaceae, with sedges, shrubs, and trees. Swamps are nutrient rich, oxygenated, productive sites, which have luxuriant tree and shrub cover. The water table is high, persisting as standing water for long periods of time, or as subsurface groundwater flow. Peat forming mosses are absent or minimal, and the peat is often highly humified and woody. Wetlands that are periodically inundated, resulting in the addition of nutrient rich waters and mineral soil are marshes. Marsh waters are oxygen rich and circumneutral to slightly alkaline. Vegetation is dominated by reeds, rushes, and sedges. The Wetlands Working Group defines

shallow open waters as small non-fluvial bodies of standing water. Emergent vegetation is lacking, but floating and rooted aquatic macrophytes may be present. Midsummer water depth is less than 2 meters. The next level of hierarchical division for this classification system is the wetland form. It is based upon the surface morphology of the wetland, presence of surface patterns, tidal effects, and proximity to water bodies. These forms reflect differences caused by environmental factors.

OBJECTIVES

Elk Island National Park was chosen as a suitable location for the documentation of the effect of early Holocene climate on peatland development because of: a) the regional sensitivity to early Holocene climate that has been documented in the paleolimnological and pollen record, b) its close proximity to the southern limit of coniferous trees and, c) its geographical position in a southern extension of the mixedwood section of the Boreal Forest. It was well known that a diversity of wetland types exist within the park, allowing the study of wetland processes on a variety of wetland types. The objectives of this thesis are:

- a) To inventory and classify the wetlands of Elk Island National Park.
- b) To inventory surface water chemistry and physical parameters (peat depth, water level fluctuations, and elevation) characterizing wetlands in the park.
- c) To outline relationships between vegetation, surface water chemistry, and physical parameters.
- d) To examine the influence of the early Holocene thermal maximum on the development of wetlands in the park.

- e) To identify the hydrosereal successional patterns that have occurred in the wetlands through an evolutionary study of peat macrofossils.
- f) To identify some of the processes controlling peatland development in the park, and examine the influence of allogenic and autogenic factors on peatland development.

STUDY AREA

Elk Island National Park is located 37 km east of the eastern limit of the city of Edmonton, in the center of a physiographic region known as the Alberta Plain (Lang 1974). On the Alberta Plain are several hills, known as the Beaver Hills, which rise above the level of the surrounding landscape. About 30 m of overburden, consisting of morainic materials deposited from disintegrating ice sheets separate the surface from the underlying bedrock (Bayrock and Hughes 1962). The sedimentary bedrock in this area is of the Belly River Formation (Emerson 1977), and consists of non-marine sandstones, siltstone, and mudstones. These rocks were formed during the Upper Cretaceous and are erosional products from previously existing hills (Lang 1974). Tills in Elk Island National Park are formed from hummocky stagnant ice moraine (Lang 1974), giving the area a rugged topography characterized by knobs, kettles, till hummocks and till ridges (Scace and Associates Ltd. 1976). Prairie mounds are a distinctive feature of dead-ice moraine and within the Beaver Hills these mounds are circular, with a basal diameter of about 91 m, and a rim that rises 5 m above the surrounding land. Within the central depression, a thin layer of lacustrine deposit can usually be found (Emerson 1977).

There is evidence, in the form of glaciofluvial deposits and landforms, that the Beaver Hills experienced a greater supply of surface water during the ice stagnation than at present (Emerson 1977). At Elk Island, this evidence occurs as small meltwater stream trenches at Tawayik and Oster Lakes. This is a route from which drainage water collected in the southern half of Elk Island National Park flowed northward into glacial Lake Edmonton (Emerson 1977). Glaciolacustrine sediments were also deposited in a hummocky area southwest of Astotin Lake (Crown 1977), and the southeastern part of the Park (Lang 1974).

Climate

The mean annual temperature for Elk Island Park is 1.5°C (Parks Canada 1986), with a mean summer temperature of 9.3°C. The region surrounding the park receives 400-500 mm of precipitation annually, with between 325 and 391 mm as rain. June, July and August are the wettest months. The area experiences on average between 1278 to 1560 growing degree days a year, and 109 to 140 frost free days.

Vegetation

There has been much controversy in the past regarding the concepts and boundary locations of the main phytogeographic regions or forest regions surrounding Elk Island National Park. In 1955, Moss placed Elk Island in the Parkland Prairie phytogeographic region of Alberta (Parks Canada 1986); later Rowe (1972) placed Elk Island in the Boreal mixedwood forest region. In a recent ecological land

classification by Strong and Leggat (1981), Elk Island was placed in the Boreal mixedwood. Consultants for Parks Canada (Parks Canada 1986) considered the Park to be in the Parkland-Boreal Forest Transition Ecoregion, as it contains areas of aspen groves and mixedwood forest. According to the latest Ecoclimatic regions of Canada publication (Ecoregions Working Group 1989), Elk Island National Park is in the Transitional Grassland Ecoclimatic Region, a transition zone between semi-arid prairie, and moister boreal forests. Vegetation studies done for Parks Canada at Elk Island identified 15 communities comprised of *Populus tremuloides*, *Populus tremuloides*-*Populus balsamifera*, *Betula papyrifera*, *Picea glauca*, and *Picea mariana*. Three of the fifteen were shrub communities of *Betula*, *Salix* and *Ledum groenlandicum*. Six grass and sedge communities were described: *Puccinellia distans*, *Poa pratensis*, *Bromus inermis*, *Carex*, *Carex-Salix*, and *Typha*. The most common assemblages were found to be the *Populus* assemblage with *Corylus cornuta* and *Aralia nudicaulis* as co-dominants; a shrub assemblage with *Poa pratensis*; the grass assemblage *Poa pratensis*; and the *Carex* and *Typha* assemblages. In terms of land area, the most widespread community types are the *Typha-Carex*; *Carex*; and *Carex-Petasites* dominated vegetation occupying 53% of the park's land area (Parks Canada 1986).

METHODS

Vegetation Inventory

Preliminary airphotographic reconnaissance and information from previous vegetation and soils studies available at Elk Island Park (Crown 1977, Parks Canada 1986) indicated that approximately ten wetland types (including peatlands) could be identified within the park. Based on these surveys, 63 wetlands were chosen for the vegetation inventory. Wetland selection was based on adequate duplication of the ten major wetland types, adequate geographic coverage of the park and reasonable access. Fig. 1 outlines the location of wetland sites inventoried. At each of the 63 wetlands, stands were identified based on uniform vegetation, generally arranged in zones. From one to three relevés each of 10 square meters were randomly placed within each stand. Vegetation was described on a percentage cover basis, and mean cover values of the one to three relevés were calculated for each stand. Voucher specimens were collected for vascular plants and bryophytes and are deposited in the University of Alberta herbarium (ALTA). Authority names and nomenclature for vascular plants follows Moss (1983), except for Cyperaceae (Taylor 1983); the genus *Sphagnum* follows (Vitt and Andrus 1977), the genus *Drepanocladus* (Janssens 1983), other bryophytes (Ireland 1982), hepatics (Schuster 1953), and lichens (Hale 1979). Physical parameters measured at each relevé included: water depth measured in open pools or shallow pits, hummock height, and pH of surface waters. Site elevation was taken from Canada Dept. of Energy, Mines and Resources (1987) 1:50,000

ELK ISLAND NATIONAL PARK

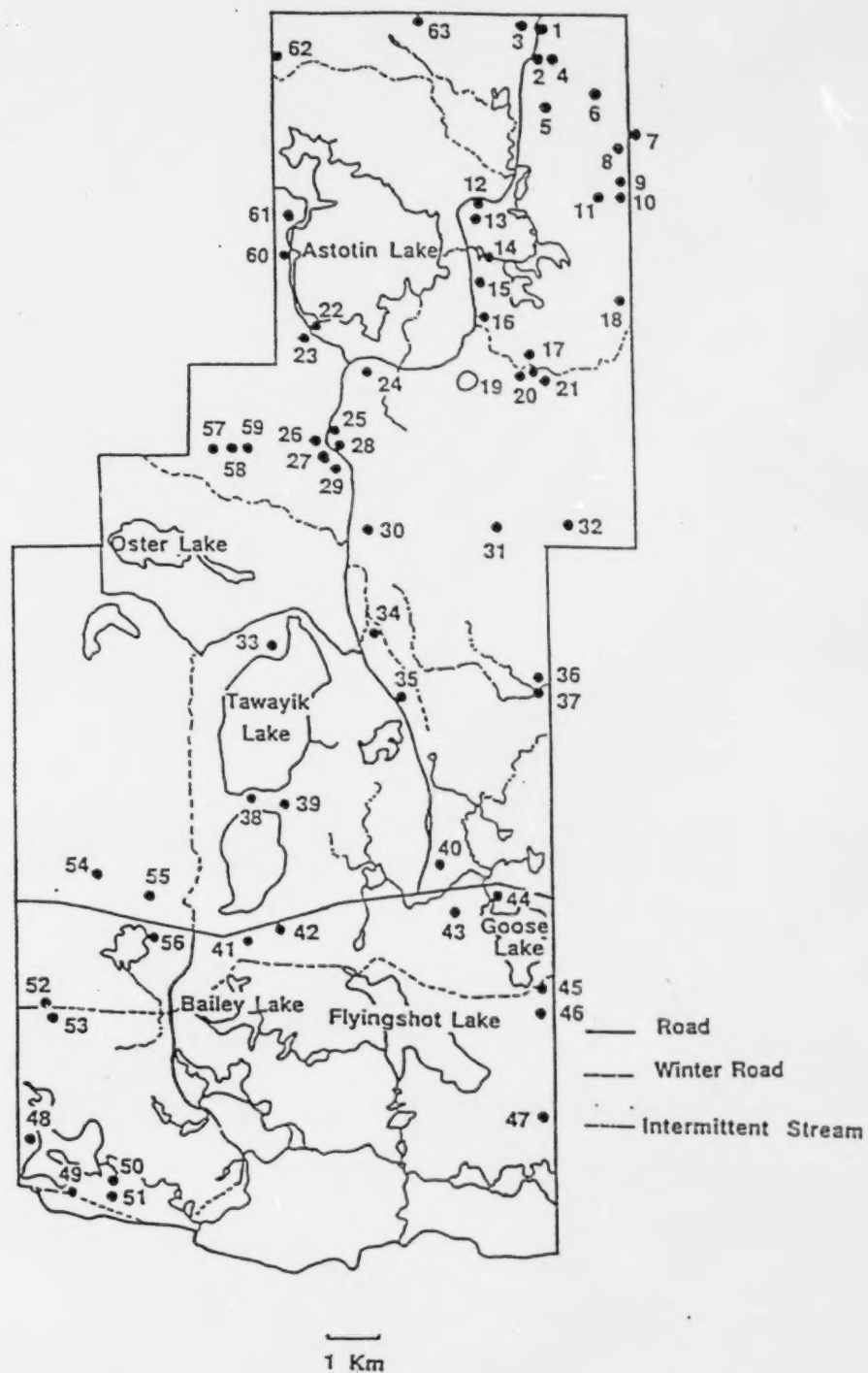


Figure 1. Location of 63 Wetlands Used In The Vegetation Stand Inventory

contour maps, and the number of streams entering and exiting from 1:30,000 aerial photographs of the Park.

Water Chemistry

On May 15-16 and September 18-20, 1989, water samples were collected from open pools or shallow pits from the stands. Samples analyzed for specific conductivity and cations were stored in acid washed 60 ml polyethylene and polypropylene bottles. Samples analyzed for total phosphorus were stored in 500 ml polypropylene bottles, while samples analyzed for total Kjeldahl nitrogen, nitrate, and ammonium were placed in 60 ml and 500 ml polystyrene bottles. All nutrient (TKN, nitrate, ammonium, total phosphorus) samples were analyzed within 24-48 hours; cation samples were preserved with 1 ml of 1.5 N HCL and analyzed at a later date. All pH measurements were taken in the field using a portable Beckman 10 pH meter. Specific conductivity measurements were done in the lab with a Radiometer conductivity meter. Results for specific conductivity were adjusted to 25°C and corrected for hydrogen ion activity (Sjörs 1952). Cations were analyzed at Forestry Canada, Northern Forestry Center, Edmonton, Alberta, on an inductively coupled plasma spectrophotometer. Nitrate and ammonium were analyzed at the University of Alberta, Department of Zoology, on a Technicon autoanalyzer, and total Kjeldahl nitrogen samples were persulfate digested using methods outlined by D'Elia *et al.* (1976).

Stand data were analyzed by TWINSPAN, a Fortran program that orders multivariate data into a classification scheme (Hill 1979). Spearman's Rank Correlation statistic, a non-parametric statistic, was

used to obtain correlations between the Twinspan classification, water chemistry, and physical parameters.

Wetland Development Analysis

Classification of the vegetation data by TWINSpan identified in thirteen vegetation groups. Based on this analysis, six representative peatlands were subjectively selected from the 63 sites for a development analysis (Fig. 2). In order to analyze developmental trends, sites were selected representing the seral stages of bog, fen and marsh. A bog which had been disturbed by beaver activity was added after the first field season in response to a request from the staff at EINP. Cores were removed at each site in a transect. Depth to mineral substrate was determined by probing with iron rods at 30.5 m. intervals to develop basin configuration at five sites. Triplicate peat cores were removed manually with a modified 5 cm diameter Macaulay peat corer. One core was used for macrofossil analysis, the second for peat chemistry and the third for radiocarbon dating. Macrofossil analysis follows the methodology of Janssens (1988), in which subsamples of a known volume are removed from the core, soaked for three days in Aerosol O.T. Solution (a non-foaming wetting agent), and wet sieved with a 150 μ m soil sieve. Relative percentages of the macrofossils present in the subsample were determined using a dissecting microscope. Seeds, bracts, and reproductive structures too small for percentage estimates were counted directly. Zonation of the peat profile into peat phases was determined visually based on species composition and ecology. The central core from each transect has been illustrated in detail with all

ELK ISLAND NATIONAL PARK

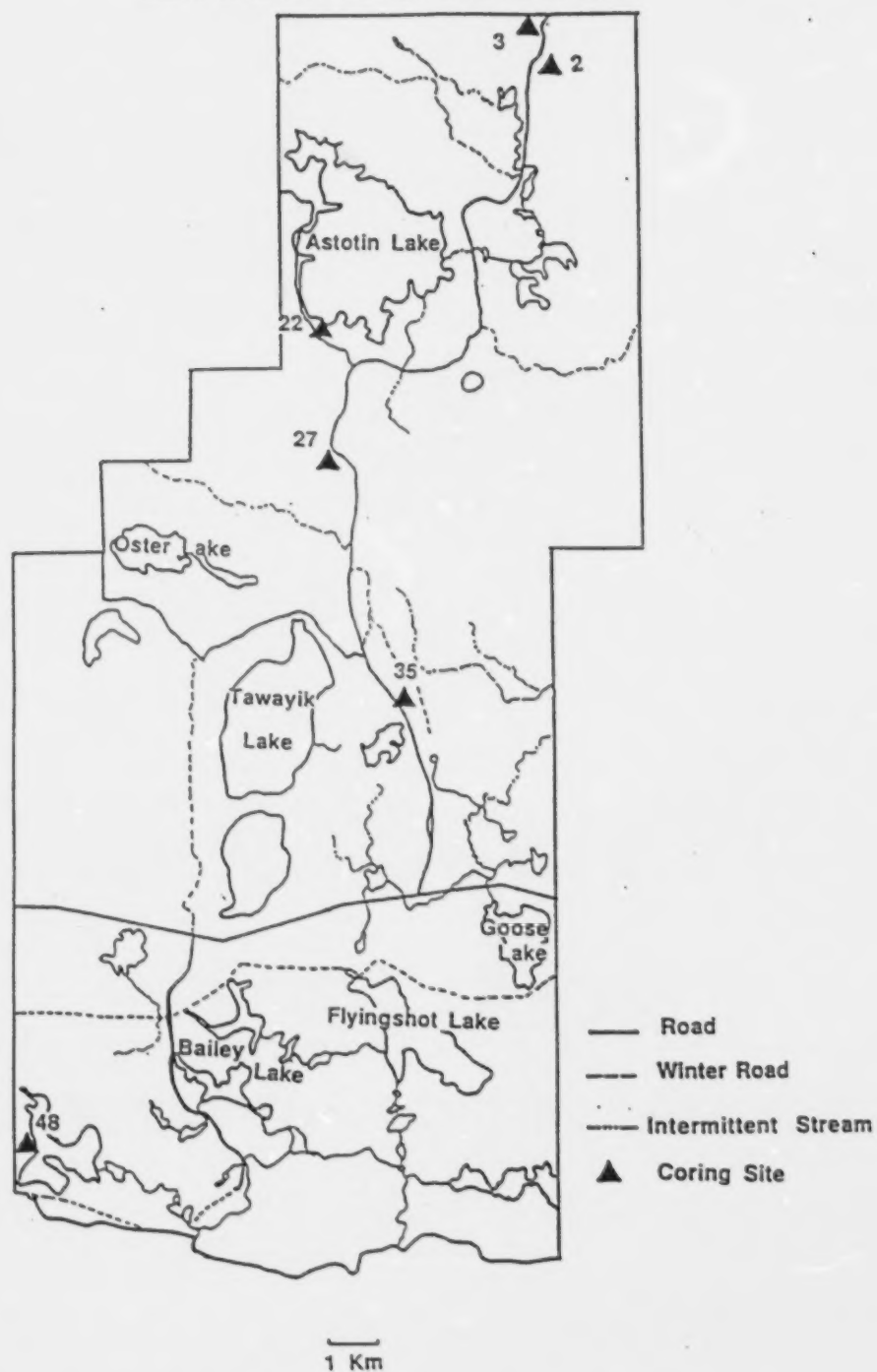


Fig. 2. Location of Six Peatlands Used For Evolutionary Analysis

macrofossil fragments and seeds identified. For the remainder of the cores from each transect, summary diagrams are presented outlining the important macrofossils. A detailed listing of all macrofossil fragments and seeds identified in the summary diagrams is presented in Appendix I. Paleoenvironmental reconstructions of height above water table and pH were calculated based on the known modern autecology of specific moss species. Weighted species means of height above the water table and local pH were calculated on 1331 relèves measured throughout western Canada (Gignac *et al.* 1991a,b).

$$\mu_k = \frac{\sum_{i=1}^n Y_{ik} X_i}{Y_{+k}}$$

Where Y_{ik} is the abundance measured as % cover of species k on reléve i , X_i is the pH or height above the water table for reléve i , and Y_{+k} is the total abundance of species k .

Hydrogen ion measurements were log transformed to approach a normal distribution before calculating weighted means, then backtransformed to the original scale. This method is more accurate than using untransformed data because the resulting means account for the skewness in the data (Gignac pers. com.). Weighted means for pH and height above the water table for mosses in the modern reference data set are presented in Appendix II. Inferred pH and height above the water table (x) for macrofossil sample (m) was calculated as follows:

$$X_m = \sum_{k=1}^p Y_{km} \mu_k / Y_{+m}$$

where Y_{km} is the abundance of moss species k in macrofossil sample m , μ_k is the modern weighted mean of moss species k , and Y_{+m} is the total abundance of mosses in macrofossil sample m .

Peat Chemistry

Peat chemistry samples were removed from each core at 10 cm intervals and major macrofossil boundaries as seen in the field. Each sample was weighed, oven dried at 105°C for 1-2 days and reweighed to derive bulk density measurements. Samples were then ground with a mortar and pestle. Subsamples of 0.2-0.3 g were removed, dry ashed at 470°C, and acid digested with 6 ml of 1.5 N HCL and 1 ml of concentrated HNO₃. Digested solutions were then filtered through Whatman #42 filter paper and analyzed by inductively coupled plasma spectrophotometry. Results are given in mg/Kg of peat. Radiocarbon dates were obtained at the Radiocarbon and Tritium Laboratory in Vegreville, Alberta. After the macrofossil analysis were completed the peat samples were assigned to one of eight peat type categories based on the results. In order to assess how the peat types based on macrofossil composition differed chemically a SAS/STAT canonical correlation and discriminant analysis were performed. The discriminant analysis creates composites of the chemical variables showing maximum differences among the peat types. The canonical correlation defines the relationships between

the chemical variables (Thorndike 1978). Group means were plotted in discriminant space to graphically show the relationships between the peat types. Multivariate statistics test the hypothesis that the class means are equal in the population.

RESULTS

Vegetation Classification

274 relevés from the 63 wetlands contained 260 species (Table 1) from 156 stands. TWINSpan analysis classified these stands into 13 ecologically meaningful stand groups.

Group 1 consists of one unique stand located at site 49 a brackish environment. Dominant species are *Triglochin palustris*, *Triglochin maritima*, and *Eleocharis palustris*. Species present in minor abundance are *Ranunculus gmelinii*, *Scirpus validus*, and *Rumex britannica*. Species diversity is low in this stand with 2.3% of the total species present.

Group 2 is a small group of 10 stands located predominantly along lake edges. Characteristic species are *Typha latifolia*, *Lemna minor*, *Carex atherodes*, *C. aquatilis*, *Sium suave*, and *Ranunculus gmelinii*. Bryophytes are absent except for minor occurrences of *Drepanocladus polycarpus*. Species diversity is low with only 20% of the total species present.

Group 3 contains 22 stands, characterized by high *Lemna* and *Carex* cover, and can be found at wetland margins, on shallow peats, and in areas recently flooded by beaver (*Castor fiber*). The *Carex*

Table 1
Twinspan Vegetation Groups
Wetlands of Elk Island National Park

Group	1	2	3	4	5	6	7	8	9	10	11	12	13
Species													
<i>Bidens cernua</i>		+	+		+	+	+				0.67		
<i>Lemna trisulcata</i>		0.8	+		+						0.67		
<i>Utricularia vulgaris</i>		0.6	+		+			+					
<i>Glyceria pulchella</i>			+	+		+						+	
<i>Riccia fluitans</i>		+	+		+						+		
<i>Eleocharis palustris</i>	4	1.2	2		+	+	+				+		
<i>Lemna minor</i>		1.3		+	+	+	+				0.67		
<i>Polygonum amphibium</i>		+	0.86		+	+	+						
<i>Ranunculus gmelinii</i>	1	1	+		+	+	0.56				+		
<i>Scirpus validus</i>	1	+	+			+							
<i>Triglochin palustris</i>	3			+		+							
<i>Typha latifolia</i>		3.4	0.73	+	+	+					+		
<i>Carex etherodes</i>		1.3	2.18			1.03	+		+				
<i>Carex bebbii</i>		+	+		+	+							
<i>Beckmannia syzigachne</i>		+				+							
<i>Glyceria grandis</i>			0.91		+	+							
<i>Hordeum jubatum</i>			+			+							
<i>Hippuris vulgaris</i>		+		+	+								
<i>Utricularia minor</i>		0.6	+	+	+								
<i>Carex rostrata</i>			1.45	0.58	1.73	0.69	0.56				+		
<i>Cicuta bulbifera</i>		+	+	+	+	+						+	
<i>Galium trifidum</i>	1	0.91	+	0.93	1.28	0.56			+			+	
<i>Sium suave</i>		1.2	0.82	+	+	0.52	+				+		
<i>Calamagrostis stricta</i>		+	+	0.58	+	0.66							
<i>Epilobium leptophyllum</i>		+		+	+					+			
<i>Mentha arvensis</i>		+	+	+	+	0.79	0.67	+			0.67		
<i>Stellaria longifolia</i>		+	+	+	+	0.97	0.78	+	+		0.67		
<i>Carex aquatilis</i>		1.7	2.55	1.54	3.57	1.86	+	+	+	1.5	1		
<i>Cirsium arvense</i>			+		+	+							
<i>Epilobium ciliatum</i>		+	+	+	+	+	0.56					+	
<i>Stachys palustris</i>			+			0.59	+						
<i>Urtica dioica</i>		+	+		+	+	+						
<i>Scholochloa festuacea</i>		+	+		+	+							
<i>Salix petiolaris</i>		+		+	+	0.55							
<i>Cicuta virosa</i>		+	+	+	+	+	+						
<i>Geum aleppicum</i>			+	+	+	+							
<i>Rumex britannica</i>	1	+	+	+	0.87	0.93	+		+				+
<i>Scutellaria galericulata</i>		+	+	+	+	1.21	0.67		+		+	+	
<i>Salix serotima</i>			+	+	+	0.83			+	1	0.67	+	+
<i>Aster borealis</i>			+	+	+							+	
<i>Impatiens capensis</i>		+	+	+	+	2						1.33	
<i>Lycopus uniflorus</i>		+	+		+	0.66	0.56	+			0.67		
<i>Agrostis scabra</i>			+	+	+	0.93	+	+	+				+
<i>Poa palustris</i>		+	+		+	0.66	+	+					
<i>Salix bebbiana</i>			+	+	+	+	+		+				
<i>Salix discolor</i>		+	+	+	+	1.24	+		+	1	+		
<i>Salix exiguus</i>				+	+	+							+
<i>Aster puniceus</i>			+	+	+	1.1	0.67				+	1	
<i>Caltha palustris</i>			+	+	+	0.59	1.44					0.67	
<i>Geum rivale</i>		+	+		+	0.62	0.89					1	
<i>Lonicera involucrata</i>			+	0.67	+	0.89	+				+	0.67	
<i>Petasites sagittatus</i>			+	0.54	+	1.07	0.56					0.67	+
<i>Ribes glandulosum</i>						+	+					+	
<i>Ribes lacustre</i>			+			+	0.56				+		
<i>Bromus ciliatus</i>				+		0.52	+	+					+
<i>Amblystegium serpens</i>			+	+	+	+	+						
<i>Climacium dendroides</i>		+	+	+	+	1	1		+		0.67	+	
<i>Hypnum pratense</i>				+	+	+	+				+	0.67	
<i>Marchantia polymorpha</i>				+	+	+	1.33					0.67	
<i>Plagiomnium ellipticum</i>			+	0.92	0.67	0.97	2.11				+	1.67	
<i>Potentilla palustris</i>		+	+	1.83	1.67	1.38	1.56		+	1	+	+	0.6
<i>Plagiomnium medium</i>					+	+	+						
<i>Equisetum fluviatile</i>				1.08	0.67	+	2.33			1		1.67	
<i>Eriophorum polystachion</i>				0.88	+	+	+						
<i>Lysimachia thyrisiflora</i>			+	+	+	0.66	0.78		+		0.67	0.67	
<i>Menyanthes trifoliata</i>				1.79	+	+	+			1.5	+	1.33	
<i>Brachythecium mildeanum</i>			+	1.46	+	0.76	0.78	+	+			1.33	0.6
<i>Calliergonella cuspidata</i>				+		+	+					0.67	
<i>Drepanocladus aduncus</i>			+	0.67	0.67	+	+				+	+	
<i>Drepanocladus polycarpus</i>		0.6	+	1.33	1.23	1.07	+	+	+		+	0.67	+
<i>Cornus stolonifera</i>					+	+	+					+	

Table 1 continued

Group	1	2	3	4	5	6	7	8	9	10	11	12	13
Species													
<i>Ribes oxycanthoides</i>				+	+	+	+						+
<i>Viola palustris</i>					+	+						+	
<i>Salix pedicularis</i>				1.38	+	+		+	+				
<i>Drepanocladus capillifolius</i>				+	+	+						+	
<i>Salix candida</i>				+	+	+	0.56						
<i>Equisetum arvense</i>				+	+	+							
<i>Triglochin maritima</i>	3			0.71	+	+							
<i>Carex chordorrhiza</i>				+	+	+							
<i>Carex diandra</i>		+		1.79	+	+	+	+				+	
<i>Carex canescens</i>				+	+	+							
<i>Carex lasiocarpa</i>				2.38	0.77	+		+		1			1.6
<i>Drepanocladus vernicosus</i>				1.5	1.1	2.9	2.22	0.67			3	1	
<i>Calamagrostis canadensis</i>	+	0.91		1.25	+	+			1	1			
<i>Calla palustris</i>				+	+	+						0.67	
<i>Tomenthypnum nitens</i>				0.75	+	+	+	+	+		0.67		+
<i>Salix pyrifolia</i>	+	+		+	0.67	+	+	+	+		1	2	0.6
<i>Rubus acaulis</i>				0.83	+	1.31	1.44	+			+		
<i>Drepanocladus exannulatus</i>				+	+	+		+	+	+	+		
<i>Sphagnum squarrosum</i>				+	+	+		+				+	
<i>Dryopteris carthusiana</i>					+	+		+				0.67	
<i>Mitella nuda</i>					+	+	1.33	+	+		+	+	
<i>Ribes hudsonianum</i>					+	+	+	+	+		+	0.67	+
<i>Rubus idaeus</i>					+	+	+				1	+	
<i>Viola renifolia</i>				+	+	+	+					0.67	
<i>Viola spp.</i>					+	+						+	
<i>Pyrola secunda</i>			+	+		+							
<i>Rhizomnium gracile</i>			+	+	+	+		+	+		0.67	+	+
<i>Alnus tenuifolia</i>			+	+	+	+	+	+	+			1.33	
<i>Galium boreale</i>	+					+	+	+				1.33	
<i>Pyrola asarifolia</i>			+	+		+	+	+	+				+
<i>Salix spp.</i>			+	+	+	+	+	+	+				+
<i>Bryum pseudotriquetrum</i>			+	0.79	+	+	+	+	+				
<i>Calliergon stramineum</i>				+	+	+	+	+	+		0.67		+
<i>Epilobium angustifolium</i>				+	+	+	+	+	+		0.67		+
<i>Aulacomnium palustre</i>				1.33	+	+	+	1.2	0.83	+		1.67	1.6
<i>Calliergon giganteum</i>				0.88	+	+	1.22	+			2.33	1.33	
<i>Helodium blandowii</i>				+	+	+	+	+			0.67	+	
<i>Betula pumila</i>				1.21	+	+	+	0.8			1.33	+	1.6
<i>Lophocolea heterophylla</i>				+	+	+	+	+	+			+	
<i>Larix laricina</i>				0.79	+	+	+	+		2		1	+
<i>Carex limosa</i>				+	+	+	+	+		2			
<i>Betula nealaskana</i>				0.58	+	+	1.56	1.13			2.33	1	1.4
<i>Plagiothecium denticulatum</i>			+	+	+	+	+	+	+			1	
<i>Cornus canadensis</i>				+	+	+	+	+	+			1.67	0.6
<i>Pohlia nutans</i>			+	+	+	+	+	+	+			1.67	
<i>Picea glauca</i>				+	+	+	1	+	0.83		0.67	2	1
<i>Carex disperma</i>				+	+	+	+	+				1.67	
<i>Cinna latifolia</i>			+	+	+	+	+	+				1	
<i>Brachythecium campestre</i>					+	+	+	+	+	2.5			0.8
<i>Betula glandulosa</i>					+	+	+	0.67	0.67	1	0.67		0.8
<i>Carex curta</i>		+		0.58	+	+	+	+		1	0.67		+
<i>Carex paupercula</i>				+	+	+	+	0.67		4	1.67		1.2
<i>Sphagnum teres</i>				+	+	+	+	1.93	+	2	1		1.2
<i>Eriophorum vaginatum</i>				0.67	+	+	+	0.93	1.17	1.5	2	2	2.4
<i>Sphagnum warnstorffii</i>				+	+	+	+	3.4	2.83	1	2.67	2	3.2
<i>Ledum groenlandicum</i>	+			+	+	+	+	1.6	1.67	2.5	1	0.67	1.6
<i>Oxycoccus microcarpus</i>				+	+	+	+	2.67	2.33		2	0.67	3
<i>Rubus chamaemorus</i>					+	+	+	0.93	1.17	1	2	1	1.2
<i>Smilacina trifolia</i>				+	+	+	+	1.67	0.83		1	0.67	2
<i>Vaccinium myrtilloides</i>					+	+	+	1.93	0.83	2	0.67	1	1.8
<i>Polytrichum strictum</i>					+	+	+	2.07	1	2.5	1.67	1	2.2
<i>Sphagnum angustifolium</i>				+	+	+	+	2.33	1		1.67	1	2.4
<i>Sphagnum mangellanicum</i>					+	+	+	1.27			0.67	+	+
<i>Sphagnum russowii</i>					+	+	+	1.47	3		+	1	1
<i>Picea mariana</i>				+	+	+	+	1.11	2	2.33		1.67	1.8
<i>Vaccinium vitis-idaea</i>				+	+	+	+	+	+			+	+
<i>Dicranum undulatum</i>				+	+	+	+	+	0.83				+
<i>Lophozia spp.</i>					+	+	+	+	1.17				2
<i>Polytrichum juniperinum</i>					+	+	+	1.6	1.17		+		0.8
<i>Sphagnum fuscum</i>					+	+	+	0.67	2.5			+	+
<i>Sphagnum nemoreum</i>				+	+	+	+	+	1		+	+	+
<i>Cladonia mitis</i>								+	0.67				+
<i>Cladonia rangiferina</i>								+	0.83				
<i>Cladonia chlorophaea</i>								+	0.8	1.17			0.6
<i>Betula papyrifera</i>			+	+		+	+						

Table 1 continued

Group	1	2	3	4	5	6	7	8	9	10	11	12	13
Species													
<i>Hylocomium splendens</i>						+	+		0.83			1	
<i>Pleurozium schreberi</i>						+	+	0.67	2			2.33	0.6
<i>Ptilidium crista-castrensis</i>						+	+	+	+			1	+
<i>Sphagnum centrale</i>							+	+	+			1.33	+
Species with minor occurrences													
<i>Juncus bulbosus</i>			+										
<i>Monolepis nuttalliana</i>			+										
<i>Ranunculus lapponicus</i>		+	+										
<i>Rorippa palustris</i>		+	+			+							
<i>Sperganium angustifolium</i>		+											
<i>Viola canadensis</i>			+										
<i>Phalaris arundinacea</i>		+				+							
<i>Myriophyllum exallescens</i>		+											
<i>Potamogeton pectinatus</i>			+										
<i>Drepanocladus lapponicus</i>			+										
<i>Funaria hygrometrica</i>			+										
<i>Lysimachia thysiflora</i>		+											
<i>Ricocarpus natans</i>		+	+		+								
<i>Rhizomnium ellipticum</i>			+		+								
<i>Achillea millefolium</i>		+				+							
<i>Equisetum pratense</i>				+		+							
<i>Ribes americanus</i>			+			+							
<i>Senecio eremophilus</i>		+	+			+	+						
<i>Carex lacustris</i>			+		+	+							
<i>Salix maccalliana</i>						+							
<i>Salix scouleriana</i>						+							
<i>Amelanchier alnifolia</i>						+							
<i>Achillea sibirica</i>						+							
<i>Anemone canadensis</i>						+							
<i>Anemone cylindrica</i>						+							
<i>Cardamine pensylvanica</i>						+	+						
<i>Equisetum palustre</i>						+							
<i>Fragaria vesca</i>			+			+							
<i>Erysimum</i> sp.						+							
<i>Galium triflorum</i>						+	+						
<i>Geum macrophyllum</i>				+		+							
<i>Lathyrus ochroleucus</i>						+							
<i>Lathyrus venosus</i>						+							
<i>Potentilla norvegica</i>						+							
<i>Potentilla paradoxa</i>			+			+							
<i>Ribes triste</i>							+						
<i>Ribes viscosissimum</i>						+							
<i>Solidago canadensis</i>						+							
<i>Solidago gigantea</i>							+						
<i>Viola nephrophylla</i>						+							
<i>Carex sychnocephala</i>						+							
<i>Elymus glaucus</i>						+							
<i>Carex brunnescens</i>						+							
<i>Brachythecium plumosum</i>							+						
<i>Cephalozia connivens</i>						+							
<i>Leptobryum pyriforme</i>						+							
<i>Plagiomnium cuspidatum</i>						+							
<i>Peltigera canina</i>						+	+						
<i>Rhizomnium pseudopunctatum</i>				+		+							
<i>Salix pseudomonticola</i>				+	+	+	+						
<i>Parnassia palustris</i>				+	+	+							
<i>Ranunculus macounii</i>				+	+	+							
<i>Rosa acicularis</i>				+	+	+							
<i>Senecio indecorus</i>					+	+							
<i>Spiranthes romanzoffiana</i>				+		+							
<i>Carex retrorsa</i>				+		+	+						
<i>Thuidium recognitum</i>				+		+							
<i>Salix arbusculoides</i>				+									
<i>Salix candida</i>				+	+	+	+						
<i>Moneses uniflora</i>				+									
<i>Scheuchzeria palustris</i>				+									
<i>Carex tenuiflora</i>				+									
<i>Campyllum polygamum</i>				+									
<i>Cephalozia bicuspidata</i>				+									
<i>Drepanocladus revolvens</i>				+									
<i>Meesia triquetra</i>				+									
<i>Peltigera aphthosa</i>				+									
<i>Sphagnum contortum</i>				+									
<i>Calamagrostis inexpansa</i>	+			+	+	+				+			
<i>Petasites palmatus</i>				+	+	+						+	

Table 1 continued

Group	1	2	3	4	5	6	7	8	9	10	11	12	13
Species													
<i>Carex trisperma</i>					+	+							+
<i>Hypnum lindbergii</i>				+		+			+				
<i>Equisetum sylvaticum</i>						+							+
<i>Campylopus radialis</i>						+						+	
<i>Habenaria hyperborea</i>				+		+						+	
<i>Pinus banksiana</i>				+	+			+					
<i>Lycopodium annotinum</i>						+							+
<i>Cladonia scabruscula</i>					+		+					+	
<i>Ceratodon purpureus</i>						+			+				
<i>Salix planifolia</i>					+	+							
<i>Drepanocladus uncinatus</i>						+	+		+				+
<i>Populus tremuloides</i>				+		+			+				
<i>Cladonia cucullata</i>				+				+					
<i>Andromeda polifolia</i>				+						+			+
<i>Drosera rotundifolia</i>					+					+			+
<i>Kalmia polifolia</i>											+		+
<i>Cephalozia lunulifolia</i>							+				+		+
<i>Calypogeia sphagnicola</i>							+	+			+		+
<i>Scirpus hudsonianus</i>									+				
<i>Trifolium repens</i>								+					
<i>Myrica anomala</i>								+	+				
<i>Ptilium pulcherrimum</i>								+					
<i>Sphagnum fallax</i>								+					
<i>Sphagnum riparium</i>								+					
<i>Sphagnum subsecundum</i>				+					+				+
<i>Cladonia cornuta</i>						+		+	+				
<i>Cladonia deformis</i>								+	+				
<i>Cladonia furcata</i>								+	+				
<i>Cladonia gracilis</i>								+	+				
<i>Cladonia verticillata</i>								+	+				
<i>Cetraria nivalis</i>						+		+	+				+
<i>Cetraria cucullata</i>								+	+				
<i>Dicranum polysetum</i>												+	
<i>Lepidozia reptans</i>												+	
<i>Tetraphis pellucida</i>									+			+	
<i>Cladonia coniocraea</i>												+	

Numbers are mean abundance values of each TWINSpan vegetation group on a scale from 1-4, cut levels were 0, 2, 20, 50
 + = a mean abundance TWINSpan value of < or = 0.5

species are broad-leaved and consist of *Carex aquatilis*, *Carex rostrata*, and *Carex atherodes*. Other species that occur are the grasses *Beckmannia syzigachne*, *Hordeum jubatum*, *Glyceria grandis*, and *Calamagrostis canadensis*, and other angiosperms *Polygonum amphibium*, *Typha latifolia*, *Galium trifidum*, and *Sium suave*. Bryophytes are not major components of this group and consist of only a few occurrences of *Plagiomnium ellipticum*, *Brachythecium mildeanum*, *Drepanocladus polycarpus*, and *D. aduncus*. Thirty-two percent of the total species are found in these stands.

Group 4 contains 24 stands that can either be found in floating mats at lake edges, as slightly raised center mats, or at the edges of peatlands. This stand group contains a mixture of the narrow-leaved carices, (*Carex lasiocarpa*, *C. diandra*), and broad-leaved carices, (*Carex aquatilis*, *Carex rostrata*). *Larix laricina* and *Salix pedicellaris*, as well as the grasses *Calamagrostis canadensis* and *C. stricta*, can be present. Other common vascular plants include *Menyanthes trifolitata*, *Potentilla palustris*, *Betula pumila*, *Betula neoalaskana*, *Equisetum fluviatile*, *Eriophorum polystachion* and *Eriophorum vaginatum*. Bryophytes form important components of these stands, particularly *Brachythecium mildeanum*, *Plagiomnium ellipticum*, *Drepanocladus polycarpus*, *D. vernicosus*, *Tomenthypnum nitens*, *Calliergon giganteum*, and *Calliergonella cuspidata*. Forty-seven percent of the total species are present in these stands.

Group 5 contains the largest number of stands (30), and was sampled under a variety of conditions. The stands are characterized by high cover of *Carex aquatilis*, and *C. rostrata*, two common broad-leaved species. Bryophytes are less important than in group 4.

Common secondary species are *Calamagrostis canadensis*, *Carex lasiocarpa*, *Galium trifidum*, *Rumex britannica*, *Petasites sagittatus*, *Potentilla palustris*, *Drepanocladus polycarpus*, *Drepanocladus aduncus*, and *Plagiomnium ellipticum*. Species diversity is relatively high, with 49% of the total species present.

Group 6 contains 29 stands that are generally located on either shallow peat or on the edges of peatlands. These stands are dominated by *Salix* trees and shrubs with high grass cover in the understory layer. *Salix discolor* is the most common *Salix* species, followed by *S. serissima*, and *S. pyrifolia*. Grasses are dominated by *Calamagrostis canadensis* and *Agrostis scabra*, while *Carex aquatilis* and *C. atherodes* are the most common carices. Flowering plants are diverse and include *Potentilla palustris*, *Cicuta bulbifera*, *Scutellaria galericulata*, *Petasites sagittatus*, *Stellaria longifolia*, *Rumex britannica*, *Galium trifidum*, *Aster puniceus*, and *Rubus acaulis*. *Drepanocladus polycarpus* is the only bryophyte of importance in this group. Species diversity is the highest in this stand group with 71% of the total species present.

Group 7 is a second series of stands that appears commonly along the edges of the wetlands. This group of 10 stands has the highest percentages of *Alnus tenuifolia*, *Salix candida*, and *Betula neoalaskana*. *Picea mariana*, *P. glauca*, and *Larix laricina* are also found, but with low cover. The grass, *Calamagrostis canadensis*, is also an important component. Other vascular plants that are prevalent include: *Equisetum fluviatile*, *Potentilla palustris*, *Impatiens capensis*, *Vaccinium vitis-idaea*, *Ribes hudsonianum*, *Rubus acaulis*, and *Aster puniceus*. Bryophytes are common in this vegetation,

particularly *Brachythecium mildeanum*, *Calliergon giganteum*, *Aulacomnium palustre*, *Sphagnum warnstorffii*, *Pleurozium schreberi*, *Plagiomnium ellipticum*, *Climacium dendroides*, and *Marchantia polymorpha*. Thirty-eight percent of the total species are present in this group.

Group 8 contains 10 stands found in the more open peatlands. Vegetation is dominated by *Sphagnum* mosses and ericaceous shrubs. The *Sphagnum* species include: *S. fuscum*, *S. magellanicum*, *S. angustifolium*, *S. russowii*, and smaller amounts of *S. warnstorffii*, and *S. nemoreum*. Ericaceous shrubs consist of *Ledum groenlandicum*, *Oxycoccus microcarpus*, *Vaccinium myrtilloides*, and *V. vitis-idaea*. Vascular plant species are limited and dominated by *Rubus chamaemorus* and *Eriophorum vaginatum*. Lichens, particularly *Cladina mitis*, *C. rangeriferina*, and *C. chlorophaea* are present. Species present in this group are 26% of the total.

Group 9 also consists of 6 stands found in *Sphagnum*-dominated peatlands, but with a heavy black spruce (*Picea mariana*) cover (Table 1). Ericaceous shrubs are significant in Group 9 as well as the *Sphagnum* species *S. fuscum*, *S. magellanicum*, *S. angustifolium*, *S. nemoreum*, and *S. warnstorffii*. Feather mosses and lichens have high abundance, particularly *Hylocomium splendens*, *Pleurozium schreberi*, *Ptilium crista-castrensis*, *Cladina mitis*, *C. rangiferina*, and *C. chlorophaea*. Thirty percent of the total species are found in this group.

Group 10 consists of two *Sphagnum*-dominated stands that have high *Larix laricina* cover. These stands are restricted to site 7. In addition to *Larix*, *Salix discolor*, *S. serrissima*, and *Betula glandulosa*

are prevalent. Other important vascular plants are *Oxycoccus microcarpus*, *Smilacina trifolia*, *Eriophorum vaginatum*, *Carex aquatilis*, *C. paupercula*, *C. limosa*, *C. curta*, and *Menyanthes trifoliata*. *Sphagnum teres*, *S. warnstorffii*, and *S. angustifolium* are the important *Sphagnum* in this group. Species diversity is low at 11%.

Group 11 is a group of three *Sphagnum*-dominated stands that occur predominantly on the edge, and are most similar to stands of Group 13. Group 11 contains, in addition to *Sphagnum*, species that are normally restricted to much more minerotrophic conditions. These species are *Alnus tenuifolia*, *Betula pumila*, *Betula neoalaskana*, *Carex aquatilis*, *Calamagrostis canadensis*, *Bidens cernua*, *Lemna minor*, *L. trisculata*, *Lycopus uniflorus*, *Viola renifolia*, *Epilobium angustifolium*, *Mentha arvensis*, *Stellaria longifolia*, *Aster puniceus*, *Lysmachia thrysifolia*, and *Menyanthes trifoliata*. These stands also contain bryophytes that are indicative of minerotrophic conditions. These are *Climacium dendroides*, *Hypnum pratense*, *Drepanocladus exannulatus*, *D. polycarpus*, *Calliergon stramineum*, and *C. giganteum*,

Group 12 is comprised of dense *Picea glauca*, *P. mariana*, and *Larix laricina* stands that contain an ericad understory, *Sphagnum*, and feather mosses. Ericaceous shrubs consist of *Ledum groenlandicum*, and *Vaccinium vitis-idaea*. *Sphagnum* species of importance are *S. warnstorffii*, *S. centrale*, *S. angustifolium*, *S. magellanicum*, and *S. russowii*. Several species of bryophytes and vascular plants occur here that are less frequently associated with *Sphagnum*. These are the bryophytes *Plagiomnium ellipticum*, *Plagiothecium denticulatum*, *Pohlia nutans*, *Brachythecium mildeanum*, and *B. campestre*; and the vascular plants *Geum rivale*,

Galium boreale, *Equisetum fluviatile*, *Cornus canadensis*, *Cinna latifolia*, and *Impatiens capensis*. Thirty-two percent of the total species are present in this group.

Group 13 is a group of 5 *Sphagnum*-dominated stands that have a high cover of *Betula pumila*, *B. neoalaskana*, and *Picea mariana*, mixed with some *Larix laricina*, *Betula glandulosa*, *Salix pyrifolia*, and *Betula papyrifera*. Ericaceous shrubs are important components of this group, as well as *Eriophorum vaginatum* and *Rubus chamaemorus*. Minerotrophic indicators are present, including *Drepanocladus vernicosus*, *Aulacomnium palustre*, *Carex curta*, and *C. disperma*. Twenty-eight percent of the total species are present in this group.

Correlation of Vegetation to Water Chemistry and Environmental Factors

The stand classification produced by TWINSpan is significantly correlated with 8 water chemistry and 4 physical factors (Table 2). Relationships are present between water level, pH, specific conductivity, calcium, magnesium, sodium, total phosphorus, ammonium, and total organic nitrogen. Stands are also correlated with mean water level at the site, elevation of the site, number of streams entering and exiting the wetland, and hummock height within the stand. The strongest correlations are between vegetation and water chemistry (pH, corrected specific conductivity, calcium, magnesium, sodium), as well as mean water level and hummock height. Slightly lower correlations are obtained relating vegetation

Table 2

Relationship of Wetland Type to Water Chemistry and Environmental Factors

Wetland Type	Saline Deep Marsh			Shallow Marsh			Swamp			Mod. Rich Swamp			Poor Fen			Poor Fen			Bog			Corr. Prob.
	1	2	3	5	6	7	12	4	13	10	11	9	8	3.6	.0001	8	3.5	.0001	27	16	4	.0001
Stand Groups	5.2	6.4	6.4	6.1	6.2	5.6	5.9	6.2	4.5	4	4.4	3.5	3.6	.0001								
pH	899	534	478	158	326	320	233	162	62	59	42	27	16	.0001								
Corr. Spec. Cond. μ S	45	65	41	27	26	43	41	33	7	2	6	7	4	.0001								
Calcium mg/l	7	18	10	9	9	12	14	10	3	1	2	2	1	.0001								
Magnesium mg/l	790	125	31	3	14	22	5	5	3	4	2	2	3	.0001								
Sodium mg/l	6384	2166	2527	2013	2977	2628	2013	1577	2645	1362	2843	2918	2979	.0134								
Organic N μ g/l	64	175	14	9	10	8	7	8	23	14	16	20	13	.3353								
Nitrate μ g/l	250	82	133	73	28	62	146	23	82	23	36	162	251	.0346								
Ammonium μ g/l	1110	252	520	248	650	609	221	135	411	195	374	352	482	.0484								
Total Phosphorus μ g/l	0	21	17	17	7	6	4	10	1	4	6	-9	-15	.0001								
Mean Water Level (cm)	22	9	12	15	16	16	20	12	24	15	14	31	26	.0001								
Hummock Height (cm)	2400	2370	2364	2388	2368	2391	2383	2377	2380	2375	2400	2400	2378	.0220								
Elevation (Ft.)	3	2.7	2.5	2.2	2.3	2.4	0.7	2.2	2	3	1.7	2	1	.0048								
# Streams	1.5	1.4	1.1	2.2	1.2	1.3	2.3	3.7	2.4	2.4	1.7	2.5	2.7	.4399								
Peat Depth (M)	1	10	22	30	29	10	3	24	5	2	3	6	10									
Number of stands																						

Corr. Prob. = Spearman Rank Correlation Probability

Corr. Spec. Cond. = Corrected Specific Conductivity

stands to nitrogen and phosphorus levels within the wetland, elevation, and the number of inlet and outlet streams.

Integration of the thirteen TWINSpan vegetation stands into the wetland classification based on the National Wetlands Working Group (1988) results in seven wetland types: saline marsh, deep marsh, shallow marsh, swamp, moderate-rich fen, poor fen and bog (Table 2). The placement of each stand group into a wetland type is based on vegetation definitions of the wetland types and vegetational similarities amongst the stand groups.

The pH is slightly acidic to circumneutral across marsh, swamp, and moderate-rich fen wetland types (Table 2), whereas it declines to about 4.5 in poor fen and 3.5 in bog groups. Similarly, corrected specific conductivity is highest in marsh wetlands (158-899 μS), intermediate in swamp wetlands (233-326 μS), and declines significantly in the moderate-rich fen (162 μS), and poor fens (42-62 μS). Corrected specific conductivity is lowest in the bog wetlands at 16-27 μS . This gradient is repeated for calcium, magnesium, and sodium surface water cation concentrations. Marshes have 27-65 mg/l Ca, 7-18 mg/l Mg, 3-790 mg/l Na; swamps have 26-43 mg/l Ca, 9-14 mg/l Mg, 5-22, g/l Na; moderate-rich fens have 33 mg/l Ca, 10 mg/l Mg, 5 mg/l Na; poor fens have 2-7 mg/l Ca, 1-3 mg/l Mg, 2-4 mg/l Na; and bogs have 4-7 mg/l Ca, 1-2 mg/l Mg, 2-3 mg/l Na.

Nutrient concentrations are not as strongly correlated to stand groups as are cation concentrations and do not follow the ionic gradient outlined in Table 2. Organic nitrogen is highest in the saline marsh at 6384 $\mu\text{g/l}$. Most of the other wetland types have between 2013-2979 $\mu\text{g/l}$ of organic N. Two fen groups, moderate-rich fen

Group 4 and poor fen Group 10, have comparatively low organic N at 1577 and 1362 $\mu\text{g/l}$.

Surface water nitrate concentrations are highest in deep marsh wetlands (Stand Group 2) at 175 $\mu\text{g/l}$, followed by saline marshes (64 $\mu\text{g/l}$). The remaining stand groups contain between 6.6-19.9 $\mu\text{g/l NO}_3$.

Ammonium concentrations are highest in the saline marsh and bog Group 8 at 250 $\mu\text{g/l}$. Intermediate values of 133-162 $\mu\text{g/l}$ were recorded in shallow marsh Group 3, swamp Group 12, and bog Group 9. Lowest values (23-36 $\mu\text{g/l}$) were found in swamp Group 6, moderate-rich fen Group 4, poor fen Group 10, and poor fen Group 11.

Total phosphorus is highest in saline marsh at 1110 $\mu\text{g/l}$. Other relatively high values are recorded in swamp Groups 6 (650 $\mu\text{g/l}$), and 7 (609 $\mu\text{g/l}$). Lowest total phosphorus concentrations are found in the moderate-rich fen Group 4 (135 $\mu\text{g/l}$), and poor fen Group 10 (195 $\mu\text{g/l}$).

Mean water levels are below the surface vegetation only in the bog wetland types as indicated by negative water levels, and at the surface in the saline marshes. Fens and swamps have relatively shallow water table depths, 1.3-9.5 cm and 3.5-7.2 cm respectively. Shallow marsh vegetation is on average covered by 17 cm of water, while the deep marshes were covered by 21 cm of water. Hummock height tends to be lowest in marsh wetlands (9-22 cm), intermediate in fens and swamps (12-23 cm), and highest in bogs (26-31 cm). A positive correlation with site elevation shows a tendency for marshes to be located at lower elevations, swamps and fens at intermediate elevations, and bogs at the highest elevations. Similarly, bogs (and

one of the swamps) have the fewest number of streams entering and exiting the wetlands, while fens, other swamps and marshes have more.

Evolutionary Development

All radiocarbon dates are given in Table 3. The oldest basal radiocarbon date is 6720 $\pm$ 120 yBP. There is no relationship to the amount of peat material accumulated and the age of the deposit nor to the age of the deposit and the type of wetland. Within a given site however, successively younger dates appear towards the peatland edges, in shallower peats. At each site, accumulation rates are highest in the central part of the peatland. Highest accumulation rates occur in the moderate-rich fen sites, followed by the bog sites 27 and 22. Lowest accumulation rates occur in shallow marsh site 35.

Site 2, Core 2-1

The basal radiocarbon date for this peat profile is 5620 $\pm$ 130 yBP at 420 cm (Fig. 3). Basal sediments contain seeds of many aquatic plants such as *Typha* sp., *Scirpus* sp., and *Zannichella palustris*. Oospores of *Chara* sp., the ehippia of cladocera, and the shells of various ostracods and gastropods are also found. Farther up the peat profile (350 cm) bryophytes become prevalent. Leaves of *Drepanocladus crassicosatus*, *D. aduncus* var. *kneiffii*, *D. exannulatus*, *D. aduncus*, *D. revolvens*, *D. vernicosus*, and *D. polycarpus*, as well as *Meesia triquetra*, *Calliergon trifarum*, *C. stramineum*, and *C.*

Table 3
ELK ISLAND RADIOCARBON DATES - Relative to 1950

Wetland Type	Site	Position in Transect	Depth (cm)	Lab. No.	C-14 Age (BP)	Depth Dated	Peat Type Dated	Accumulation Rate
Bog	27	Center	190-195	1020C	2620+/-130	0-195	Poor fen/fen bound.	.73 mm/yr.
	27	Center	295-300	817C	4670+/-200	196-300	Aquatic/limnic.	.49 mm/yr.
	27	Midway	164-169	819C	3510+/-100	0-169	Aquatic/limnic.	.47 mm/yr.
	27	Edge	100-104	820C	1690+/-100	0-104	Fen	.60 mm/yr.
Bog	22	Center	201-211	1022C	4080+/-160	0-211	Aquatic/limnic	.50 mm/yr.
	22	Midway	171-180	1021C	4910+/-110	0-180	Aquatic/limnic	.36 mm/yr.
Mod.-rich Fen	2	Center	415-420	726C	5620+/-130	0-420	Aquatic/limnic	.74 mm/yr.
	2	Midway	545-550	812C	6720+/-120	0-550	Aquatic/limnic	.81 mm/yr.
	2	Edge	477-480	813C	5740+/-240	0-480	Aquatic/limnic	.83 mm/yr.
Mod.-rich Fen	3	Center	460-465	1025C	5560+/-110	0-465	Aquatic/limnic	.83 mm/yr.
	3	Midway	290-295	815C	4780+/-140	0-295	Aquatic/limnic	.61 mm/yr.
	3	Edge	165-170	816C	3270+/-100	0-170	Aquatic/limnic	.51 mm/yr.
Shallow Marsh	35	Center	135-140	821C	4250+/-260	0-140	Aquatic/limnic	.32 mm/yr.
	35	Edge	117-125	1019C	2880+/-140	0-125	Aquatic/limnic	.42 mm/yr.
Swamp	48	Center	220-225	823C	4990+/-230	0-225	Aquatic/limnic	.45 mm/yr.
	48	Midway	205-210	1024C	5310+/-90	0-210	Aquatic/limnic	.39 mm/yr.
	48	Edge	56-70	825C	2910+/-240	0-70	Aquatic/limnic	.20 mm/yr.

Table 3 outlines the location of the peat cores, the depth and type of peat material dated, and the relative accumulation rate.
BP = Years before 1950, Lab. No. = Laboratory Number, Mod.-rich = Moderate-rich, Bound. = Boundary.

Figure 3.

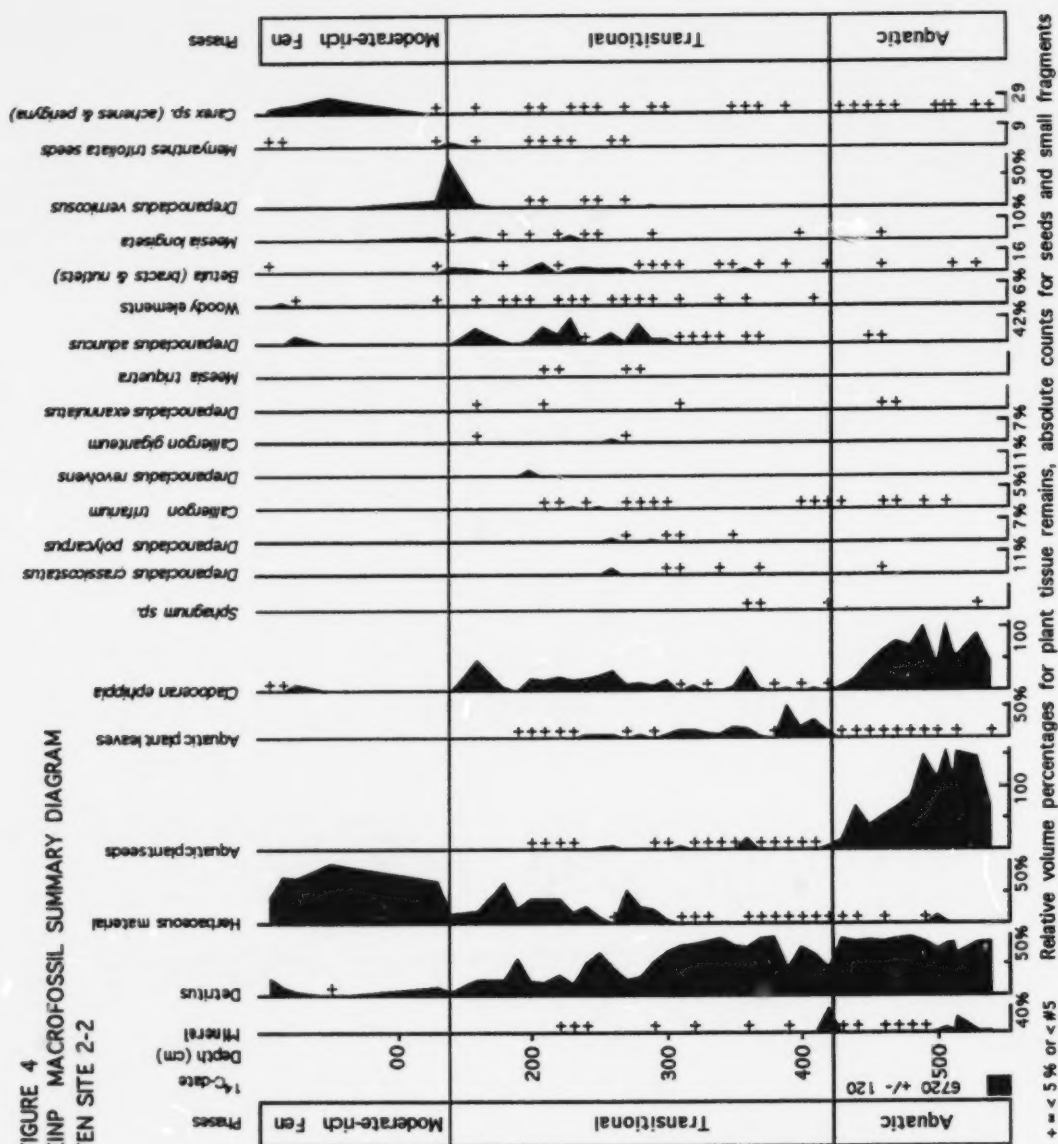
Elk Island National Park Macrofossil Profile, Fen Site 2-1. Located in back pocket.

giganteum are present. The remains of aquatic plants (*Potamogeton*, *Ceratophyllum*), are still prevalent in these strata and seeds of the aquatic plants *Ranunculus sceleratus*, and *Najas flexilis* are common. Seeds of semi-terrestrial plants (*Menyanthes trifoliata*, *Triglochin maritima*, *Carex* sp.) appear at 250 cm. A distinct change occurs in the peat profile at 150 cm. Detritus values decline, corresponding to an increase in herbaceous roots, herbaceous material, *Meesia triquetra*, *M. longiseta*, *Calliergon trifarum*, *C. stramineum*, *C. giganteum*, *Carex limosa/paupercula*, and *Carex* sp. achenes.

Core 2-2

Basal sediments of core 2-2 at 550 cm, are radiocarbon dated at 6720 +/- 120 yBP (Fig. 4). Seeds of the aquatic plants (*Chenopodiaceae*, *Scirpus* sp., *Zannichella palustris*, *Typha* sp., *Potamogeton* sp., *Ceratophyllum demersum*, *Ranunculus sceleratus*) are the predominant macrofossils in the basal peat sediments (Appendix I). Cladoceran ephippia reach the highest proportions as well as ostracod and gastropod shells. Above 420 cm, aquatic seeds and cladoceran ephippia numbers decline abruptly and fossil remains of *Drepanocladus crassicostratus*, *D. polycarpus*, *D. revolvens*, *D. exannulatus*, *D. aduncus*, *D. vernicosus*, *Calliergon trifarium*, *C. giganteum*, and *Meesia triquetra* appear. At 130 cm herbaceous material, herbaceous roots, and the number of *Carex* seeds sharply increase, corresponding to a decline in detritus and cladoceran ephippia.

FIGURE 4
EINP MACROFOSSIL SUMMARY DIAGRAM
FEN SITE 2-2



Core 2-3

The radiocarbon date at 480 cm is 5740 +/- 240 yBP (Fig. 5).

Sediments at this depth are dominated by aquatic plant leaves (*Myriophyllum* sp., *Ceratophyllum demersum*, *Potamogeton* sp., aquatic plant seeds (*Zannichella palustris*, *Ceratophyllum demersum*, *Scirpus* sp., *Ranunculus sceleratus*, *Typha* sp., *Myriophyllum* sp., *Rumex* sp.), *Chara* oospores, cladoceran ephippia, and ostracod shells (Appendix I). Above 225 cm, leaves of the bryophytes *Drepanocladus crassicosatus*, *D. aduncus*, *Meesia triquetra* and *M. longiseta* appear. Seeds of the semi-terrestrial *Menyanthes trifoliata* also appear and there is an increase in the amount of *Betula* and woody elements. Above 115 cm the amount of herbaceous material increases substantially while aquatic plants and cladoceran ephippia decline. In these strata the remains of *Drepanocladus vernicosus*, *D. revolvens* and *Rhizomnium pseudopunctatum* can be found.

Site 3, Core 3-1

Basal peats at core 3-1 are 465 cm deep, 5560 years old, and contain the remains of aquatic seeds (*Scirpus* sp., *Zannichella palustris*, *Ceratophyllum demersum*, *Ranunculus sceleratus*), ostracod and gastropod shells, *Chara* sp. oospores, and cladoceran ephippia (Fig. 6). Above 375 cm, leaves and seeds of the aquatic plants *Myriophyllum* sp., *Ceratophyllum demersum*, and *Potamogeton* become prevalent, as well as the bryophytes *Drepanocladus aduncus*, *D. aduncus* var. *kneiffii*, *D. crassicosatus*, *Calliergonella cuspidata*, and *Sphagnum* sp. Seeds of semi-terrestrial plants such as *Menyanthes trifoliata*, *Carex*

FIGURE 5
EINP MACROFOSSIL SUMMARY DIAGRAMS
FEN SITE 2-3

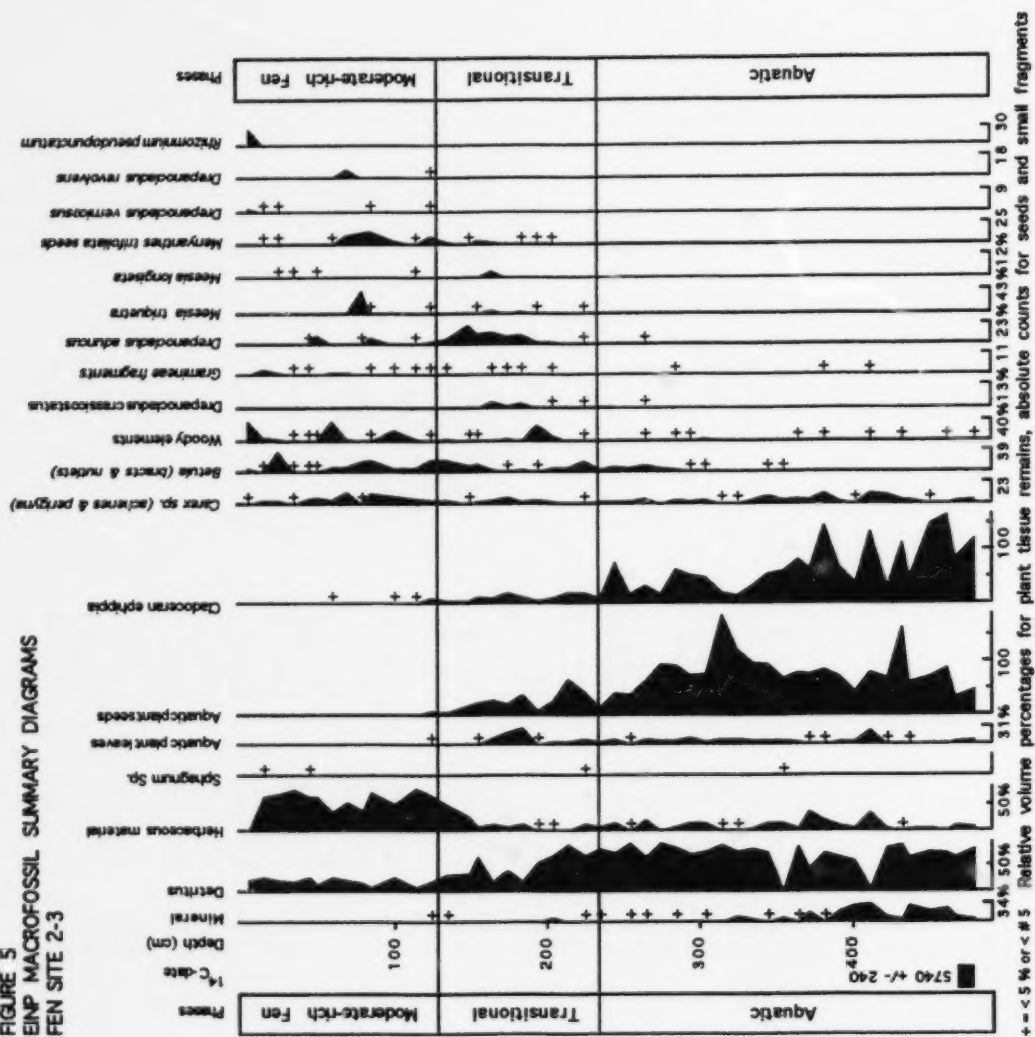


Figure 6.

Elk Island National Park Macrofossil Profile, Fen Site 3-1. Located in back pocket.

sp., and the terrestrial species *Betula* are also present. Incorporation of mineral grains into the peat sediments decreases above 80 cm and in this region, herbaceous material becomes increasingly important, while aquatic plants and cladoceran ephippia decline. Above 40 cm, *Calliergon stramineum*, *Aulacomnium palustre* and *Drepanocladus vernicosus* codominate in the peat assemblage along with herbaceous material and *Carex* sp. achenes.

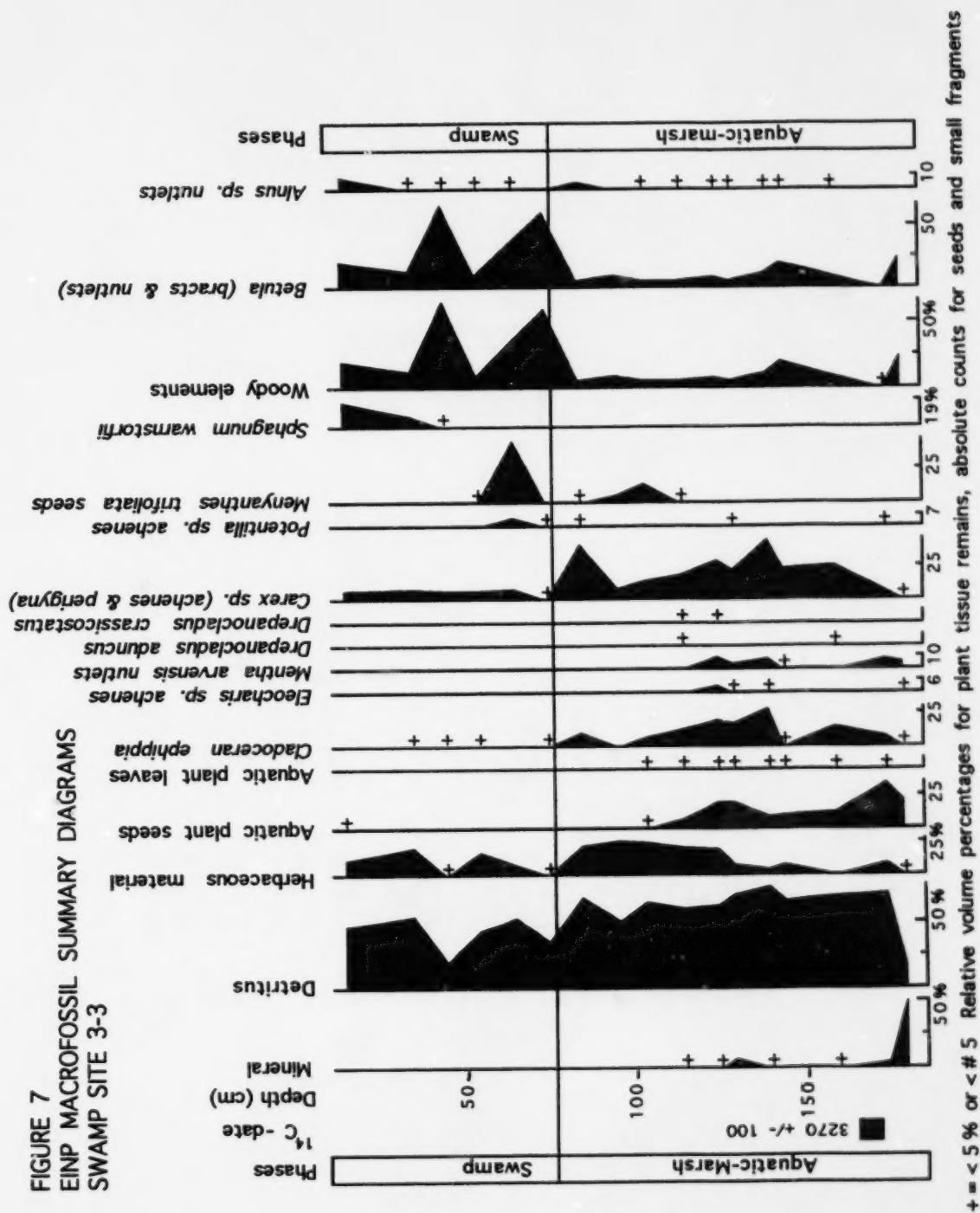
Core 3-3

At 170 cm, the basal peats of core 3-3 are 3270 +/- 100 years old. Dominating these peats are aquatic plant seeds (*Typha* sp., *Myriophyllum* sp., *Ceratophyllum demersum*, *Ranunculus sceleratus*, *Najas flexilis* (Appendix I), and cladoceran ephippia (Fig. 7). In addition to the aquatic plants, there are propagules and remains of terrestrial and semi-terrestrial plants (*Mentha arvensis*, *Potentilla*, *Carex*, *Betula*, and *Alnus*). Herbaceous material is more abundant than in the basal sediments of core 3-1. Bryophyte remains include *Drepanocladus aduncus* and *D. crassicostatus*. Above 96 cm, aquatic seeds and cladoceran ephippia decline, while semi-terrestrial and terrestrial elements rise (*Carex* sp., *Potentilla* sp., *Sphagnum warnstorffii*). Particularly dominant in these sediments are materials of a woody nature (woody elements, *Betula* bracts and nutlets, and *Alnus* nutlets).

Site 22, Core 22-1

Four thousand and eighty years before present, aquatic plants and cladoceran ephippia began to accumulate in the basin of site 22 (Fig.

FIGURE 7
EINP MACROFOSSIL SUMMARY DIAGRAMS
SWAMP SITE 3-3



8). Leaves and microscopic remains of *Carex* sp., *Betula* sp., *Sphagnum warnstorffii*, *S. centrale*, *S. capillifolium*, *S. russowii*, and *Polytrichum strictum* were deposited. Above 129 cm the remains of aquatic plants decline and semi-terrestrial plants (*Carex* sp., *Betula* sp., *Sphagnum* spp.) dominate. Recent peat sediments are dominated by *Sphagnum capillifolium*, woody elements, *S. russowii*, and *Polytrichum strictum*.

Core 22-2

Peat depth at core 22-2 is shallower, but older (4910 yBP) than core 22-1. Mineral grains comprise the greatest portion of the basal sediments (Fig. 9). Farther up the peat profile, detritus and aquatic plants dominate with fragments of *Drepanocladus aduncus*, *Sphagnum magellanicum*, and *S. russowii*. A transitional aquatic/terrestrial phase occurs above 90 cm. Aquatic seeds and cladoceran ephippia are still present in the sediments, but herbaceous material has increased and seeds of the terrestrial herbs (*Sium suave*, *Potentilla* sp., *Carex* sp., *Betula* sp.) occur. This section is followed by sediments dominated by *Sphagnum*, *Ledum groenlandicum*, *Carex* sp., woody elements, and *Betula* sp. Near the surface, the peat community abruptly changes to one dominated by herbaceous material, *Polytrichum strictum*, and *Calliergon stramineum*.

Core 22-3

The earliest peat forming community at core 22-3 was aquatic in nature, formed by aquatic plants, herbs and Cladocera (Fig. 10).

Figure 8.

Elk Island National Park Macrofossil Profile, Bog Site 22-1. Located in back pocket.

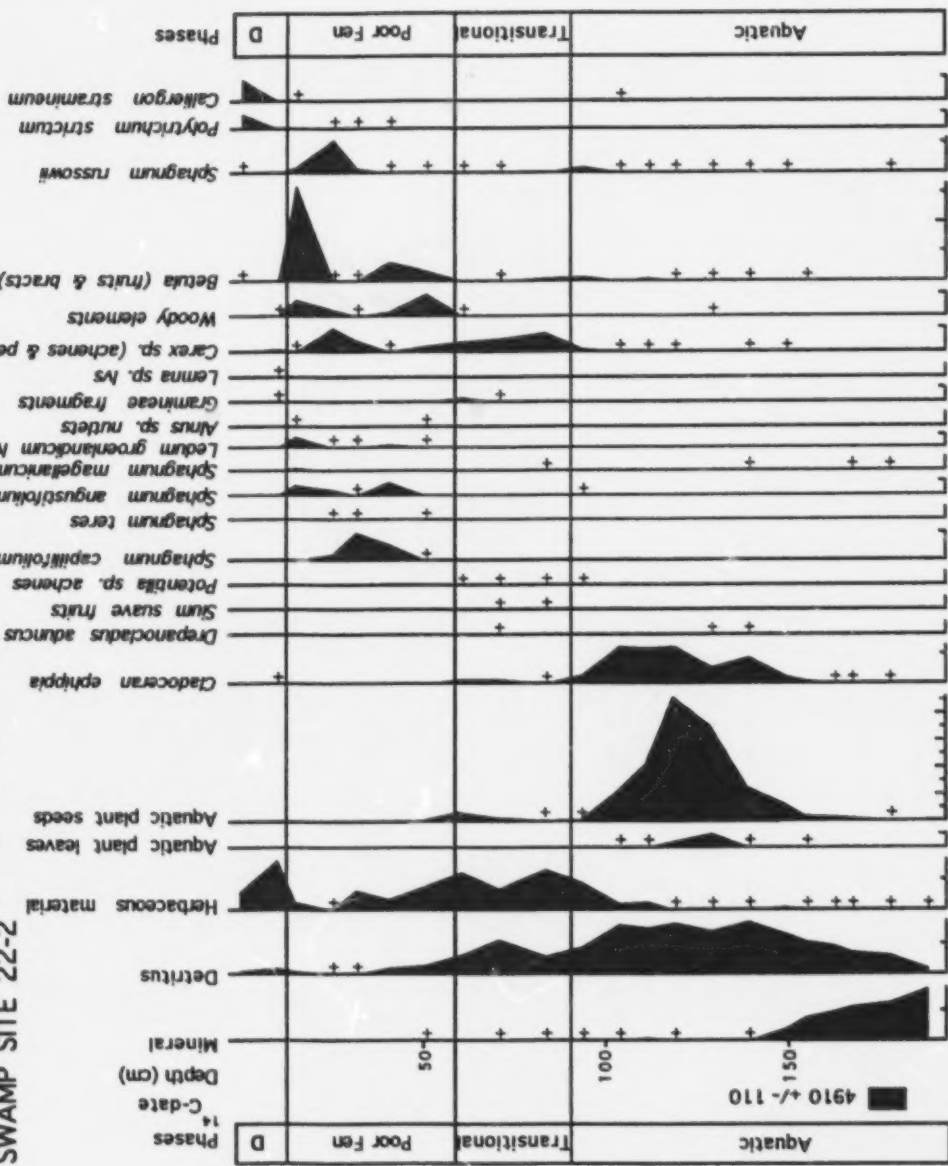
FIGURE 9
EINP MACROFOSSIL SUMMARY DIAGRAMS
SWAMP SITE 22-2

Phases: Aquatic, Transitional, Poor Fen, D

Depth (cm): 0, 50, 100, 150

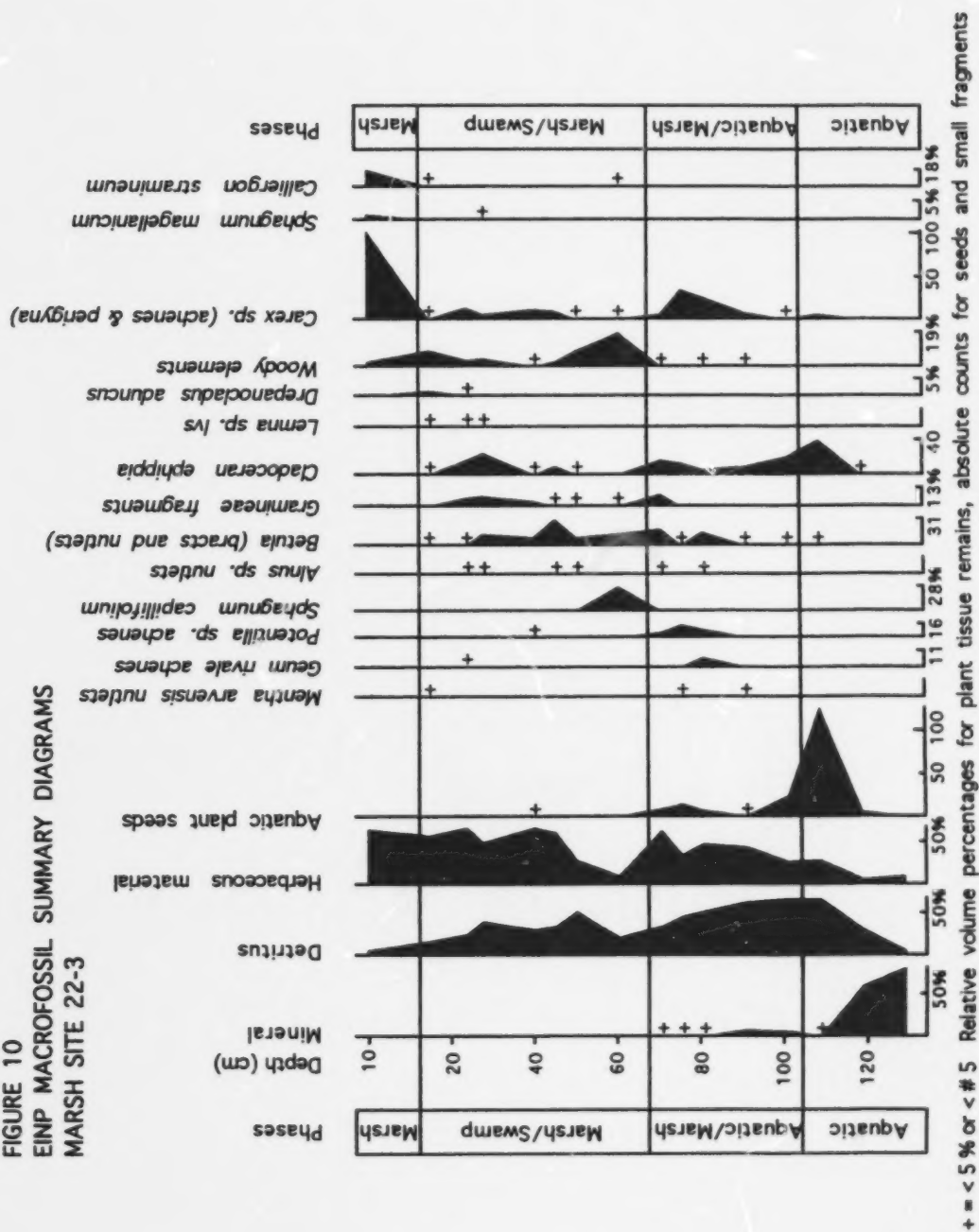
Plant Tissue Remains (Absolute Counts for Seeds and Small Fragments):

- Aquatic plant seeds: 0, 50, 100, 150, 200, 250, 300, 350, 400, 450, 500, 550, 600, 650, 700, 750, 800, 850, 900, 950, 1000, 1050, 1100, 1150, 1200, 1250, 1300, 1350, 1400, 1450, 1500, 1550, 1600, 1650, 1700, 1750, 1800, 1850, 1900, 1950, 2000, 2050, 2100, 2150, 2200, 2250, 2300, 2350, 2400, 2450, 2500, 2550, 2600, 2650, 2700, 2750, 2800, 2850, 2900, 2950, 3000, 3050, 3100, 3150, 3200, 3250, 3300, 3350, 3400, 3450, 3500, 3550, 3600, 3650, 3700, 3750, 3800, 3850, 3900, 3950, 4000, 4050, 4100, 4150, 4200, 4250, 4300, 4350, 4400, 4450, 4500, 4550, 4600, 4650, 4700, 4750, 4800, 4850, 4900, 4950, 5000, 5050, 5100, 5150, 5200, 5250, 5300, 5350, 5400, 5450, 5500, 5550, 5600, 5650, 5700, 5750, 5800, 5850, 5900, 5950, 6000, 6050, 6100, 6150, 6200, 6250, 6300, 6350, 6400, 6450, 6500, 6550, 6600, 6650, 6700, 6750, 6800, 6850, 6900, 6950, 7000, 7050, 7100, 7150, 7200, 7250, 7300, 7350, 7400, 7450, 7500, 7550, 7600, 7650, 7700, 7750, 7800, 7850, 7900, 7950, 8000, 8050, 8100, 8150, 8200, 8250, 8300, 8350, 8400, 8450, 8500, 8550, 8600, 8650, 8700, 8750, 8800, 8850, 8900, 8950, 9000, 9050, 9100, 9150, 9200, 9250, 9300, 9350, 9400, 9450, 9500, 9550, 9600, 9650, 9700, 9750, 9800, 9850, 9900, 9950, 10000
- Aquatic plant leaves: 0, 50, 100, 150, 200, 250, 300, 350, 400, 450, 500, 550, 600, 650, 700, 750, 800, 850, 900, 950, 1000, 1050, 1100, 1150, 1200, 1250, 1300, 1350, 1400, 1450, 1500, 1550, 1600, 1650, 1700, 1750, 1800, 1850, 1900, 1950, 2000, 2050, 2100, 2150, 2200, 2250, 2300, 2350, 2400, 2450, 2500, 2550, 2600, 2650, 2700, 2750, 2800, 2850, 2900, 2950, 3000, 3050, 3100, 3150, 3200, 3250, 3300, 3350, 3400, 3450, 3500, 3550, 3600, 3650, 3700, 3750, 3800, 3850, 3900, 3950, 4000, 4050, 4100, 4150, 4200, 4250, 4300, 4350, 4400, 4450, 4500, 4550, 4600, 4650, 4700, 4750, 4800, 4850, 4900, 4950, 5000, 5050, 5100, 5150, 5200, 5250, 5300, 5350, 5400, 5450, 5500, 5550, 5600, 5650, 5700, 5750, 5800, 5850, 5900, 5950, 6000, 6050, 6100, 6150, 6200, 6250, 6300, 6350, 6400, 6450, 6500, 6550, 6600, 6650, 6700, 6750, 6800, 6850, 6900, 6950, 7000, 7050, 7100, 7150, 7200, 7250, 7300, 7350, 7400, 7450, 7500, 7550, 7600, 7650, 7700, 7750, 7800, 7850, 7900, 7950, 8000, 8050, 8100, 8150, 8200, 8250, 8300, 8350, 8400, 8450, 8500, 8550, 8600, 8650, 8700, 8750, 8800, 8850, 8900, 8950, 9000, 9050, 9100, 9150, 9200, 9250, 9300, 9350, 9400, 9450, 9500, 9550, 9600, 9650, 9700, 9750, 9800, 9850, 9900, 9950, 10000
- Herbaceous material: 0, 50, 100, 150, 200, 250, 300, 350, 400, 450, 500, 550, 600, 650, 700, 750, 800, 850, 900, 950, 1000, 1050, 1100, 1150, 1200, 1250, 1300, 1350, 1400, 1450, 1500, 1550, 1600, 1650, 1700, 1750, 1800, 1850, 1900, 1950, 2000, 2050, 2100, 2150, 2200, 2250, 2300, 2350, 2400, 2450, 2500, 2550, 2600, 2650, 2700, 2750, 2800, 2850, 2900, 2950, 3000, 3050, 3100, 3150, 3200, 3250, 3300, 3350, 3400, 3450, 3500, 3550, 3600, 3650, 3700, 3750, 3800, 3850, 3900, 3950, 4000, 4050, 4100, 4150, 4200, 4250, 4300, 4350, 4400, 4450, 4500, 4550, 4600, 4650, 4700, 4750, 4800, 4850, 4900, 4950, 5000, 5050, 5100, 5150, 5200, 5250, 5300, 5350, 5400, 5450, 5500, 5550, 5600, 5650, 5700, 5750, 5800, 5850, 5900, 5950, 6000, 6050, 6100, 6150, 6200, 6250, 6300, 6350, 6400, 6450, 6500, 6550, 6600, 6650, 6700, 6750, 6800, 6850, 6900, 6950, 7000, 7050, 7100, 7150, 7200, 7250, 7300, 7350, 7400, 7450, 7500, 7550, 7600, 7650, 7700, 7750, 7800, 7850, 7900, 7950, 8000, 8050, 8100, 8150, 8200, 8250, 8300, 8350, 8400, 8450, 8500, 8550, 8600, 8650, 8700, 8750, 8800, 8850, 8900, 8950, 9000, 9050, 9100, 9150, 9200, 9250, 9300, 9350, 9400, 9450, 9500, 9550, 9600, 9650, 9700, 9750, 9800, 9850, 9900, 9950, 10000
- Detritus: 0, 50, 100, 150, 200, 250, 300, 350, 400, 450, 500, 550, 600, 650, 700, 750, 800, 850, 900, 950, 1000, 1050, 1100, 1150, 1200, 1250, 1300, 1350, 1400, 1450, 1500, 1550, 1600, 1650, 1700, 1750, 1800, 1850, 1900, 1950, 2000, 2050, 2100, 2150, 2200, 2250, 2300, 2350, 2400, 2450, 2500, 2550,



	50%	50%	50%	27%	24%	19%	8%	41	38%	100	50%	25%	39%
D=	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed	Disturbed
+ =	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%	< 5%
	Relative volume percentages for plant tissue remains, absolute counts for seeds and small fragments												

FIGURE 10
EINP MACROFOSSIL SUMMARY DIAGRAMS
MARSH SITE 22-3



Mineral grains are frequently mixed with the peat sediments. Farther up the peat profile, herb percentages increase, along with terrestrial species such as *Mentha arvensis*, *Geum rivale*, *Potentilla*, *Alnus* sp., *Betula* sp, Gramineae, woody elements, and *Carex* sp. In strata above 70 cm, aquatic plants decline, while *Sphagnum capillifolium*, *Drepanocladus aduncus*, *Carex* sp., and *Betula* sp. increase. In the top 20 cm of peat, *Carex* sp., and herbaceous material dominate along with *Sphagnum magellanicum* and *Calliergon stramineum*.

Site 27, Core 27-1

The aquatic macrofossils that comprise the basal peats at core 27-1 are 4670 +/- 200 years old, and are slightly different from the previous sites in that only a few aquatic species are present and in small amounts (Fig. 11). These consist predominantly of the leaves and seeds of *Myriophyllum* sp., *Ceratophyllum demersum*, *Potamogeton* sp., *Typha* sp., *Najas flexilis*, *Ranunculus sceleratus*, and *R. circinatus*. A high number of semi-terrestrial species are also present in these basal sediments (*Drepanocladus aduncus* var. *kneiffii*, *D. exannulatus*, *Calliergon stramineum*, *Sphagnum* section *Cuspidata*, *S. magellanicum*, *Caltha palustris*, *Menyanthes*, *Carex*, and *Betula*). Above 240 cm there is a dramatic increase in the quantity of herbaceous roots, and herbaceous material as well as an increase in the quantity of bryophyte remains (*Drepanocladus aduncus*, *D. vernicosus*, *Calliergon stramineum*, and *Tomenthypnum nitens*), and *Betula* nutlets. This is followed by a lengthy section of peat that is dominated by members of the genus *Sphagnum* (*S. magellanicum*, *S.*

Figure 11.

Elk Island National Park Macrofossil Profile, Bog Site 27-1. Located in back pocket.

angustifolium, *S. warnstorffii*, *S. capillifolium*), ectomycorrhizal roots, *Picea* needles and fungal mycelia. Surficial peats are comprised of only a few *Sphagnum* species (*S. magellanicum*, *S. fuscum*) and *Polytrichum strictum*.

Core 27-4

The basal sediments of core 27-4 are the youngest found in this study (1690 $\pm$ 100 yBP, Fig. 12). Contrary to all previous cores, this core does not contain aquatic plants, or seeds of aquatic plants, in the basal sediments. The initial peat forming community is dominated by herbs, *Carex* sp., *Betula* sp., woody elements, and small amounts of *Sphagnum teres*, *S. angustifolium* and *S. fuscum*. *Sphagnum* dominance occurs above 72 cm; predominant are the species *Sphagnum fuscum*, *S. magellanicum*, and *S. angustifolium*.

Site 35, Core 35-1

Sediments at core 35-1 are shallow, and date to 4250 $\pm$ 260 yBP (Fig. 13). Mineral grains are prominent in these basal sediments. Remains of aquatic plants are present, particularly those of *Chara*, and *Typha* sp., along with cladoceran epnippia. Above 59 cm, the remains of aquatic plants disappear and herbaceous material, herbaceous roots, *Betula* sp., and *Carex* sp., increase. No bryophytes are present in this upper phase and all categories describing woody elements remain low, indicating a herbaceous/*Carex* community.

FIGURE 12



Figure 13.

Elk Island National Park Macrofossil Profile, Marsh Site 35-1. Located in back pocket.

Core 35-2

The peat profile at core 35-2 is younger (2880 $\pm$ 140) and shallower (136 cm) than core 35-1 (Fig. 14). Once again aquatic sediments are present at the base. These consist predominantly of *Chara* oospores, *Potamogeton* sp., *Typha* sp., *Eleocharis* sp., and *Ranunculus* sp. seeds (Appendix I). Incorporated into the peat is a high percentage of mineral grains, and the seeds of *Potentilla* sp., *Mentha arvensis*, *Betula* sp., and *Carex* sp. Gradually the aquatic species decline and herbaceous material dominates with Gramineae fragments and *Carex* sp. achenes.

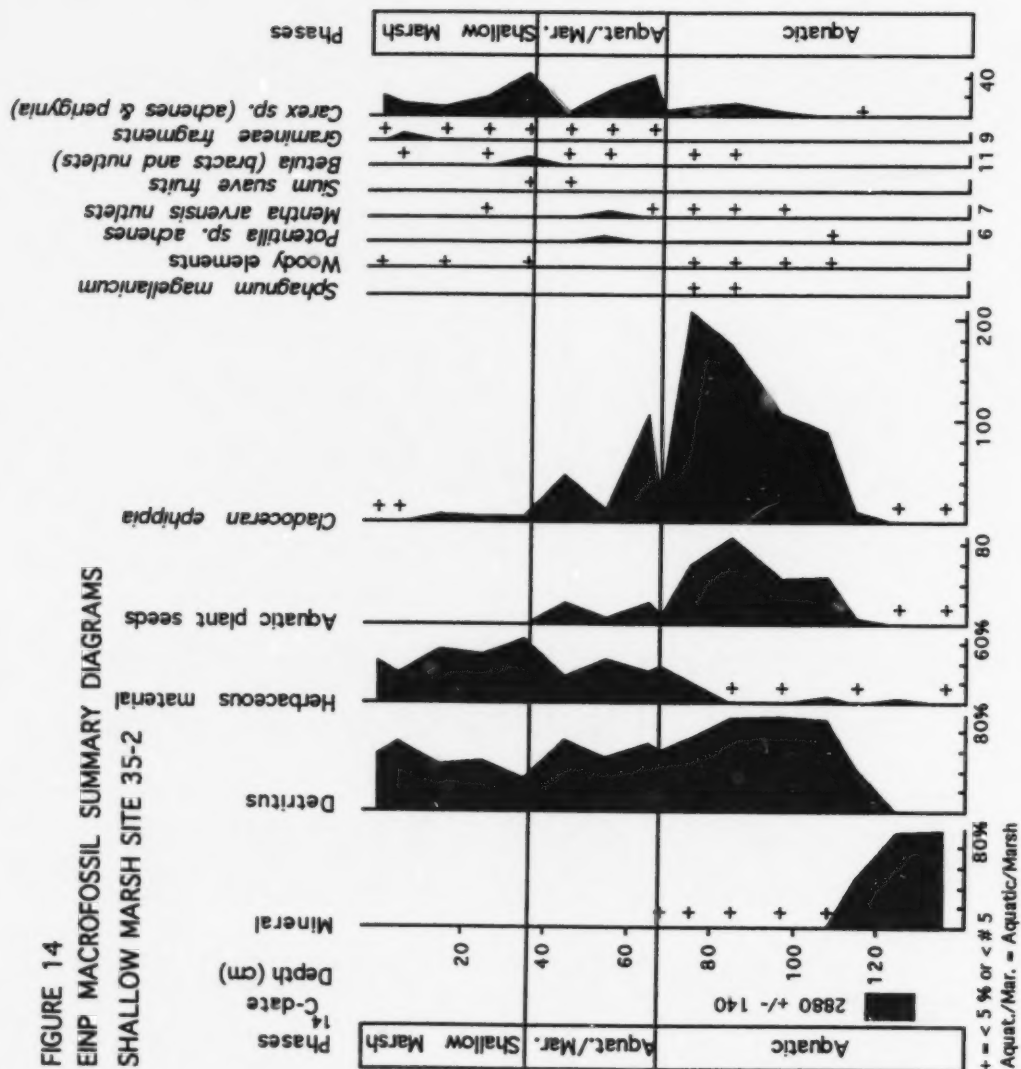
Core 35-3

Core 35-3 is shallow with just 45 cm of peat (Fig. 15). Again there are mineral grains and aquatic seeds, especially *Typha* sp. present in the basal portion (Appendix I). Charcoal fragments are abundant in this core. Herbaceous material and Gramineae fragments increase towards the surface.

Site 48, Core 48-1

Peat began to accumulate in the shallow bay of a lake at this site around 4990 $\pm$ 230 yBP. (Fig. 16). Mineral grains are frequently mixed with the peaty substrate. Leaves and seeds of the aquatic plants *Myriophyllum* sp., *Ceratophyllum demersum*, *Ruppia maritima*, *Typha* sp., *Scirpus* sp., and the oospores of *Chara* and cladoceran ephippia are found in these layers. Above 150 cm, aquatic plants and cladoceran ephippia decline, while herbaceous roots,

FIGURE 14
EINP MACROFOSSIL SUMMARY DIAGRAMS
SHALLOW MARSH SITE 35-2



Species	% fragments
<i>Carex</i> sp. (achenes & perigynia)	~10%
<i>Alnus</i> sp. nutlets	~5%
<i>Cladoceran</i> ephippia	~15%



Figure 16.

Elk Island National Park Macrofossil Profile, Marsh Site 48-1. Located
in back pocket.

Ranunculus sceleratus and *Carex* sp. achenes increase. In the third phase, above 25 cm, an abrupt increase in *Larix* and *Picea* needles, deciduous leaves and ectomycorrhizal roots occurs. This is followed by a return to sediments dominated by herbaceous components.

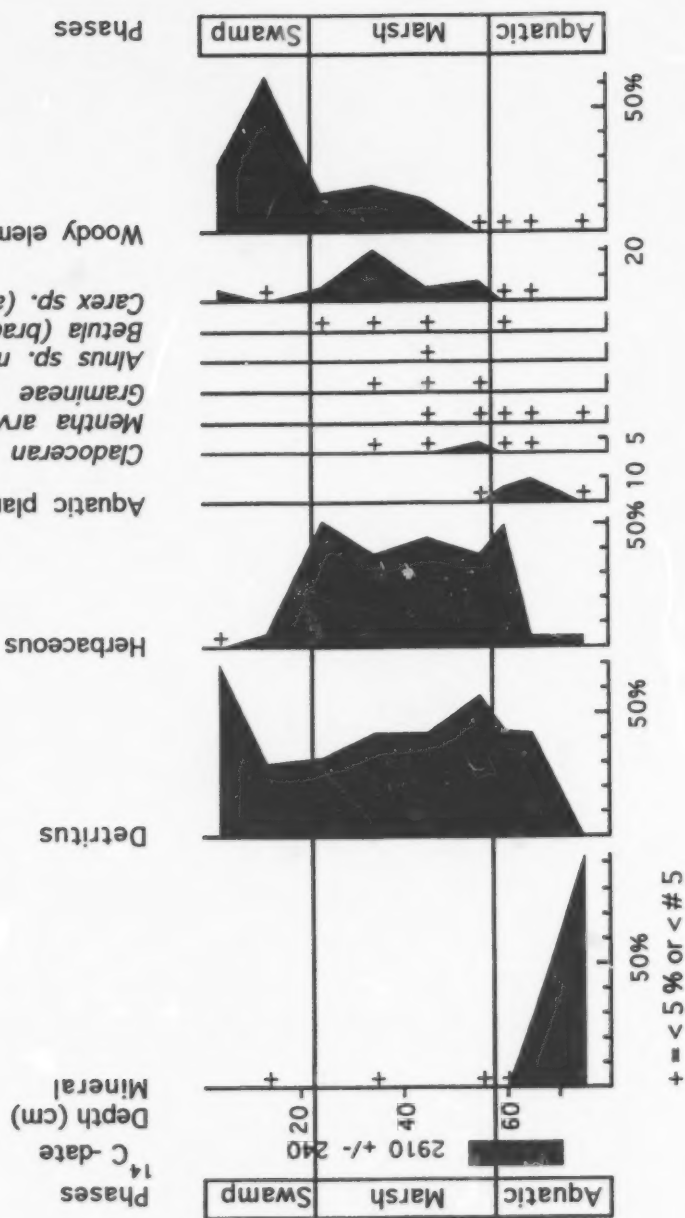
Core 48-3

Peat began to accumulate in this portion of the basin much later (2910 +/- 240 yBP) than cores 48-1 and 48-2 (Fig. 17). Mineral grains were incorporated into these initial sediments. Few aquatic macrofossils are present, consisting of the achenes of *Typha* sp., *Juncus* sp., and *Ranunculus sceleratus*. (Appendix I). Fossils of terrestrial plants (*Mentha arvensis*, *Betula*, woody elements) are present as well. Farther up the peat profile, herbaceous material, *Carex* achenes, and woody elements all increase. The most recent peat sediments are dominated by woody elements, with minor amounts of *Carex* achenes.

Peat Chemistry

Peat samples were assigned to one of eight peat types based on the peat phase identified in the macrofossil analysis and subjected to a canonical correlation and discriminant analysis. Peat type categories are based upon the macrofossil analysis: **aquatic-** limnic peat containing seeds and fossils of aquatic plants; **transitional-** limnic peat containing the remains of aquatic and few plants; **marsh-** telmatic peat containing herbaceous plant remains; **swamp-** telmatic peat containing herbaceous and woody macrofossils; **aquatic/marsh/swamp-** peat containing a mixture of the three

FIGURE 17
EINP MACROFOSSIL SUMMARY DIAGRAMS
SWAMP SITE 48-3



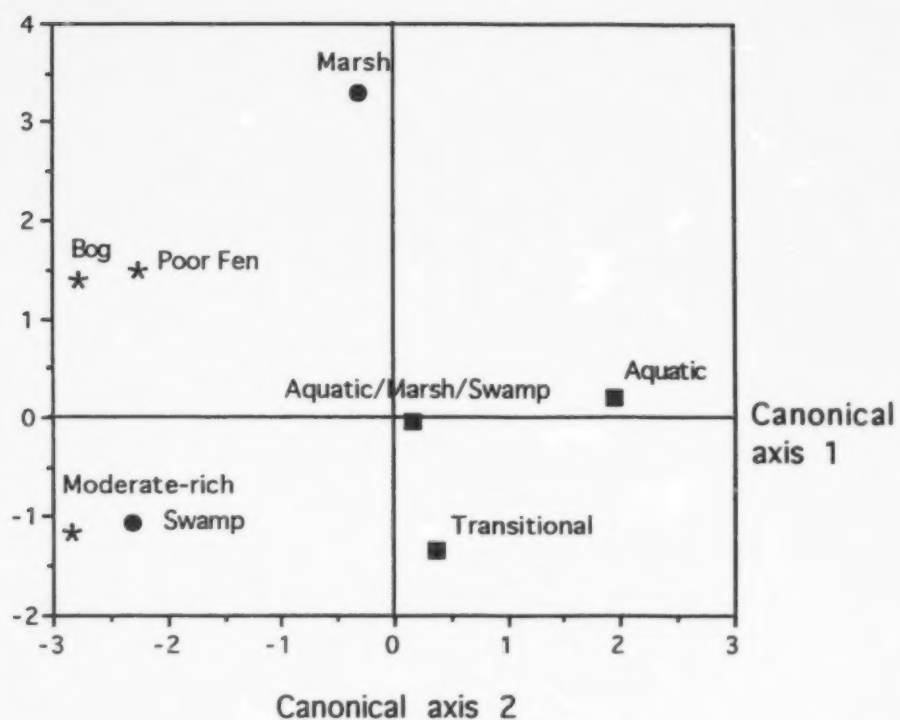
described above; **moderate-rich fen-** terrestrial peat containing fossil bryophytes of the family Amblystegiaceae; **poor fen-** terrestrial peat dominated by *Sphagnum* species but also containing the family Amblystegiaceae; and **bog-** terrestrial peat dominated by *Sphagnum* macrofossils. The canonical discriminant analysis derives a linear combination of variables that has the highest correlation within the given groups. Multivariate statistics and F approximations test the hypothesis that the class means are equal in the population. Each of the multivariate statistics is significant at the 0.0001 level indicating that these categories are significantly different (Table 4).

A graph of the class means for each peat type on the first two canonical axes shows the relative similarities between peat types. Peats that contain aquatic sediments are situated on the right side of the X axis, while marsh, swamp, fen and bogs peats are on the left (Fig. 18). According to the canonical coefficients (Table 5) the most discriminating variable on the X axis is percentage ash followed by depth. The Y axis is most strongly discriminated by site number, indicating that peat chemistry is extremely site specific. Table 6 shows the percentage of samples placed into the correct peat type category by the discriminant analysis. Most peat type categories are well defined chemically as each peat type category was correctly placed between 77-100% of the placements. The exception is peat type aquatic/marsh/swamp which contains elements of all three peat types. Two samples were placed into the adjacent marsh category, one into the aquatic category, and another into the bog category.

Table 4

Canonical Discriminant Analysis
Multivariate Statistics and F Approximations

Statistic	Value	F	Numerator Degrees of Freedom	Denominator Degrees of Freedom	Prob > F
Wilks' Lambda	0.0103	13.78	133	1834	0.0001
Pillai's Trace	2.9591	10.87	133	1974	0.0001
Hotelling-Lawley Trace	8.1948	16.9	133	1920	0.0001
Roy's Greatest Root	3.373	50.06	19	282	0.0001



- Aquatic peats
- Marsh, Swamp peats
- ★ Bog, Fen peats

Fig. 18. Outlines the relationship in discriminant space between peat types based on canonical class means of chemical variables plotted on the first two canonical axis.

Table 5

Canonical Discriminant Analysis
Total-Sample Standardized Canonical Coefficients

	Canonical Axis 1	Canonical Axis 2
Site	0.64	1.70
Core	0.05	0.17
Depth	-0.80	-1.18
Calcium	0.02	0.18
Magnesium	0.03	-0.17
Potassium	0.16	0.45
Aluminum	-0.01	0.15
Titanium	0.00	0.15
Lead	-0.14	0.13
Arsenic	0.09	-0.41
Copper	-0.24	0.13
Iron	0.25	-0.47
Manganese	-0.27	-0.64
Zinc	0.33	-0.20
Nickel	0.06	0.08
Phosphorus	-0.01	0.24
Sulphur	0.40	0.15
Ash	1.24	-0.65
Bulk density	0.03	0.07

Table 6

Into Peat Type

Aq=Aquatic, Mar=Marsh, Sw=Swamp, Mod=Moderate

Aquatic peats are characterized by high ash (66%), bulk density (0.62 g/cm^3), calcium (28,737 mg/Kg), magnesium (3,373 mg/Kg), and potassium (2,162 mg/Kg, Table 7). Aluminum, iron, manganese, and nickel concentrations are high, while phosphorus and sulphur concentrations are intermediate.

Transitional peats that contain both aquatic and fen sediments are intermediate between the two peat types in ash content, bulk density, magnesium, sodium, potassium, aluminum, lead, arsenic, manganese, zinc, nickel and phosphorus. This group is higher than aquatic or moderate-rich fen peats in sulphur.

Aquatic/Marsh/Swamp peats have high ash (33%), bulk density (0.44 g/cm^3), calcium (22,075 mg/Kg), magnesium (2,347 mg/Kg), potassium (1,506 mg/Kg), sodium (646 mg/Kg), aluminum (4,811 mg/Kg), titanium (40 mg/Kg), copper (9 mg/Kg), and zinc (121 mg/Kg)(Table 7). In comparison to other peat types, phosphorus and sulphur levels are average having 866 mg/Kg P and 6,654 mg/Kg S respectively.

Marsh peats at Elk Island National Park have the highest calcium (34,480 mg/Kg) and sodium (1,862 mg/Kg) contents (Table 7). Some elemental concentrations are reduced in this peat type as compared to other peats experiencing groundwater input (Fe 3,929 mg/Kg, Mn 82 mg/Kg, Zn 72 mg/Kg, Ni 12 mg/Kg). Sulphur and phosphorus levels are the highest in this peat type at 12,293 mg/Kg and 1,092 mg/Kg, respectively. Bulk density and ash content of this group are relatively low at 0.15 g/cm^3 and 19%.

Swamp peats have intermediate ash percentages (13%), and bulk density values (0.14 g/cm^3), but high concentrations of calcium

Table 7
Peat Chemistry

Chemical and Physical Variables	Peat type						
	Aquatic	Transitional (Aquatic/Fen)	Aq./Marsh/Swamp	Marsh	Swamp	Moderate-rich Fen	Poor Fen Bog
Ca (mg/Kg)	28,737 29,084	18,383 16,522	22,075 16,622	34,480 12,747	28,392 6,766	9,766 3,140	7,012 3,277 4,660 1,910
Mg	3,373 1,686	2,298 657	2,347 1,506	2,409 416	2,980 270	2,193 1,481	857 503 870 263
K	2,162 1,006	1,387 584	1,506 1,359	669 1,038	475 549	1,159 1,661	364 580 924 1,011
Na	477 404	210 77	646 975	1,862 1,205	516 1,114	197 112	86 80 107 45
Al	10,107 4,799	6,888 2,500	4,811 4,132	2,735 1,983	1,883 1,127	2,198 2,370	2,320 4,068 1,401 818
Ti	25 18	16 9	40 36	28 13	30 14	13 5	15 5 11 3
Pb	16 10	14 6	7 6	5 6	3 1	7 8	8 6 11 16
As	17 7	14 5	9 8	5 3	5 1	4 3	4 5 3 1
Cu	9 4	5 2	9 56	10 6	6 2	3 3	4 3 4 2
Fe	10,014 3,375	12,632 6,506	5,357 3,863	3,929 2,285	6,323 4,012	2,521 1,630	2,569 2,709 1,585 873
Mn	355 300	339 215	182 244	82 74	203 169	225 309	55 56 128 188
Zn	71 35	99 55	121 69	72 64	63 56	51 46	22 13 36 16
Ni	20 22	18 20	13 7	12 22	8 3	8 13	5 6 6 5
P	603 249	921 476	866 498	1,092 307	972 464	1,004 484	477 249 583 211
S	7,774 5,220	9,159 5,427	6,654 4,554	12,293 4,204	4,233 3,746	2,688 1,497	1,929 2,096 946 327
% Ash	66 22	35 14	33 27	19 8	13 6	11 11	8 11 6 4
Bulk Density (g/cm ³)	0.62 0.38	0.19 0.13	0.44 0.35	0.15 0.08	0.14 0.06	0.14 0.10	0.12 0.08 0.07 0.04
n	99	84	33	15	8	46	27 17

Aq. = Aquatic

Mean values plus standard deviations of the chemical and physical variables of peat samples grouped according to peat type.

(28,392 mg/Kg), magnesium (2,980 mg/Kg), and sodium (516 mg/Kg) (Table 7). In general elemental concentrations are low, except for titanium (30 mg/Kg) and iron (6,323 mg/Kg). Phosphorus levels are high (972 mg/Kg), while sulphur is intermediate (4,233 mg/Kg).

Moderate-rich fen peats contain low amounts of ash (11%), calcium (9,766 mg/Kg), and sodium (197 mg/Kg, Table 7). Elemental contents are also low, reflecting limited groundwater infiltration (Ti 13 mg/Kg, As 4 mg/Kg, Fe 2,521 mg/Kg, Zn 51 mg/Kg, Ni 8 mg/Kg). Phosphorus, sulphur, and bulk density measurements are higher than the poor fen or bog peats (1,004 mg/Kg, 2,688 mg/Kg, 0.14 g/cm³)

Poor fen peats are relatively low in ash (8%), calcium (7,012 mg/Kg), magnesium (857 mg/Kg), potassium (364 mg/Kg), sodium (86 mg/Kg), aluminum (2,320 mg/Kg), titanium (15 mg/Kg), Lead (8 mg/Kg), arsenic (4 mg/Kg), copper (4 mg/Kg), iron, (2,569 mg/Kg), Manganese (55 mg/Kg), Zinc (22 mg/Kg), Nickel (5 mg/Kg), phosphorus (477 mg/Kg), sulphur (1,929 mg/Kg), and bulk density (0.12 g/cm³, Table 7).

Bog peats are low in ash (6%), calcium (4,660 mg/Kg), magnesium (870 mg/Kg), most elements, sulphur (946 mg/Kg), and bulk density (0.07 g/cm³, Table 7)

According to the results of the discriminant analysis, it is possible to identify peat types based upon peat chemistry. Peats that contain aquatic sediments comprise the first three peat types (aquatic, transitional, aquatic/marsh/swamp). They are most easily segregated from the remaining peat types by differences in ash content and bulk density. Peat that has a component of aquatic sediments generally has an ash content greater than 20% and bulk

density greater than 0.15 g/cm^3 . Metal concentrations in these peats are characteristically high. Particularly good aquatic indicators appear to be aluminum concentrations greater than $2,735 \text{ mg/Kg}$, arsenic concentrations greater than 5 mg/Kg , iron concentrations greater than $3,929 \text{ mg/Kg}$, zinc concentrations greater than 72 mg/Kg , and nickel concentrations greater than 12 mg/Kg (Table 7).

Marsh and swamp peats are distinguishable from fen and bog peats on the basis of calcium concentrations. Marsh and swamp peats have high calcium values ($> 10,000 \text{ mg/Kg}$) and high sodium concentrations compared to fens and bogs (Table 7). Additional elements segregating the two appear to be titanium, copper, zinc, and sulphur. Ash percentages are also usually higher in marsh and swamp peats at Elk Island.

On the basis of peat chemistry it is difficult to distinguish between marsh and swamp peats at Elk Island. Swamps tend to be slightly less mineral rich, reflected in lower calcium, potassium, sodium, ash, and bulk density values, and slightly less nutrient rich reflected in lower phosphorus and sulphur concentrations. Water chemistry parameters confirm that marshes are more ionically rich and more eutrophic than swamps at Elk Island (Table 2).

The chemical parameters that define the gradient segregating bogs from fens has been well documented (Sjörs 1952, Gorham 1956, Jeglum 1971, 1972, Vitt *et al.* 1975, Horton *et al.* 1979, Slack *et al.* 1980, Glaser *et al.* 1981, Sims *et al.* 1982, Karlin and Bliss 1984, Vitt and Chee 1990, Vitt *et al.* 1990). At Elk Island, the fen peats have higher calcium, aluminum, titanium, iron, sulphur, ash, and bulk density than those from bogs (Table 7).

Peat Profiles

The chemistry of organic matter in peat deposits is the result of several processes. Peatlands have different mineral sources depending upon the hydrology of the site (Siegel 1983, 1988, Damman 1986, Siegel and Glaser 1987, Glaser *et al.* 1990). Peat deposits can sequester chemical elements through cation exchange (Malmer and Sjörs 1955, Gorham 1957, Clymo 1963, Yefimov and Yefimova 1973, Damman 1978, 1986, Hemond 1980, Lembrechts and Vanderborght 1985, Bayley *et al.* 1986) and under situations where elements can be transported into peatlands through waterflow, the peat material will become enriched with cations (Mörnsjö 1968, Zoltai and Johnson 1985) or metals (Gleason and Coope 1966, Coker and DiLabio 1979). Compaction and decomposition of the peat material with depth can account for an increase in some elements. It has been suggested by Zoltai and Johnson (1985), that sulphur increases with depth due to retention of the element and the loss of organic material through decomposition. Elements can also be released from a peatland under conditions of redox changes (Bayley *et al.* 1986, Shotyk 1986), and leaching (Damman 1986). Of the several elements analyzed and reported in this study, only calcium and copper, both of which are well known for their strong retention properties in peat (Damman 1986, Shotyk 1986), have been used to study wetland evolution.

Physical Profiles

Percentage Ash

Ash content of peat profiles in the fens at Elk Island National park decline towards the peat surface (Fig. 19). Aquatic peats contain > 40% ash, transitional aquatic/fen peats contain < 40% ash, and moderate-rich fen peats contain less than 15% ash. Ash percentages in peat profiles of the bogs (Fig. 20) experience a rapid decline between aquatic peats containing > 40% to < 20% in the transitional aquatic/fen peats or to < 5% in the poor fen and bog peats. Ash profiles at site 22 are more gradual than site 27. Ash contents in profiles from swamp/marsh wetlands (Fig. 21) decline from > 40% in the aquatic sediments to < 20% in the transitional aquatic/marsh peats. Surface sediments from both wetland types stabilize around an ash content of 10-15%.

Bulk Density

Bulk density profiles in the wetlands of Elk Island National Park (Figs. 22, 23, 24) outline a history of gradual decreasing bulk densities from basal peat to surface peats. Basal aquatic peats in the fen profiles have bulk densities > 0.2 g/cm³; transitional aquatic/fen peats are lighter (>0.1 g/cm³); while surface sediments are often denser (up to 0.5 g/cm³ in cores 2-1, 2-2, 3-1, and 3-2). Aquatic peats in bog site 22 are similar in density to fen peats (> 0.2 g/cm³), while aquatic peats at site 27 cores 2 and 3 are lighter (0.1-0.2 g/cm³). Moderate-rich fen and poor fen peats at site 27 have a bulk density of approx. 0.1 g/cm³, and are variable in core 3. Bulk density measurements of surface peats were unobtainable at site 22, because

Figure 19
Physical Profiles
Percentage Ash-Fens

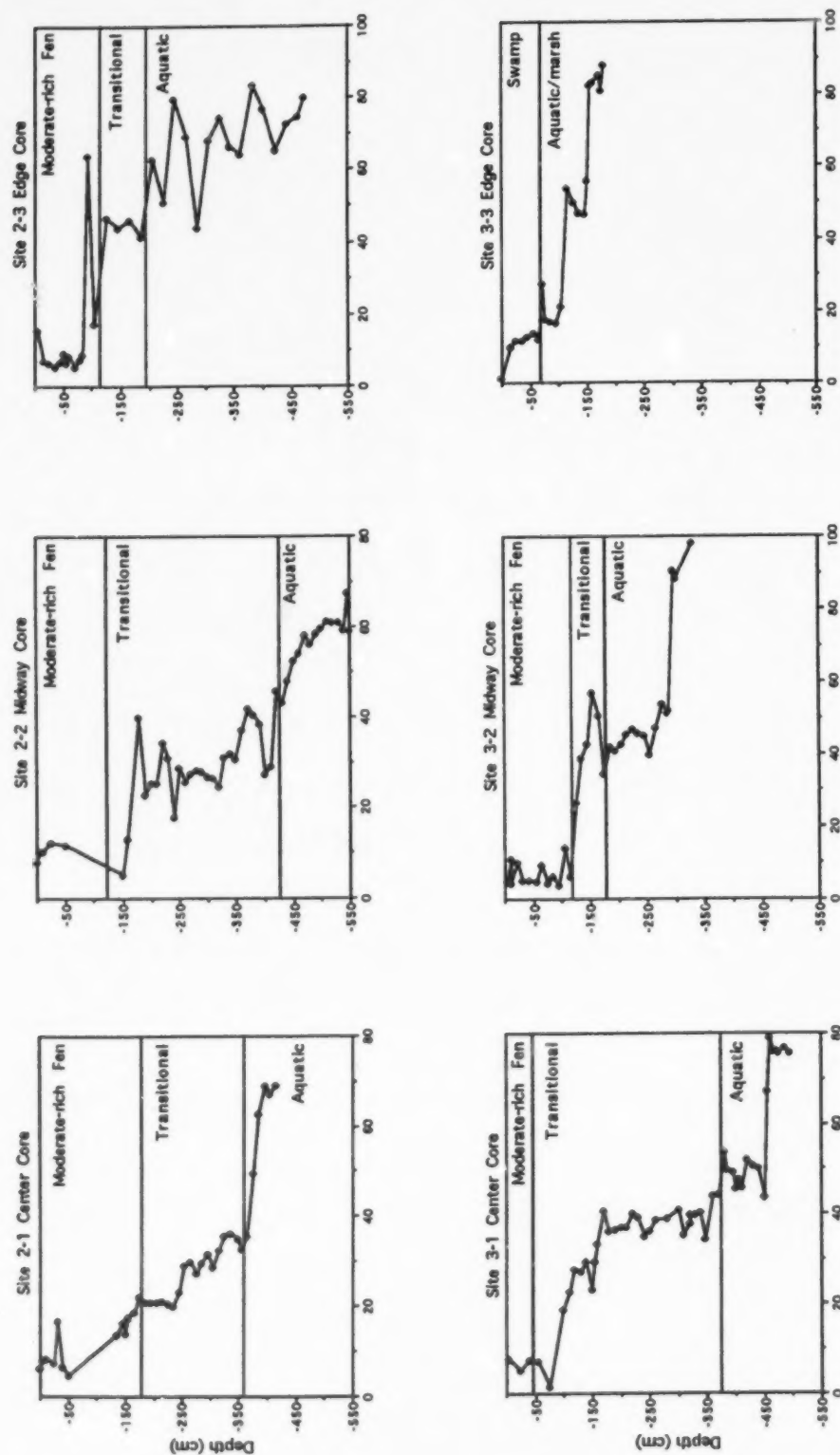


Figure 20
Physical Profiles
Percentage Ash-Bogs

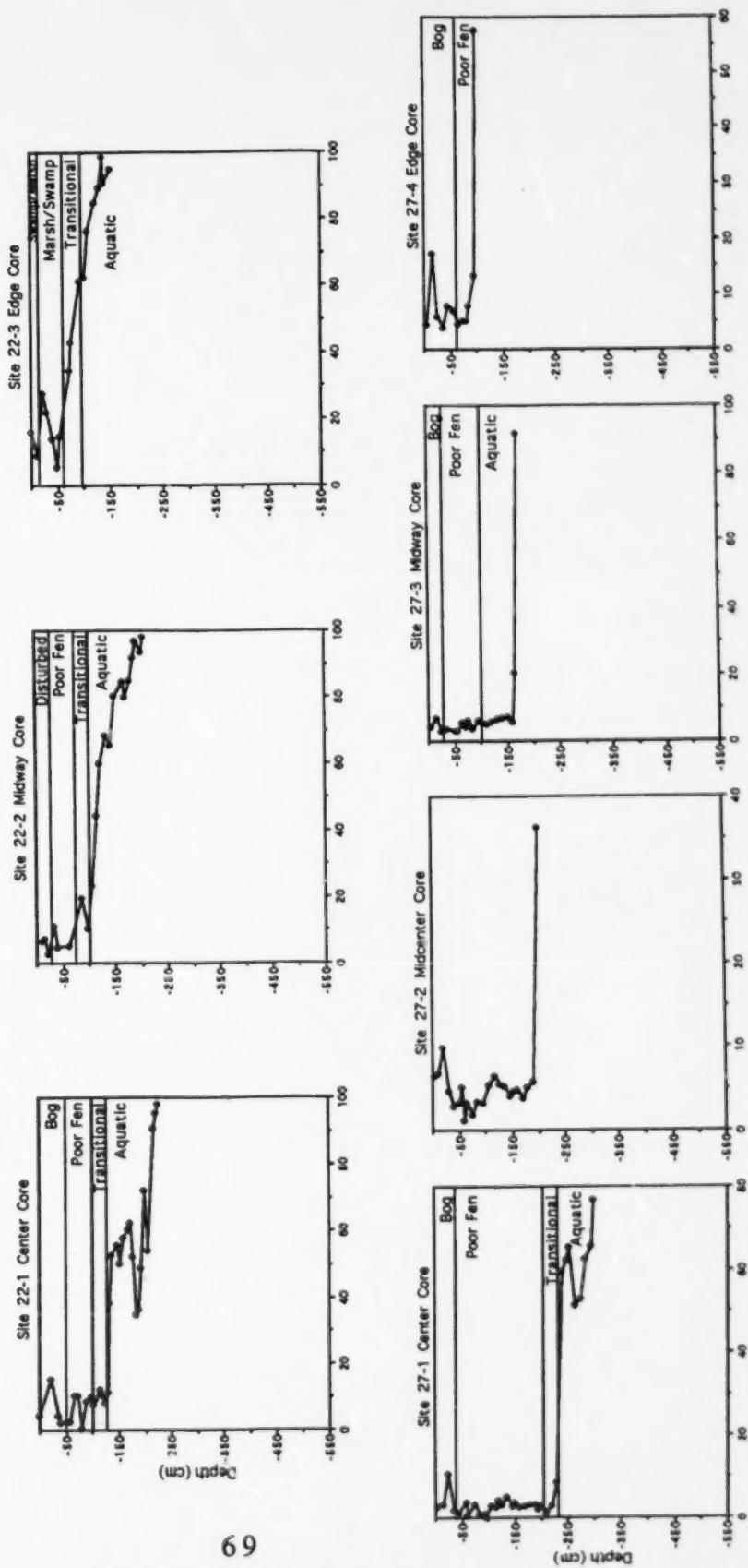


Figure 21
Physical Profiles
Percentage Ash-Marsh/Swamps

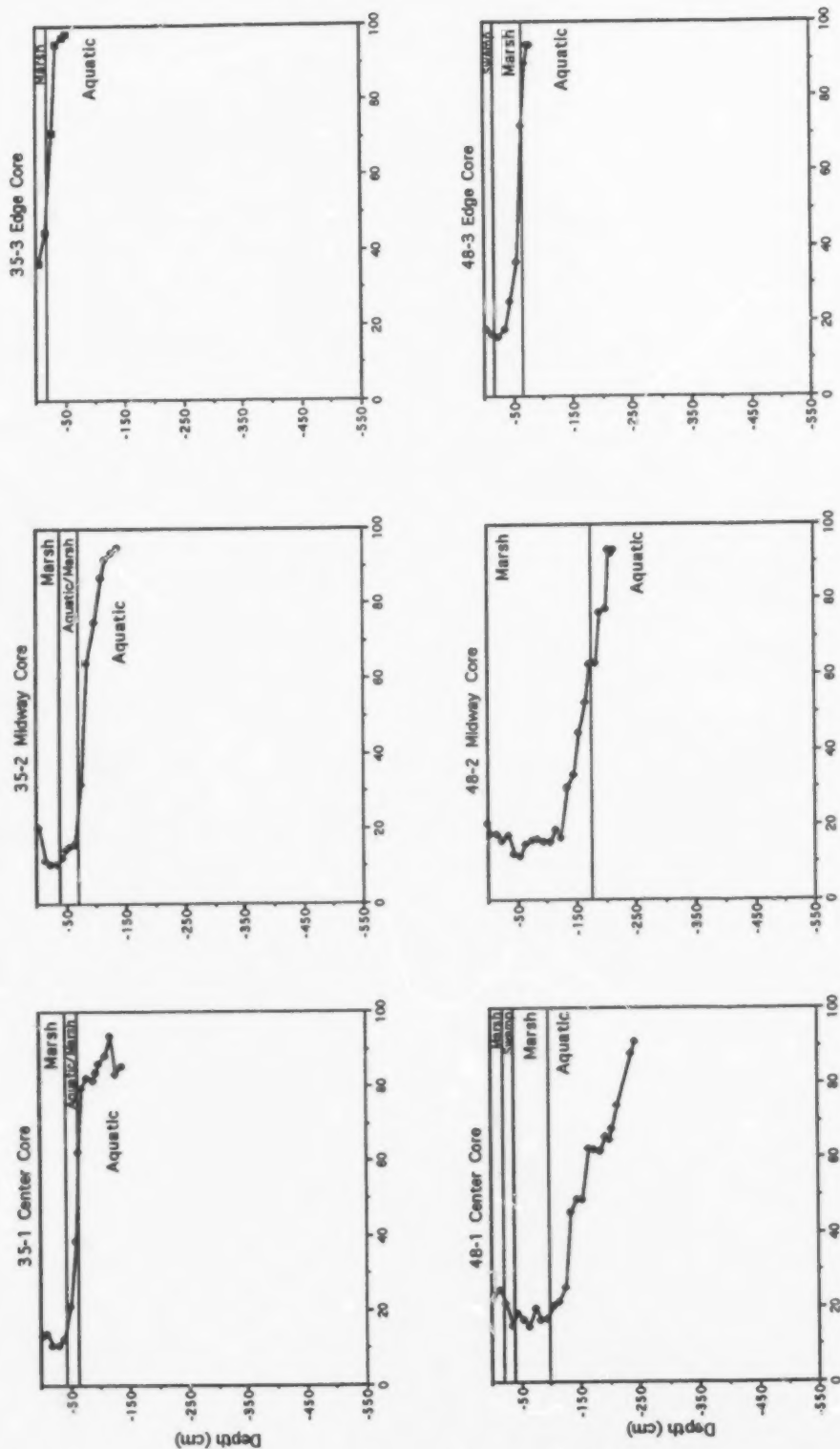


Figure 22
Physical Profiles
Bulk Density-Fens g/cm^3

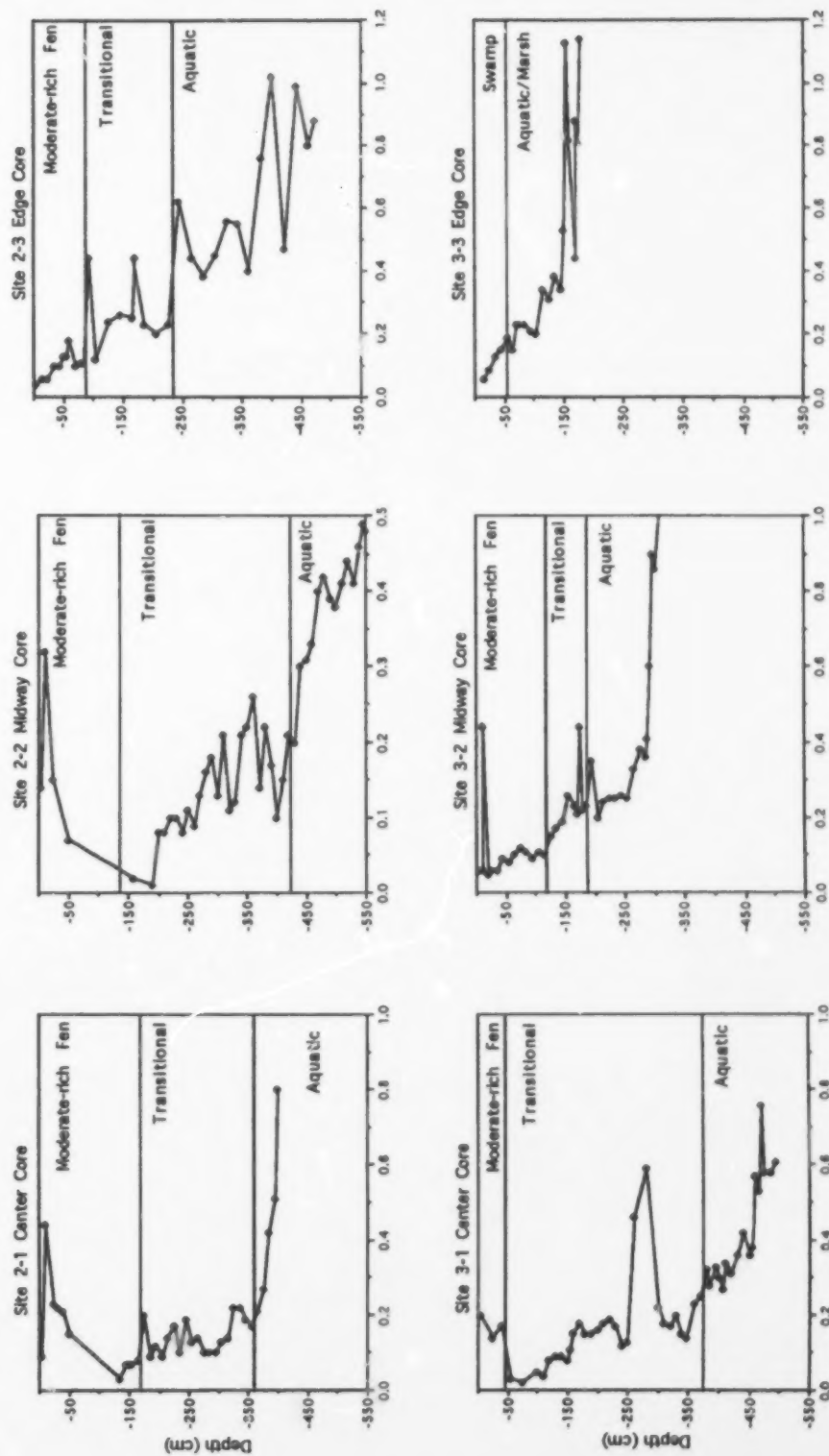


Figure 23
Physical Profiles
Bulk Density-Bogs g/cm^3

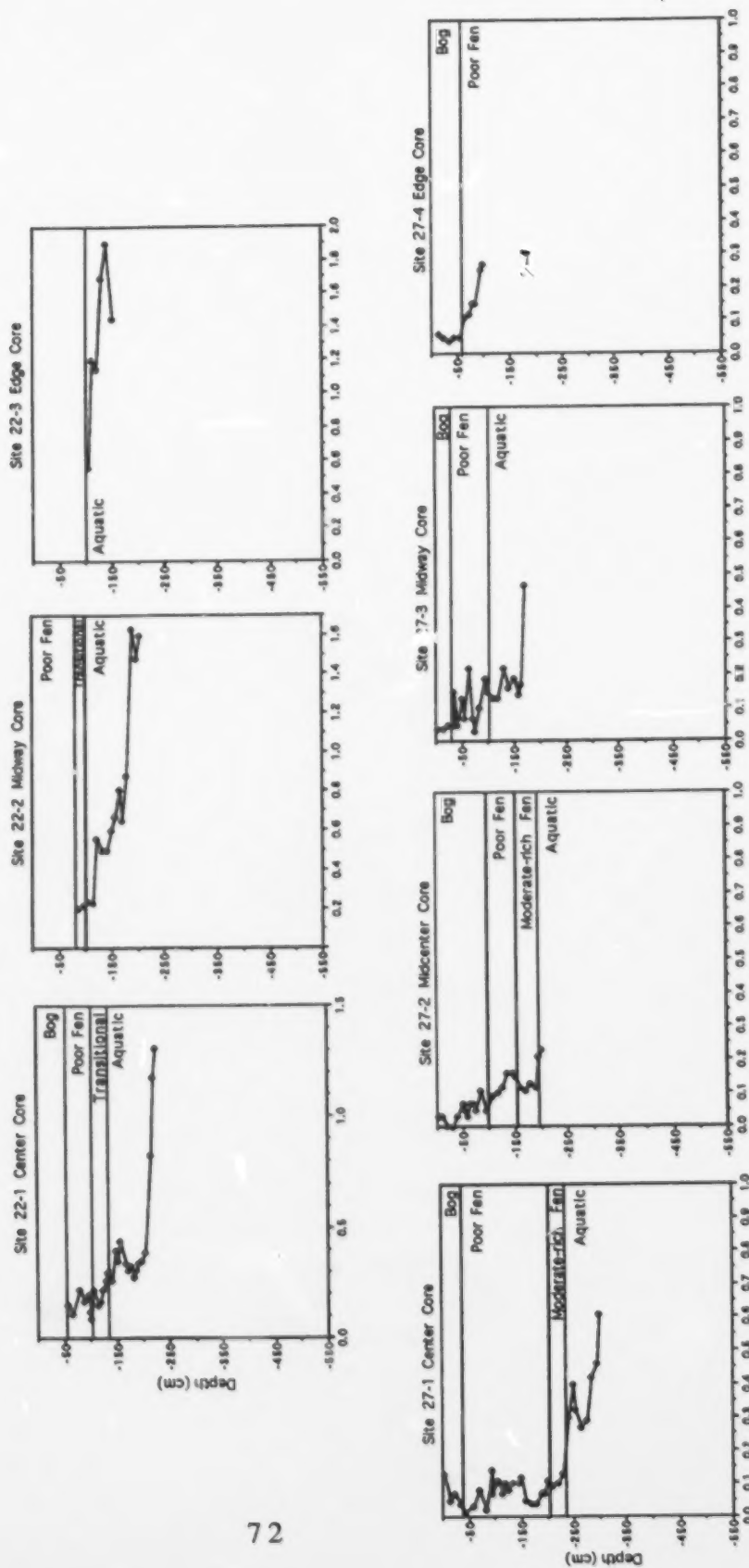
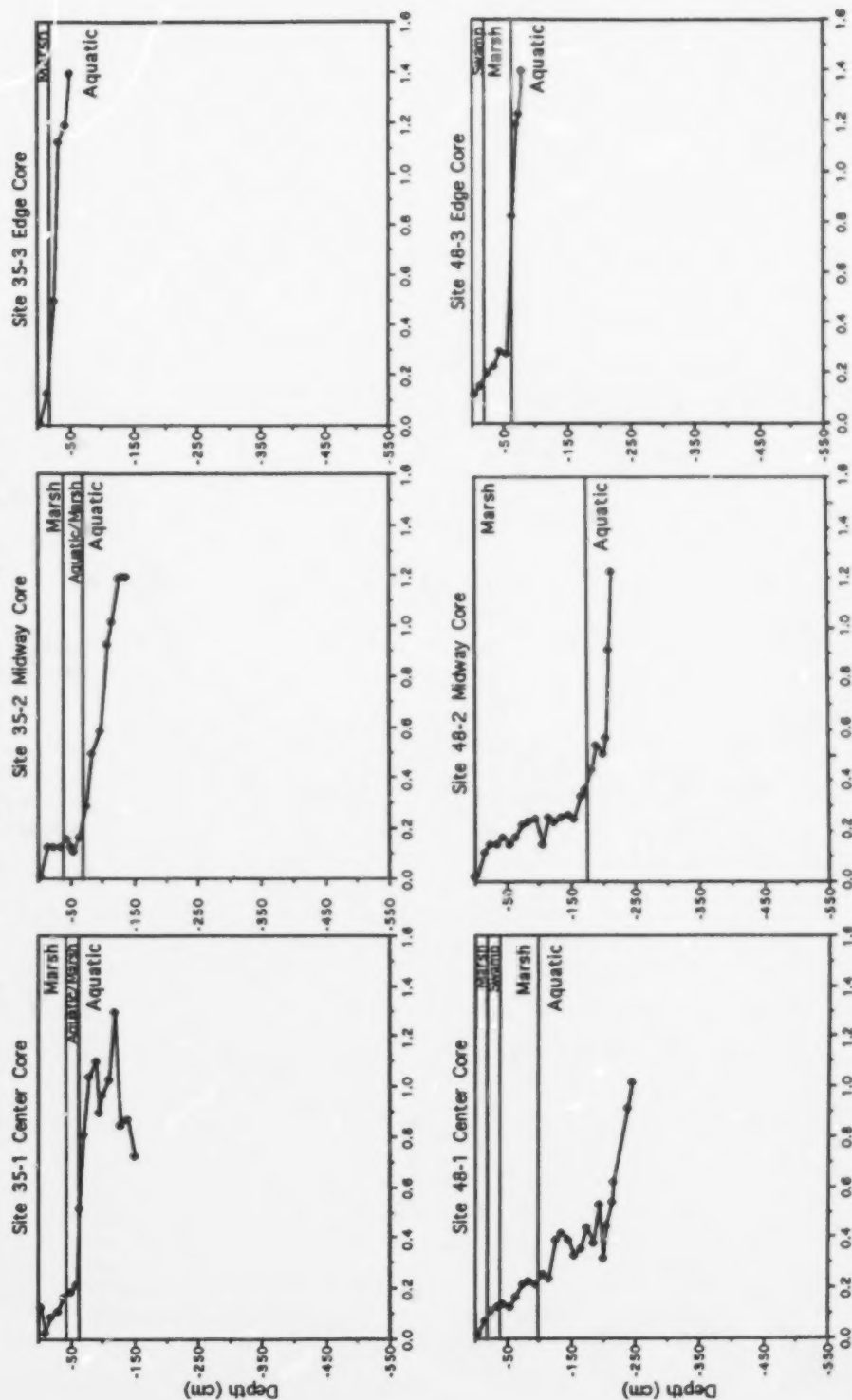


Figure 24
Physical Profiles
Bulk Density-Marsh/Swamps g/cm^3



the sediments were frozen at the time of sampling, and volumetric measurements could not be taken. Basal aquatic peats in the marshes and swamps tend to be denser than fen and bog basal peats ($>0.3 \text{ g/cm}^3$). These profiles decline rapidly to bulk density values usually less than 0.2 g/cm^3 . Bulk density values in the surface peats of marshes and swamps are lowest at the surface compared to bog and fens which can have denser peats at the mire surface.

Chemical Profiles

Calcium

In calcium and copper profiles, there is a visible difference between cores taken in the center of the wetland and cores taken from the edge. Profiles of cores taken from the center portions of fens, have high calcium concentrations in the aquatic sediments (1×10^5 - $7 \times 10^5 \text{ mg/Kg}$, Fig. 25). This does not occur in edge core 2-3, as calcium values are stable around 1×10^5 . At edge core 3-3 aquatic/marsh sediments contain 5×10^5 - $3 \times 10^5 \text{ mg/Kg}$ of calcium peaking to $35 \times 10^5 \text{ mg/Kg}$ in the swamp sediments. In the central cores, calcium concentrations decline gradually through the deep section of transitional peats, and stabilize at about $1 \times 10^5 \text{ mg/Kg}$ in the moderate-rich fen peats. Central bog cores from bog site 27, (Fig. 26) contain approximately $1 \times 10^5 \text{ mg/Kg}$ of calcium in the basal aquatic peats which is the lowest calcium values found through the fen profiles. Calcium values reduce very slowly through these peat profiles to lows of approximately $3 \times 10^4 \text{ mg/Kg}$ at the peat surface. Core 22-1 has a slightly different profile than the others. Basal

Figure 25
Chemical Profiles
Calcium-Fens mg/Kg

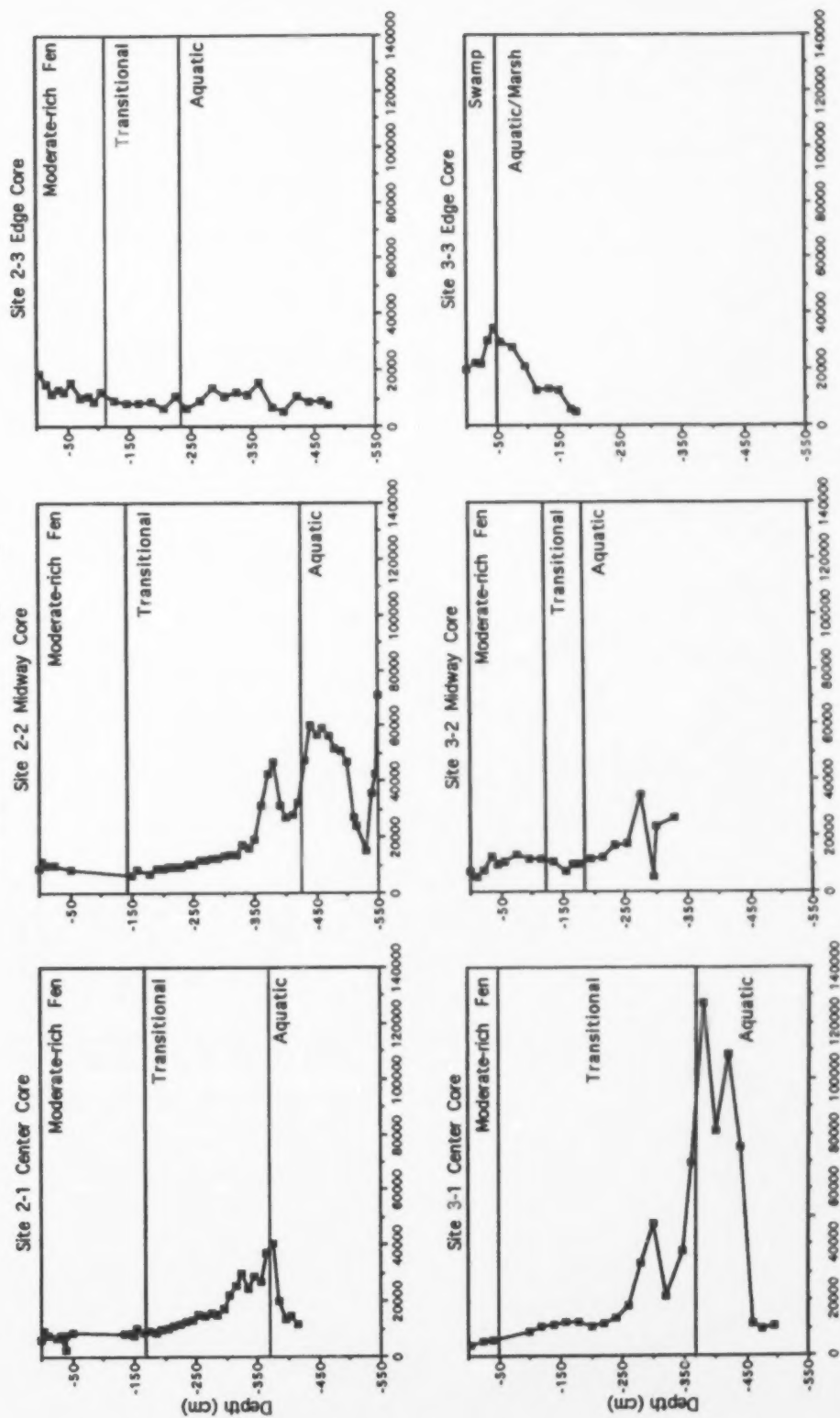
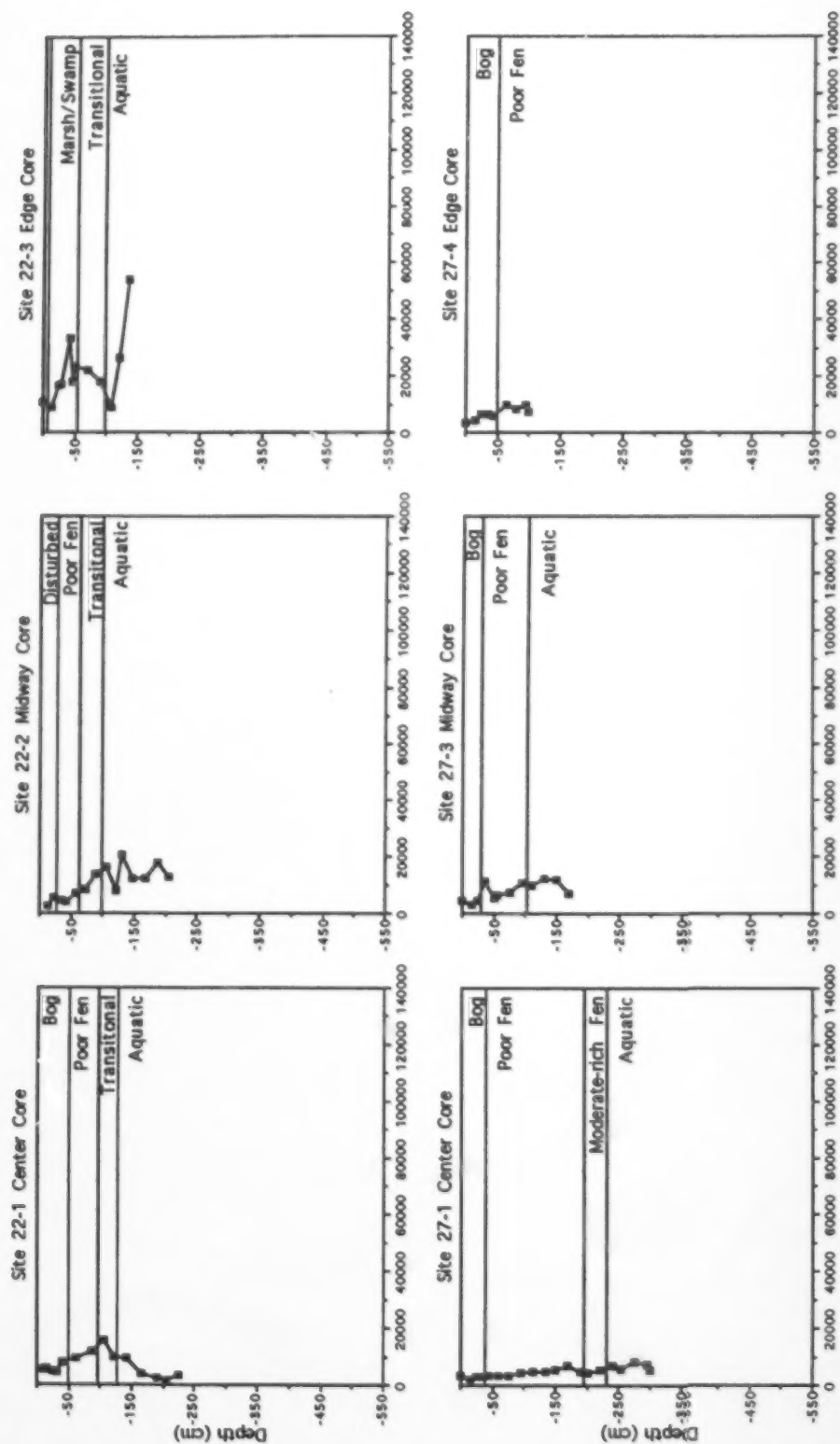


Figure 26
Chemical Profiles
Calcium-Bogs mg/Kg



aquatic peats have extremely low calcium values (1.7×10^3 - 3.7×10^3 mg/Kg) which peak 15.6×10^3 mg/Kg before declining to 4.5×10^3 mg/Kg near the peat surface. Core 22-2 has basal calcium concentrations of 13 - 20×10^3 mg/Kg which decline to 3×10^3 mg/Kg at the peat surface. Calcium profiles from central cores at marsh site 35 (Fig. 27) are unique in that surface peats have much higher calcium concentrations than basal peats. Basal aquatic sediments contain 1×10^4 mg/Kg of calcium, while peak calcium values of 35×10^3 mg/Kg occur 50 cm below the surface. The edge core (35-3) at this site has a highly variable calcium profile. At marsh site 48, calcium values are highest in the aquatic peats (70 - 150×10^3) and decline gradually to about 30×10^3 at the peat surface. At cores 48-1 and 48-3 calcium values rise at the surface to 50 - 70×10^3 mg/Kg. The calcium profile at swamp core 22-3 is distinctly more variable. Basal aquatic sediments contain calcium values of 5×10^4 mg/Kg, transitional aquatic/marsh/swamp peats average 2×10^4 mg/Kg, while surface sediments are somewhat reduced at 1×10^4 mg/Kg.

Copper

In the fens of Elk Island National Park, basal aquatic peats contain 3-14 mg/Kg of copper (Fig. 28). Profiles describe a gradual reduction in copper towards the peat surface, with some surface enrichment. Transitional aquatic/fen peats range from 1-9 mg/Kg, while surface peats can contain as high as 18 mg/Kg. Copper concentrations in bog site 27 (cores 27-1, 27-3) decrease abruptly from high levels of 18 mg/Kg in the aquatic basal peats to approximately 3 mg/Kg in the fen peat (Fig. 29). At core 27-4 poor

Figure 27
Chemical Profiles
Calcium-Marsh/Swamp mg/Kg

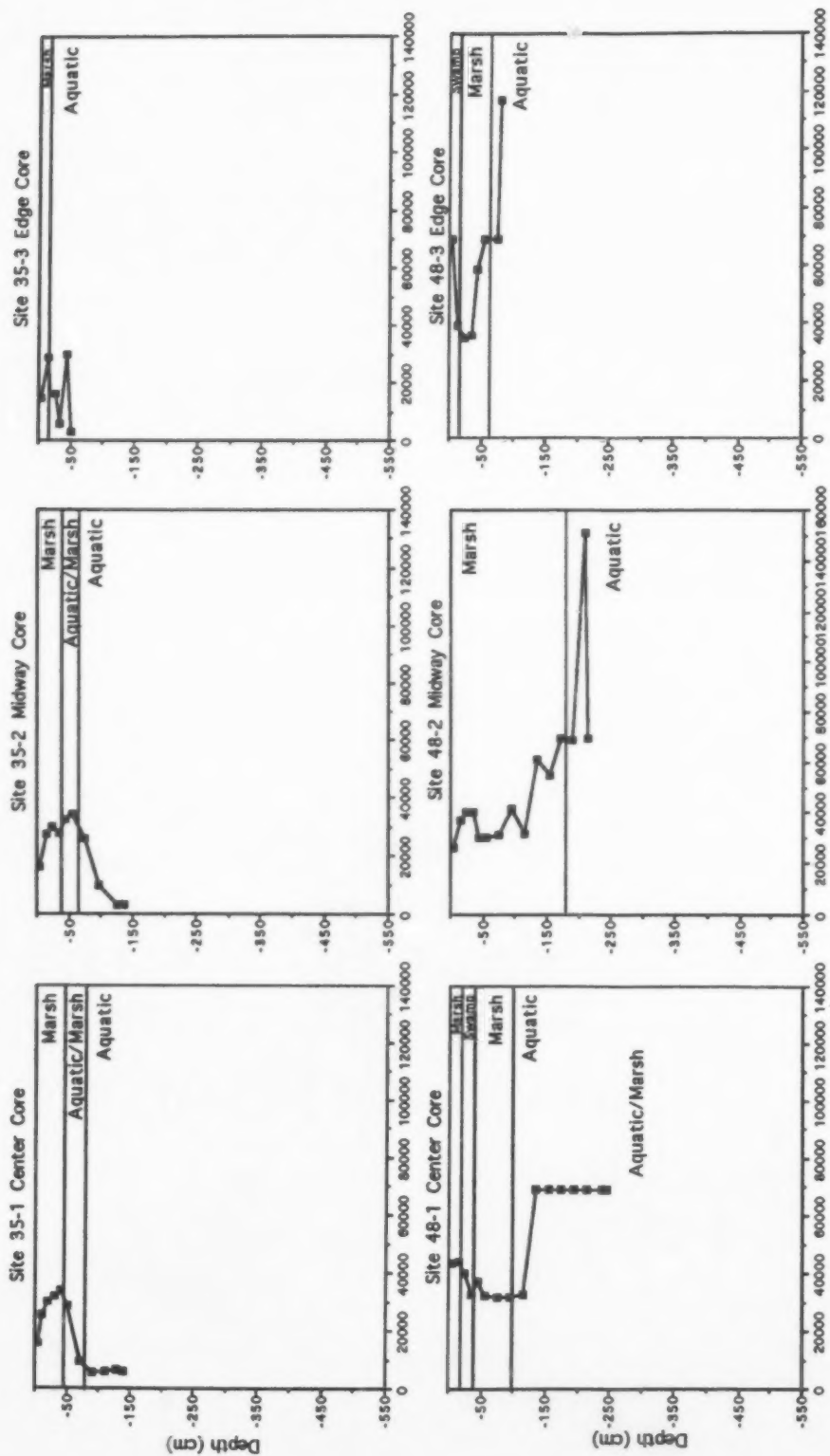


Figure 28
Chemical Profiles
Copper-Fens mg/Kg

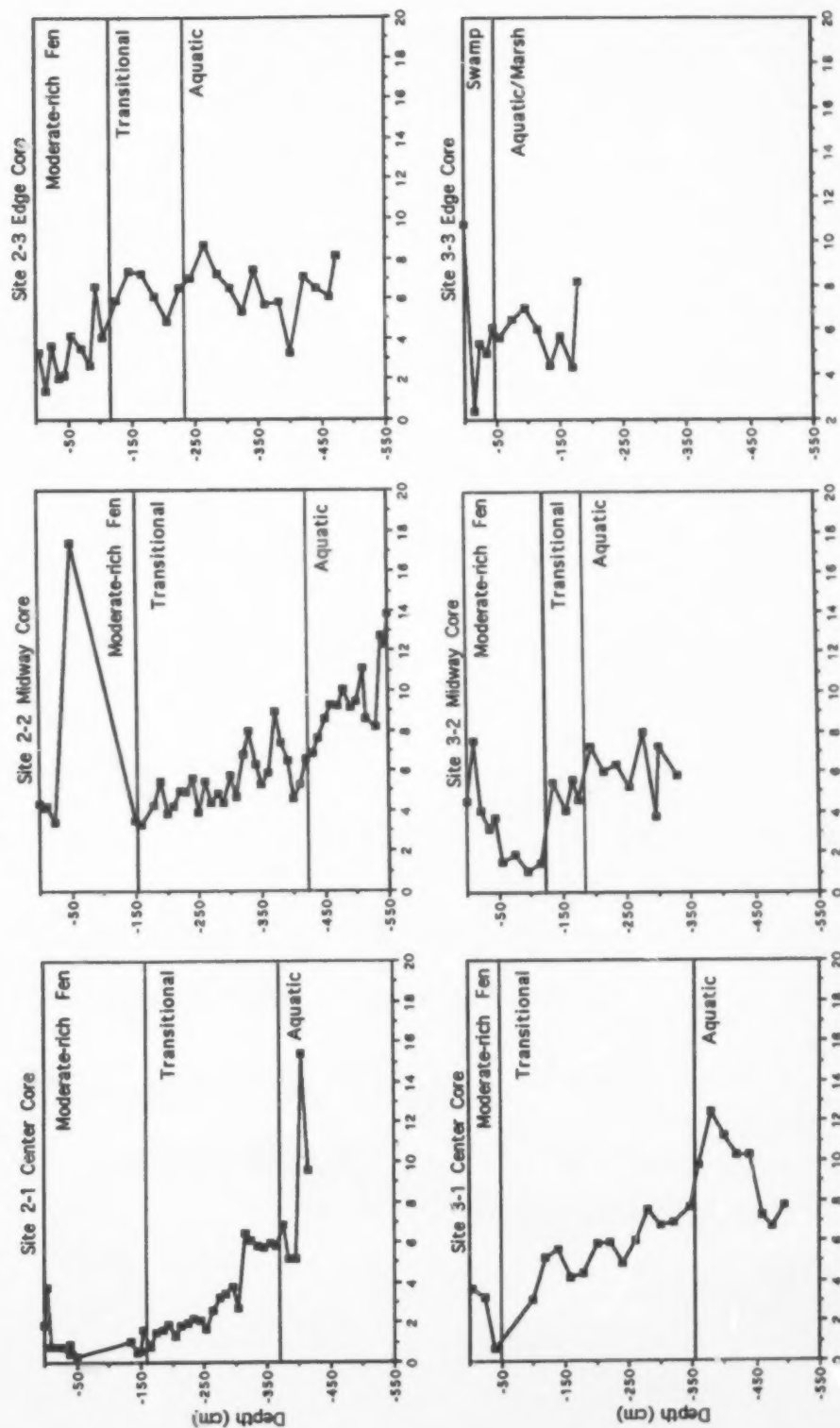
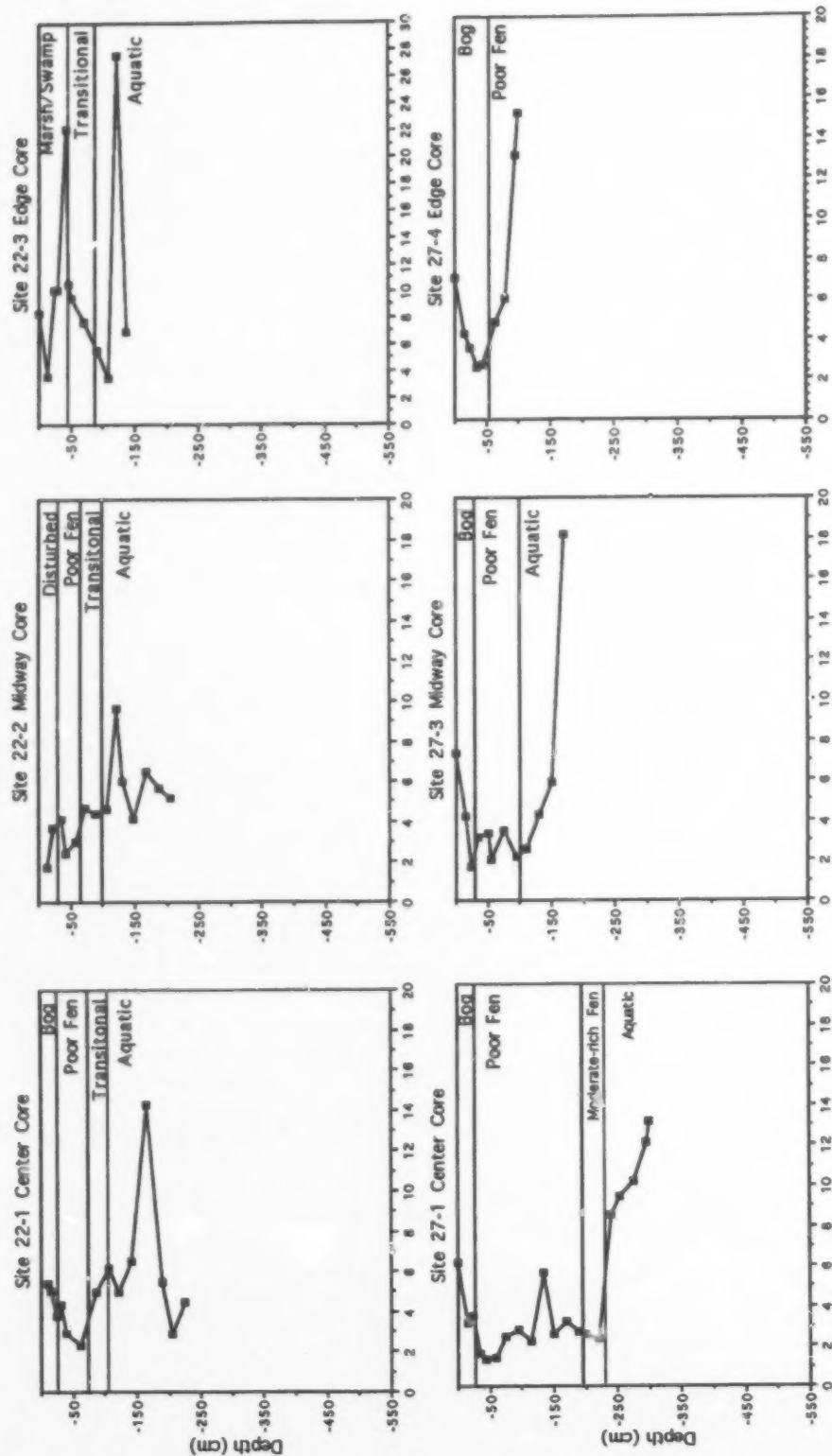


Figure 29
Chemical Profiles
Copper-Bogs mg/Kg



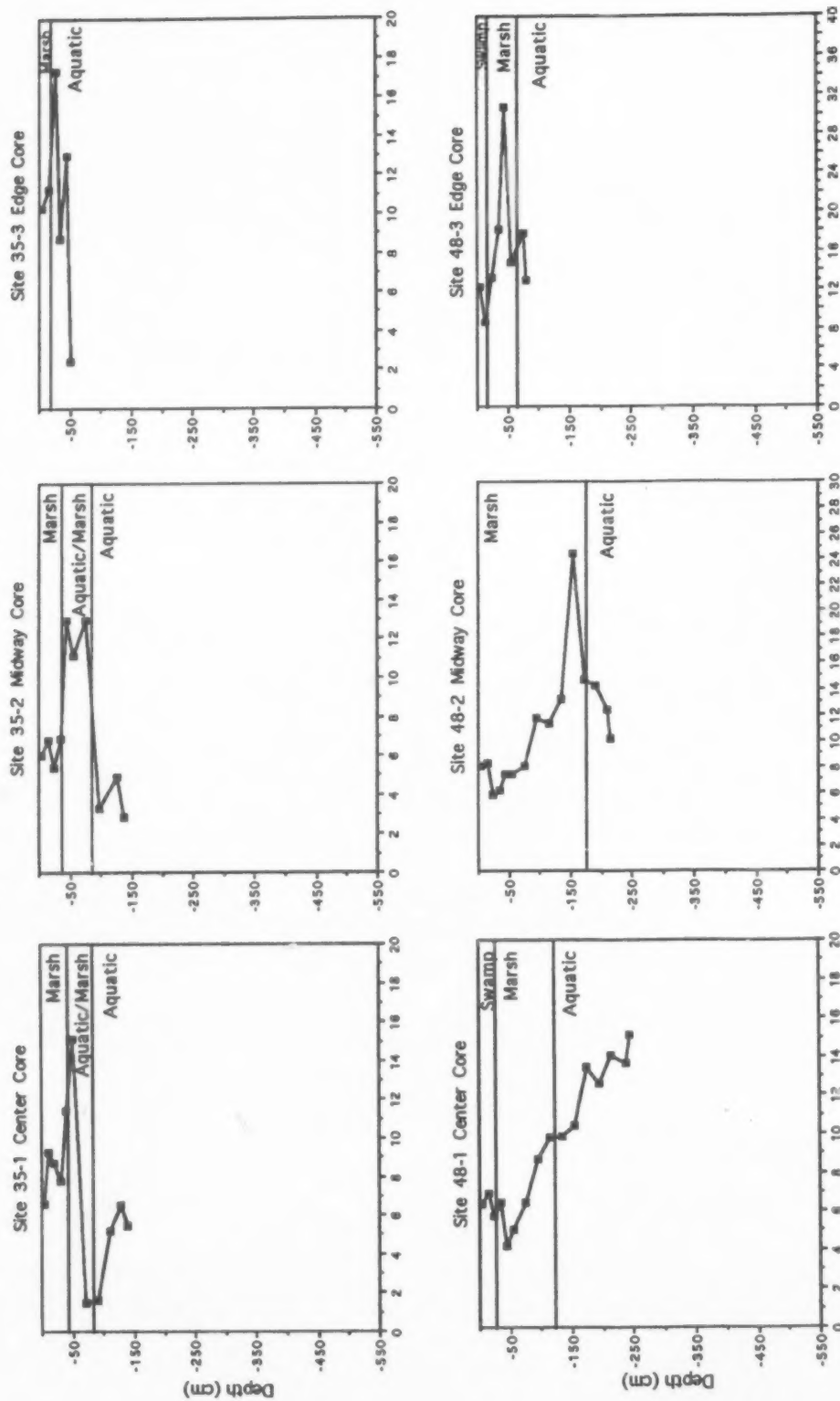
fen peats contain 5-15.8 mg/Kg. Surface enrichment of copper can be seen in all three profiles, with values ranging from 2-7 mg/Kg. At site 22 the central cores (core 22-1, 22-2), describe an initial flush of copper into the aquatic sediments peaking at 14 mg/Kg in core 22-1, followed by a reduction through the fen peats (average of 4 mg/Kg), and some surface enrichment (> 5 mg/Kg). Marsh and swamp cores have extremely variable copper profiles (Fig. 30). Initial copper flushes occur in the aquatic peats in cores 22-3, 35-1, 35-2, and 35-3, with peaks of 13-23 mg/Kg occurring in the lower marsh and aquatic/marsh sediments, and reductions in the upper marsh sediments. At cores 48-2 and 48-3, peak copper contents of 24 and 31 mg/Kg occur in the marsh peats above initial low (10-13 mg/Kg) copper values in the aquatic peats. Core 48-1 describes a gradual reduction in copper from basal aquatic sediments containing 15 mg/Kg of copper to low values of 4 mg/Kg in marsh sediments 50 cm below the surface. The surface sediments at this core are slightly enriched at about 7 mg/Kg of copper.

DISCUSSION

Classification of Wetland Sites

Group 1 consisting of one site, represents a saline marsh as indicated by the presence of *Triglochin* and *Scirpus*, and by the high surface water values of corrected specific conductivity and sodium. Moss (1983) indicates that *Triglochin* is a perennial found in brackish marshes, while Stewart and Kantrud (1972) identifies *Triglochin maritima* as a species found in brackish to subsaline waters. Waters

Figure 30
Chemical Profiles
Copper-Marsh/Swamp mg/Kg



having specific conductivities higher than 500 μS are considered to be oligosaline by Cowardin *et al.* (1979), and slightly brackish by Stewart and Kantrud (1972). Millar (1976) considers all water with specific conductivities less than 2,000 to be fresh. *Triglochin* dominated wetlands are classified as persistent emergent wetlands according to the U.S. classification system of Cowardin *et al.* (1979).

Group 2 consists of *Typha*-dominated deep marsh stands that are similar in species composition to *Scirpus-Typha* reed swamp vegetation identified by Lewis *et al.* (1928), and Moss (1953). A similar *Typha latifolia-Lemna minor* open fen vegetation was described by Jeglum (1972). *Typha*-dominated vegetation is found in emergent deep marshes (Millar 1976, Stewart and Kantrud 1972); slightly saline deep marshes (Walker and Coupland 1970); and shallow shore marshes by National Wetlands Working Group (1988). Cowardin *et al.* (1979) placed *Typha*-dominated vegetation into their palustrine persistent emergent wetlands group. This group, however is intermediate between emergent deep marsh and shallow marsh as defined by Shay and Shay (1986). The water chemistry of this group is considered to be slightly brackish by Stewart and Kantrud (1972), but fresh by Cowardin *et al.* (1979) standards.

Group 3 consists of *Carex aquatilis* and *Carex atherodes* dominated wetlands. They fit well into "meadow marshes" as defined by the National Wetlands Working Group (1988) and Walker and Coupland (1970). Similar vegetation was described as a broad leaved sedge fen by Jeglum (1972), and a low moor *Carex aquatilis* association by Lewis *et al.* (1928). *Carex atherodes* dominated vegetation is listed under shallow marsh communities by Millar

(1976), and Stewart and Kantrud (1972), while sedge dominated wetlands are classified as palustrine persistent emergent wetlands according to Cowardin *et al.* (1979). In Shay and Shay's (1986) classification this group is intermediate in composition between shallow marsh and wet meadow categories.

Group 4 consists of bryophyte dominated moderate-rich fens. This group bears some similarity to Glaser *et al.* (1990) spring fen channel vegetation with its mixture of *Carex lasiocarpa*, *Carex aquatilis*, and *Larix laricina*; Jeglum's (1972) narrow-leaved sedge fen category; and Vitt and Chee's (1990) moderate-rich fen species of Stand Group 2. Moss (1953) considered this vegetation to be a marsh-bog transition and Lewis *et al.* (1928) a low moor caricetum. According to the U.S. classification system of Cowardin *et al.* (1979), this wetland community is a palustrine persistent emergent wetland, while the Canadian System of Classification considers these to be sedge-grass type basin fens (National Wetlands Working Group 1988).

Group 5 consists of shallow marsh stands that appear to have developed in a more stable environment than Group 3 as they contain a larger number of bryophyte species and *Carex lasiocarpa*, a species more common in fens. This group is considered to be a meadow marsh by the National Wetlands Working Group (1988), and Walker and Coupland (1970); a broad leaved sedge fen by Jeglum (1972); fen emergent vegetation by Stewart and Kantrud (1972); a low moor *Carex aquatilis* association by Lewis *et al.* (1928); a shallow marsh by Millar (1976); and a shallow marsh/wet meadow intermediate by Shay and Shay (1986). This group has some affinities with

widespread hummock species described by Vitt and Chee (1990) for fens in Alberta.

Group 6 consists of *Salix* dominated swamps similar to either the low shrub or tall shrub fens of Jeglum (1972); willow swamps of Moss (1953); *Salix-Calamagrostis canadensis* association of Lewis *et al.* (1928); and shoreline shrub thickets of Ovenden and Brassard (1989). Millar (1976) considered *Salix* dominated vegetation to be parkland versions of a wet meadow. In the U.S., this vegetation would be classified as a scrub-shrub wetland (Cowardin *et al.* 1979).

Group 7 consists of alder dominated swamps, that resemble tall shrub thickets found in the Yukon described by Ovenden and Brassard (1989); alder-willow swamps described in Alberta by Moss (1953); and tall shrub fens identified in central Saskatchewan by Jeglum (1972). This group would also be considered as a parkland wet meadow wetland by Millar (1976).

Group 8 consists of *Sphagnum* bogs that have open canopies due to a recent fire. They are similar to the *Picea mariana-Ledum groenlandicum-Sphagnum magellanicum* association described in Alberta by Vitt *et al.* (1975), and mire expanse communities found in eastern North America by Gauthier (1988). Similar vegetation associations are found in *Sphagnum* bogs described by Moss (1953), and open bogs of Jeglum (1972). This type of wetland is classified as a basin bog in the Canadian system (National Wetlands Working Group 1988), and as a moss-lichen wetland in the American system (Cowardin *et al.* 1979).

Group 9 consists of *Sphagnum* dominated bogs that have dense *Picea mariana* canopies. Jeglum (1972) accurately described this

association as "wooded bogs". This group appears to be a dry version of the *Picea mariana-Sphagnum* association described in Alberta by Moss (1953). Lewis *et al.* (1928) considered this vegetation type to be the mature hydrarch climax stage for the Edmonton area. In the American wetland classification system this association is equivalent to needle-leaved evergreen and scrub-shrub wetlands (Cowardin *et al.* 1979).

Group 10 consists of *Sphagnum* dominated poor fens with high *Larix* cover. This group conforms most closely to Jeglum's (1972) tamarack swamp category, and Sims *et al.* (1982) *Sphagnum*-rich treed fens in James Bay, Ontario. Cowardin's *et al.* (1979) American classification system identifies this wetland as a needle-leaved deciduous forest wetland.

Group 11 consists of poor fen stands that have significant alder and birch cover. They resemble *Sphagnum*-rich treed fens described by Sims *et al.* (1982), and *Carex aquatilis-Chamaedaphne-Sphagnum* meadows described by Ovenden and Brassard (1989) in the Yukon. According to Cowardin *et al.* (1979) this association is a moss wetland.

Group 12 consists of *Picea glauca-Picea mariana-Larix laricina* swamp associations. This association is classified as a needle-leaved evergreen wetland in the American system of wetland classification (Cowardin *et al.* 1979). This vegetation type has not been described previously in western Canada, but similar vegetation associations are present in the peat margin swamps of eastern temperate Canada (National Wetland Working Group 1988).

Group 13 consists of *Sphagnum* dominated poor fens with some *Betula*, *Picea*, *Larix*, and *Salix* cover. They are most closely related to

Ledum moors, an early *Sphagnum-Andromeda* association described near Nestow, Alberta by Lewis *et al.* (1928). In the U.S., this association would be considered a moss wetland (Cowardin *et al.* 1979).

Correlation of Vegetation to Water Chemistry and Environmental Factors

Data reported from other sources indicate that marshes are typically circumneutral to alkaline in water chemistry. Hydrogen ion activity (pH), specific conductivity, calcium, magnesium, and sodium concentrations can be variable in marshes, ranging from 5.6-8.2, 83-500 μ S, 2-72 mg/l Ca, 1-12 mg/l Mg, and 3-50 mg/l Na (Table 8). Ionic concentrations found in the marshes at Elk Island National Park are consistent with the range of values reported in Table 8. Calcium, magnesium, sodium, corrected specific conductivity, and pH levels in swamps reported from other studies indicate that swamps are also generally circumneutral to alkaline and equally as variable in their water chemistry (Table 8). Ionic concentrations in the swamps at Elk Island are similar to data presented in Table 8. Moderate-rich fens at Elk Island have lower pH, specific conductivity and cation concentrations than macrotrophic and rich fens found elsewhere in Alberta (Table 8). They are comparable to moderate-rich fens, mesotrophic fens, water tracks at Mariana Lakes, graminoid fens, and low shrub fens in pH and corrected specific conductivity. Dust from agricultural soils may be contributing to some calcium enrichment of

Table 8
Wetland Surface Water Chemistry Data - Published Reports

Type	Description	Reference	pH	Cond μ S	Ca mg/l	Mg mg/l	Na mg/l	TKN μ g/l	NO ₃ μ g/l	NH ₄ μ g/l	TP μ g/l	Notes
Marsh	Bulrush	S. Campau et al. unpublished	8.2	500				1383			21	
	Shore Marsh	National Wet Wor Gr 1988	7	95	18							
	Meadow marsh	National Wet Wor Gr 1988	6.8	97	15							
	Marsh/Fen	Vitt et al. 1992	6.7-6.8	180-186	23	8	4-6		4-9	6-7	40-77	sample depth=0.5m
	Thalypertis-reed	Wassen et al. 1989	6.4	440	72	1	51		242	285		
	Floating Marsh	National Wet Wor Gr 1988	5.8	83	9							
	Carex Aquatilis	Cowardin et al. 1979	>5		2							
Swamp	Conifer	Schwinzer 1981	7.4		38	11						
	Conifer	Schwinzer 1981	7.2		44	13						
	Conifer	Schwinzer 1981	7.2		50	12						
	Conifer	Schwinzer 1981	7		28	11						
	Mesotrophic Hard.	Zollet and Johnson 1987	6.1	379	60	19	5				210	
	Alnus	Wassen et al. 1989	6	345	20	0.5	37		235	6000		
	Mesotrophic Conit.	Zollet and Johnson 1987	5.7	192	23	9	5				140	
	Mesotrophic Thickwood	Zollet and Johnson 1987	5.5	182	20	9	7				580	
	Coniferous	National Wet Wor Gr 1988	4.7		8	4					100	
Fen	Rich	Vitt and Chee 1990	7.1-7.3	212	82	15	7	1500**	25	42	40	ORG-N
	Rich	Richelfort 1987	7.3-8	436-819	45-85	21-27	23-33		5-42	3-77		
	Macrotrophic	Zollet and Johnson 1987	6.9	436	64	17			7	53	130	ORG-N
	Moderate-rich	Vitt and Chee 1990	6.2-6.7	88	21	5	6	2234**			80	ORG-N
	Grassland	Sims et al. 1982	6.8	140**	22	4	9	1100			200	umho
	Forested Rich	Vitt et al. 1992	8.1-8.3	82-90	11-12	5	2-3		4-8	15	250-406	umho
	Grassland, Rich/Need	Sims et al. 1982	6.1	91**	17	3	6	1400			300	umho
	Low Shrub	Sims et al. 1982	6	200**	17	5	20	1700			400	umho
	Open Rich	Vitt et al. 1992	5.8	79	10	4-5	1-2		5-7	9	122-412	
	Water Tracks	Nicholson 1987	5.1-6.3	148	10	2	1.6	1079	7	11	5	
	Mesotrophic	Zollet and Johnson 1987	6	245	29	11			39	728	20	
	Houghton Lake	Richardson et al. 1978	5.1		19	4						
	Poor Fen	Vitt et al. 1992	5.4	48-60	6-8	3-4	2		5-10	13-15	98-580	umho
	Sphagnum Rich Treed	Sims et al. 1982	5.4	52**	7	1	4	1200			208	ORG-N
	Poor Fen	Vitt and Chee 1990	4.5-4.8	15-28	2	0.4	0.4	985	8	22	10-20	
	Oligotrophic	Zollet and Johnson 1987	4.7	52	1	0.3	5				190	
	Open Fen	Nicholson 1987	3.5-5.4	37.7	3	0.8	0.9	1385	10	23	30	
	Big Run	Weiler R.L. 1985	4		0.8	0.2	0.2					
	Lakelse Mine	Starr and Leine 1988	4.5-5.4	2-3								
	Coastal, B.C.	Vitt et al. 1990	4.4-6.6	0.4-0.9	0.3-0.5	1-4		350-720**	2-4	30-90		ORG-N
Bog	North Plateau	National Wet Wor Gr 1988	4.7		2	2					200	
	Oligotrophic Treed	National Wet Wor Gr 1988	4.5	62	2	0.8	2				170	
	Blank B og	National Wet Wor Gr 1988	4.4		4						200	
	Bog	Vitt et al. 1992	3.9	36-39	3-4	0-1	1-2		6-11	17-18	42-140	
	Marine Lakes	Nicholson 1987	3.6-4.5	11	2	0.5	0.8	1417	9	28	260	
	High Moor	Yellmooy and Yellmooy 1973	3.6-5.2		2	0.5	0.8		545	140		
	Lakelse Mine	Starr and Leine 1988	3.6-3.9		1-2			850-870**	1-3	20-130		ORG-N
	Coastal, B.C.	Vitt et al. 1990	4.1-4.8		<0.1	0.2-1.6	0.5-1.4					

wetlands at Elk Island National Park (Gorham *et al.* 1984) compared to peatlands located in more remote areas of North America. Calcium concentrations in the moderate-rich fens average 33 mg/l as compared to a range of 3-22 mg/l found elsewhere (Table 8). The poor fens at Elk Island have lower pH values compared to similar wetlands published in the literature (Table 8), however specific conductivity is higher. Calcium concentrations are variable but can be high (7 mg/l, Table 2), as well as magnesium (2 mg/l) and sodium (4 mg/l). This may be due to additional inputs from windblown soil of the largely agricultural surrounding area. Specific conductivity and pH values in bogs at EINP are most similar to bogs at Mariana Lakes, Alberta, and high moor bogs from the Soviet Union (Yefimov and Yefimova 1973). Calcium, magnesium, and sodium concentrations of these bogs are also slightly enriched compared to other reported sites (Table 8).

Organic nitrogen values vary in Elk Island marsh wetlands from 2013-6384 $\mu\text{g/l}$. This is predominantly soluble particulate N, which is exuded from vegetation or is a product of decomposition (Wetzel 1983). These values are much higher than TKN-N concentrations (1383 $\mu\text{g/l}$) found in a bullrush marsh in Manitoba (Carpeau *et al.* unpublished). High organic N in the saline marsh at Elk Island maybe due to high rates of decomposition and concentration of the decompositional products by evapotranspiration. Bullrush marshes are much closer in structure and chemistry to open water systems than closed stand *Carex* dominated marshes studied in Elk Island. Hypereutrophic lakes normally have $> 1200 \mu\text{g/l}$ organic N in the

epilimnion, and anaerobic hypolimnions can have as high as 10 mg/l (10,000 µg/l) organic nitrogen (Wetzel 1983).

Ammonium and nitrate concentrations in swamps at EINP are comparable to those reported in Table 8. As combined inorganic nitrogen, these values range from 30-153 µg/l; less than levels commonly found in ultra oligotrophic lakes (Wetzel 1983). This indicates that the wetland plants are either rapidly assimilating or adsorbing any available inorganic nitrogen in these wetlands, or very little organic N is being mineralized to ammonium. Fens are also extremely low in nitrate and ammonium (Table 8). Bogs often contain much higher ammonium levels than fens (Table 8). Nitrification of ammonium is inhibited by dissolved organic compounds, especially tannins, acidic conditions (Wetzel 1983), and phosphorus deficiency (Damman 1988). Any or all of these could contribute to high ammonium levels in bogs. The moisture regime, as influenced by the depth of the water table will influence the amount of N mineralization and ammonium which accumulates (Richardson *et al.* 1978). Loss of ammonium through denitrification is often inhibited or occurs at very low levels in wetlands (Hemond 1983, Urban and Bayley 1988). Inhibited denitrification plus rapid uptake by plants (Urban and Eisenreich 1988, Urban and Bayley 1988) results in undetectable or very low levels of NO₃ in bogs (Gorham *et al.* 1984).

Total phosphorus is a measure of soluble inorganic and organic P, and particulate P. The majority of total phosphorus is in an organic phase as particulate P (Wetzel 1983), and is therefore a product of decomposition. Hypereutrophic lakes contain greater than 100 µg/l of total phosphorus in the surface waters (Wetzel 1983). Surface

water in the marshes at EINP have total phosphorus concentrations ranging from 73-250 $\mu\text{g/l}$, which agrees with data acquired by Vitt *et. al* (1992). Published data (Table 8) indicates that swamps can contain anywhere from 100-580 $\mu\text{g/l}$ TP in their surface waters, while values at EINP range from 221-3155 $\mu\text{g/l}$. Total phosphorus in fens has been reported ranging from undetectable to 466 $\mu\text{g/l}$ (Table 8). Data collected from the fens at EINP generally agree with these values. Low phosphorus measurements in fen Groups 4 and 10 are related to low amounts of organic N, suggesting decomposition is limited at these sites. Both bog groups had higher total phosphorus concentrations in interstitial waters than reported from other authors (Table 8). Nitrogen and phosphorus mineralization was found by Verhoeven *et al.* (1990) to be higher in ombrotrophic bogs with a *Sphagnum* cover than in minerotrophic fens. Low water tables may be contributing to faster decomposition in the bogs at Elk Island.

The distribution of wetland species along pH and ionic gradients has been well documented for peatlands (Sjörs 1950, 1952, Vitt *et al.* 1975, Vitt and Slack 1975, Karlin and Bliss 1984, Glaser 1983, Vitt and Bayley 1984, Malmer 1986, Gignac 1987, Starr and Laine 1988, Chee and Vitt 1989, Wassen *et al.* 1989, Gignac *et al.* 1991(a), Gignac *et al.* 1991(b), Glaser *et al.* 1990, Nicholson and Vitt 1990, Renato 1990, Vitt *et al.* 1990), and lakes (Jackson and Charles 1988, Arts 1990). Species distribution is a result of habitat requirements of which bryophytes are known to be particularly sensitive (Vitt and Slack 1984, Andrus 1986, Janssens and Glaser 1986) and the ability of plants to modify their surrounding environment (Andrus 1986, Gorham *et al.* 1984, Glime *et al.* 1982, Kangas 1990). Examples of

environmental modification by plants are acidification (Gorham *et al.* 1987), organic matter build up, and physical resistance to water movement (Vitt and Kuhry 1992).

Wetland species will respond to environmental gradients by producing wetland communities that reflect regional climate, local hydrology, water chemistry, basin morphology, and disturbance. The strongest gradient controlling species distribution in Elk Island National Park appears to be an ionic gradient reflected in cation concentrations, pH, and corrected specific conductivity. Superimposed on this gradient is a nutrient gradient influenced by variations in nitrogen and phosphorus levels. Ionic gradients are controlled most strongly by water source, flow rates, and precipitation, while nutrient levels are controlled by chemical cycles within the peatland. Moisture conditions and redox potentials also play a significant role in regulating available nutrients in wetlands.

The ionic gradient in wetlands at Elk Island National Park is largely driven by the water source (ombrotrophic vs. minerotrophic) and fluctuating water tables. Low elevational shore marshes are closely tied to adjacent aquatic environments with the highest number of inlet and outlet streams, and are subject to periodic drawdown and flooding. The maintenance of species diversity and deep emergent vegetation in these sites is often dependent upon cycles of drought and flooding (Keddy and Reznicek 1986, Mallik and Wein 1986, Greening and Gerritsen 1987, Welling *et al.* 1988(a), 1988(b), Wilcox and Meeker 1991). High specific conductivities approaching those of brackish waters is indicative of semi-permanence (Stewart and Kantrud 1972) and is characteristic of the

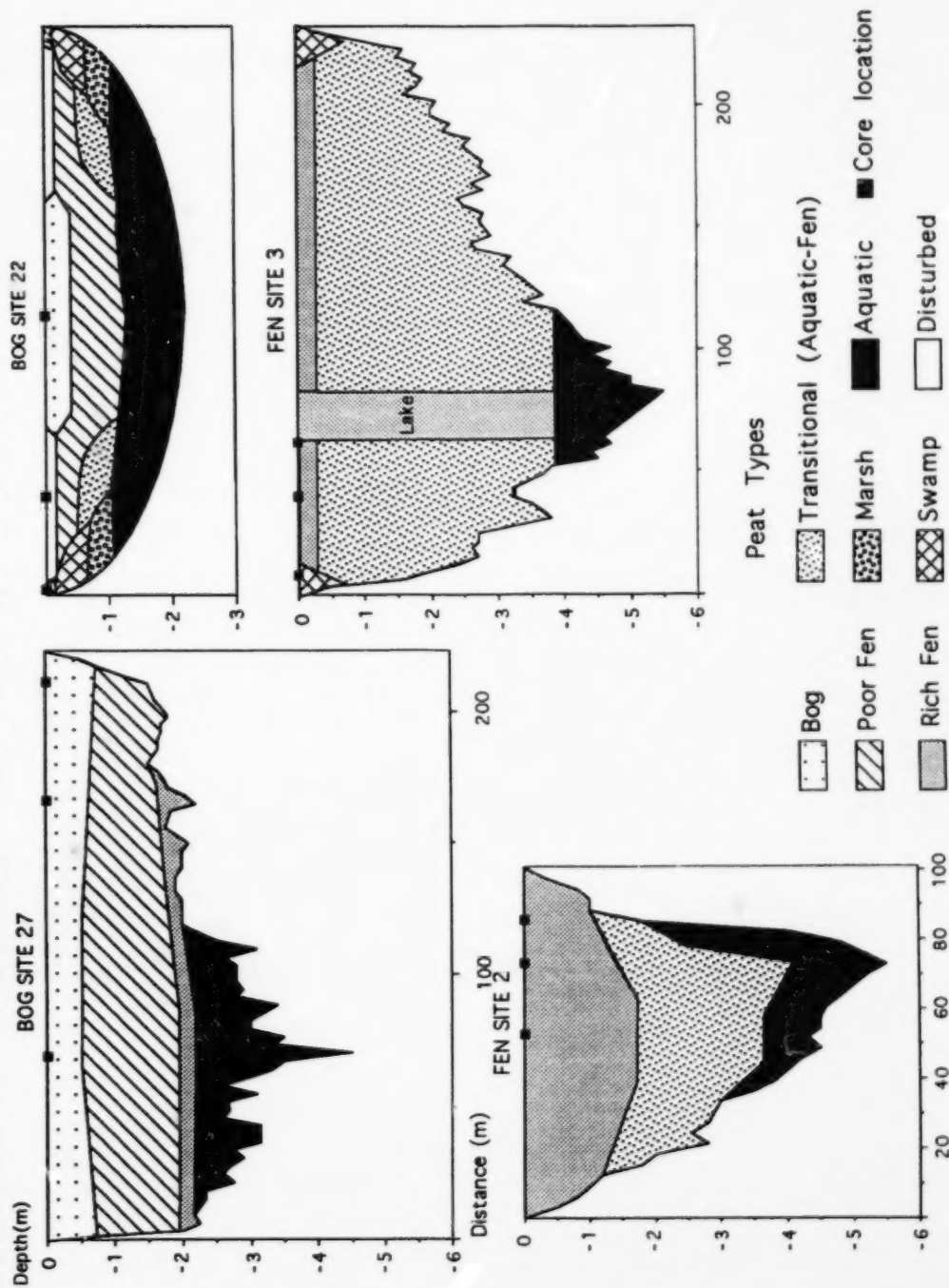
marshes at Elk Island National Park. Swamps are formed at intermediate elevations usually at the outermost fringe of the wetlands at Elk Island National Park. Cation sources for these wetlands include overland flow and underlying soils. Groundwater seepage is considered to be a major factor for swamp development, but data presented here suggest that one group of swamps at Elk Island is a product of stagnant water combined with either a shallow water table or periodic drawdown. Fens form at intermediate elevations where shallow water tables intersect the soil surface along low sloping gradients. Surface and groundwater slowly seeps across these peatlands. Basin fens receive intermediate overland flow in terms of the number of inlet and outlet streams. They usually are formed as floating mats that extend out from the basin edge into open waters. Mineral additions to the floating mats are minimized by the ability of the mat to float with rising water levels. Bogs mostly form at Elk Island National Park at the highest elevations in isolated perched depressional basins. Overland flow is minimal and water movement is restricted to one or two inlet and outlet streams.

Peatland Development

Site 2

Three phases can be recognized from the profiles of all three cores at site 2 (Fig. 31). The first phase is an aquatic phase dominated by aquatic plant and seed remains. *Carex* and bryophytes can be found in this phase in low abundance in all three cores. The presence of

Figure 31
ELK ISLAND NATIONAL PARK SUMMARY PEATLAND DEVELOPMENTAL PROFILES



mineral grains and a high percentage of detritus indicates that mixing and a high rate of decomposition occurred at the time of deposition, however, a permanent water body is suggested by the presence of *Amnicola limosa* and *Menetus cooperi* shells (Clarke 1981). The aquatic phase is followed by a transitional phase that contains both aquatic and semi-terrestrial plants. The extant phase is a moderate-rich fen community. Bryophyte species are predominant along with *Carex*, *Betula* and herbaceous material. This site has progressed from a pond-marsh where a shallow open body of water containing floating and emergent aquatic plants has slowly changed into a bryophyte/*Carex* dominated moderate-rich fen.

Site 3

The deeper portion of this basin was first occupied exclusively by aquatic plants (Fig. 31). A permanent water body is suggested by the presence of *Valvata sincera helicoidea*, a gill breathing snail (Clarke 1981). Gradually, a floating mat developed producing a large section of transitional peat. At present this floating mat is a moderate-rich fen comprised of bryophytes, sedges, and shrubs. The outer edge of this peatland flooded much later (Table 3) and initial sediments were comprised of both aquatic and semi-terrestrial plants. Close proximity to upland vegetation is indicated by the presence of woody elements, *Betula*, and *Alnus* nutlets in the basal peats. A high percentage of woody elements in the upper section of the edge profile suggests that a woody swamp developed following the earlier aquatic phase.

Site 22

Mineral soil mixed frequently with the basal aquatic sediments at site 22, and snail shells indicative of permanent water bodies are lacking (Fig. 31). *Sphagnum* fossils are present in low amounts at the base indicating that it was present in some quantity nearby. A transitional community occurred for some time, combining aquatic species with *Carex* and minor percentages of *Sphagnum* and *Drepanocladus*. In the center of the peatland, the aquatic community was quickly replaced by a *Sphagnum* dominated poor fen. This was followed by a *Sphagnum* dominated bog community, devoid of *Carex*, herbs and brown moss species. The midway core does not develop this bog phase, and the poor fen stage lasts until a few centimeters below the peat surface. Disturbance by beavers has reflooded this site altering the top few centimeters of this core to a herbaceous dominated community. Basal sediments at the edge of this site are aquatic indicating that this basin expanded in size due to increased water levels. These sediments change into a transitional assemblage consisting of aquatic species, Gramineae, *Betula*, and *Alnus*. The next phase has high percentages of woody elements, *Betula* and *Alnus* indicating that a marshy swamp developed. Surface peats show a reduction in woody elements, *Betula* and *Alnus*, and a dramatic increase in herbaceous material and *Carex* suggesting the flooding has created wetter conditions that killed the shrubs and trees previously surrounding the peatland. The profiles at this site suggest that this peatland was developing into a *Sphagnum* dominated bog, but was arrested at the margins by disturbance due to flooding. Lagg

development had progressed from an aquatic community to a marshy swamp community before disturbance occurred.

Site 27

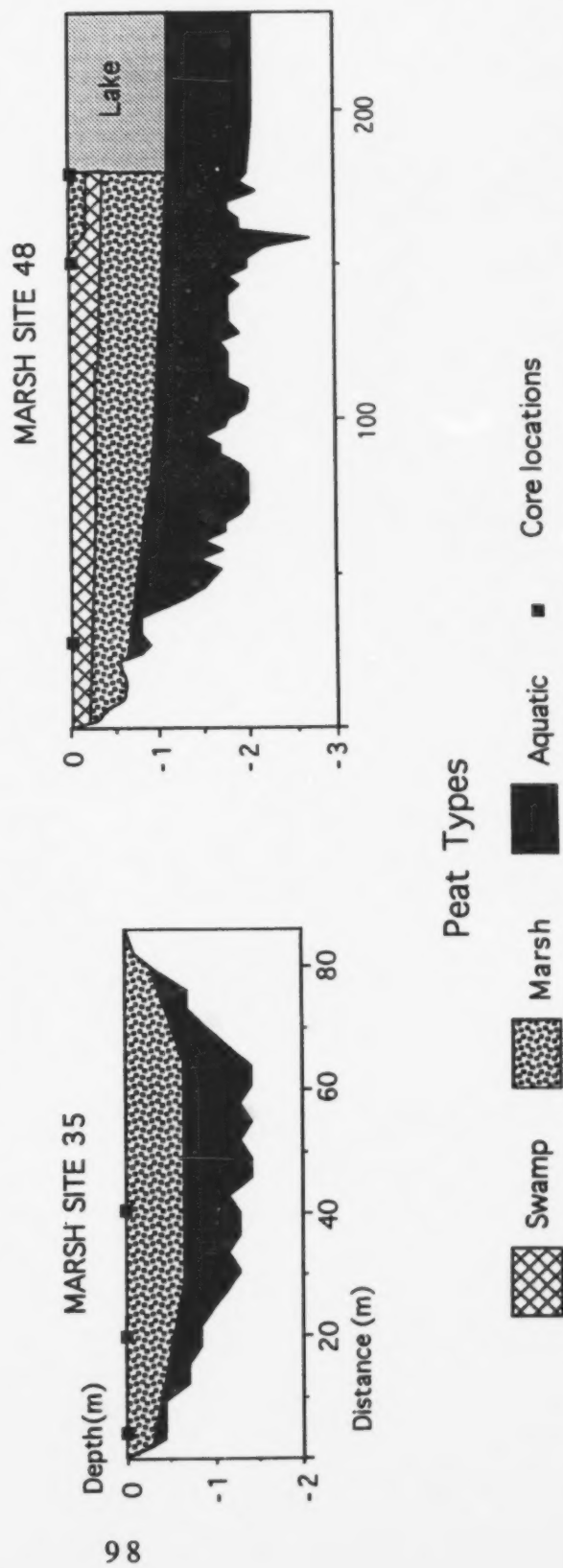
Four phases can be seen in the profile of site 27 (Fig. 31). The first stage is predominantly aquatic but there is some evidence of terrestrial plants. Snail shells indicative of permanent water bodies are absent from the basal sediments at this site (Clarke 1981). A bryophyte dominated moderate-rich fen community quickly replaced the aquatic phase, followed by a *Sphagnum* dominated poor fen. Succession has terminated with a bog community dominated by a limited number of *Sphagnum* species including *Sphagnum fuscum*. Expansion of this basin did not occur due to flooding. Basal sediments of the edge core are more terrestrial in nature and indicate that basin expansion occurred via paludification. *Sphagnum* dominated bog vegetation was attained quickly after the earlier sedge dominated poor fen community.

Site 35

This shallow basin expanded in size due to flooding. The earliest peat forming communities are aquatic (Fig. 32), but lack any gastropod shells that indicate a permanent water body (Clarke 1981). Low abundance of woody elements and high abundance of *Carex* fossils indicate that the upper peats of these cores are derived from marsh vegetation that lacked significant shrub cover.

Figure 32

ELK ISLAND NATIONAL PARK SUMMARY PEATLAND DEVELOPMENTAL PROFILES



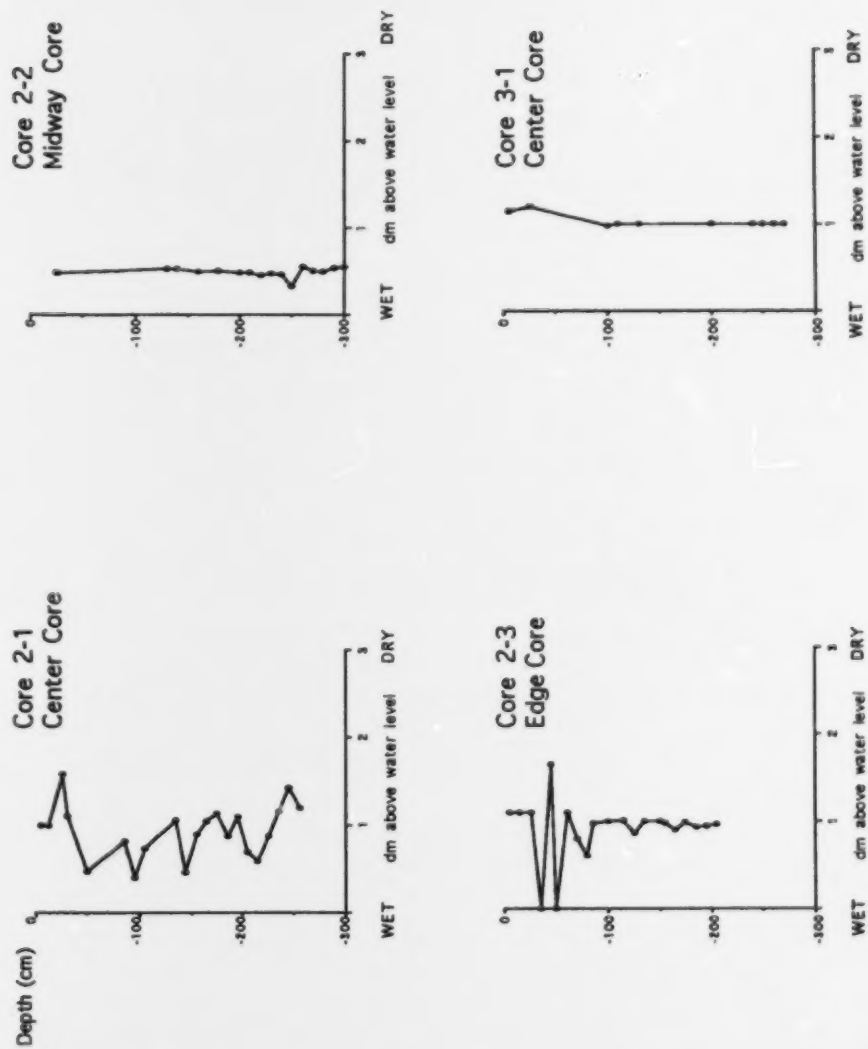
Site 48

Site 48 is presently situated adjacent to one of the larger lakes in Elk Island National Park. Lowered lake levels at the time of peat initiation are indicated by the presence of *Ruppia maritima* and *Zannichella palustris*. These species are normally found only in more saline conditions (Stewart and Kantrud 1972, Moss 1983, Husband and Hickman 1985). Lake levels rose, expanding the basin of the lake flooding the edges and initiating peat development. A permanent water body is suggested at the time of initiation by the presence of *Amnicola limosa* and *Menetus cooperi* shells (Clarke 1981). At both cores (Fig. 32), aquatic communities gave way to *Carex* dominated marsh communities. This was followed by the invasion of woody shrubs and trees, to produce a swamp community. Lake levels rose a second time killing the tree and shrub vegetation at center core. Recently a *Typha* mat has developed around the lake, as evidenced by the sudden rise of herbaceous material in the top 5 cm of the core. Rising lake levels have appeared insufficient to disturb the edge core, which is located some distance from the present shoreline.

Paleoenvironmental Reconstructions

Inferred height above water table profiles in the fens at Elk Island (Fig. 33) indicate conditions have been quite variable in the past. Cores 2-2 and 3-1, suggest a relatively stable existence, where as 2-1 and 2-3, have experienced greater water table fluctuations. Surface height above water table conditions are similar to initial inferred height above water table, indicating that the floating mats

FIGURE 33
PALEOENVIRONMENTAL RECONSTRUCTIONS
Inferred Height Above the Water Table - Fen Profiles



have not stabilized. Surface vegetation is still subjected to fluctuations in the local groundwater hydrology. Inferred height above water table profiles of the bogs at Elk Island National Park indicate that the surface became gradually drier at site 27 (Fig. 34). Surface vegetation here has been removed from groundwater influences and is no longer subjected to local flooding. Surface conditions have become significantly wetter at core 22-2 in the top 40 cm. This reflects the recent flooding which has occurred at this core.

Inferred pH profiles in the fens at Elk Island National Park (Fig. 35) indicates that acidification has not progressed significantly at these sites. Inferred surface pH values are similar to inferred basal pH values. Inferred pH profiles of the bogs (Fig. 36) demonstrates that acidification from basal pH values of 5.5 to 4.4 (Core 27-1) and 4.8 to 4.2 (Core 22-1) has taken place with an abrupt change occurring as the peatland changes from moderate-rich fen vegetation into poor fen. Disturbance of the profile at site 22 due to flooding is apparent in the top 40 cm.

The paleoenvironmental reconstructions of inferred pH of site 27 are similar to reconstructions of bogs from Saskatchewan and Manitoba (Kuhry *et al.* 1992). At Mariana Lakes (Nicholson and Vitt 1990) inferred pH curves gradually declined during the transition from fen to bog. Paleoenvironmental fen curves at EINP are similar to Mariana Lakes in that fluctuations have occurred in the past but no general trends or net changes are evident.

FIGURE 34
PALEOENVIRONMENTAL RECONSTRUCTIONS
Inferred Height Above the Water Table - Bog Profiles

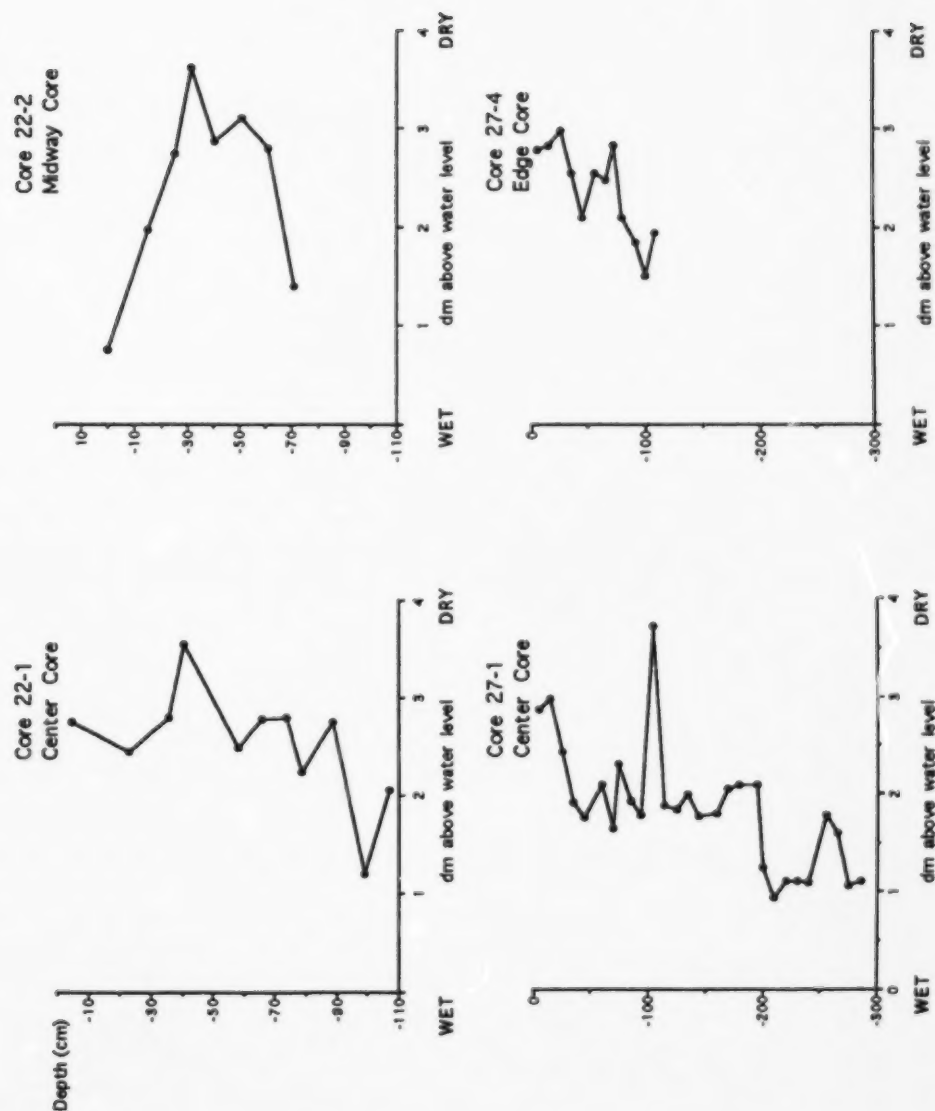


FIGURE 35
PALEOENVIRONMENTAL RECONSTRUCTIONS
Inferred pH Profiles - Fens

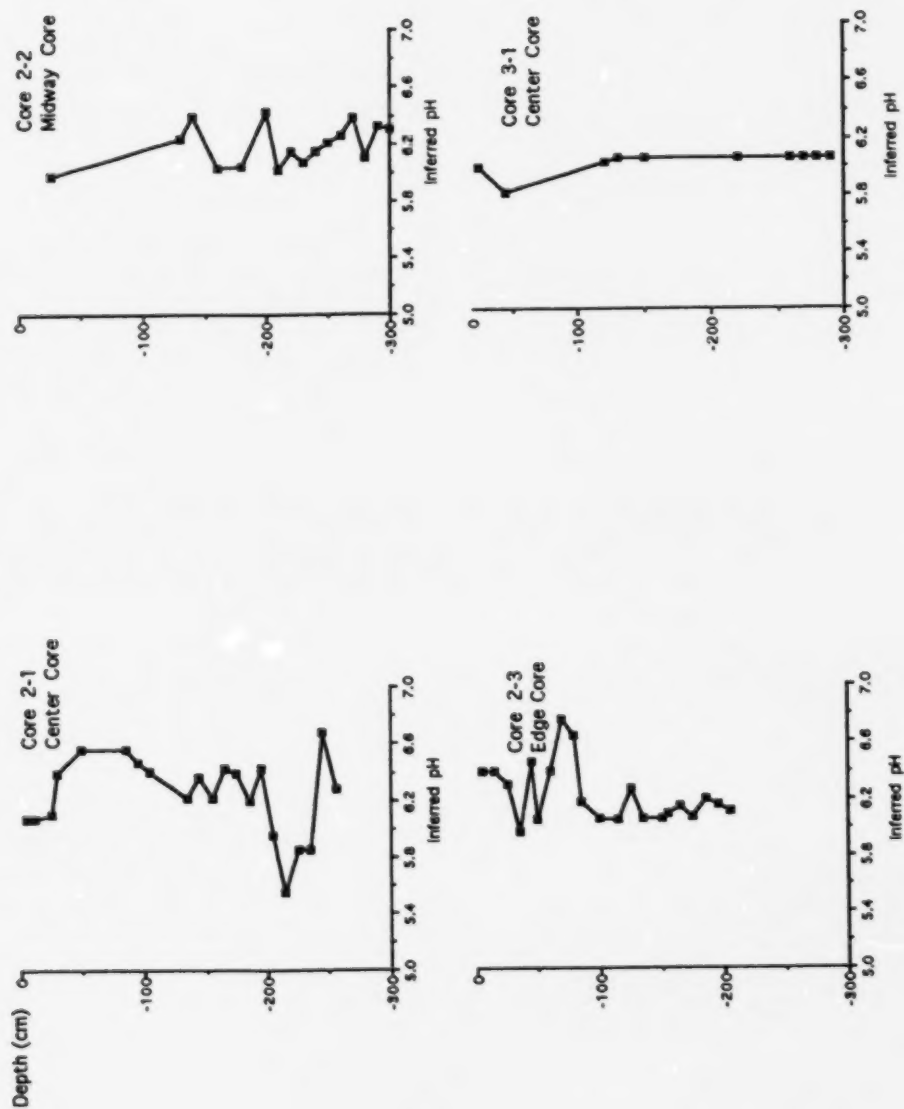
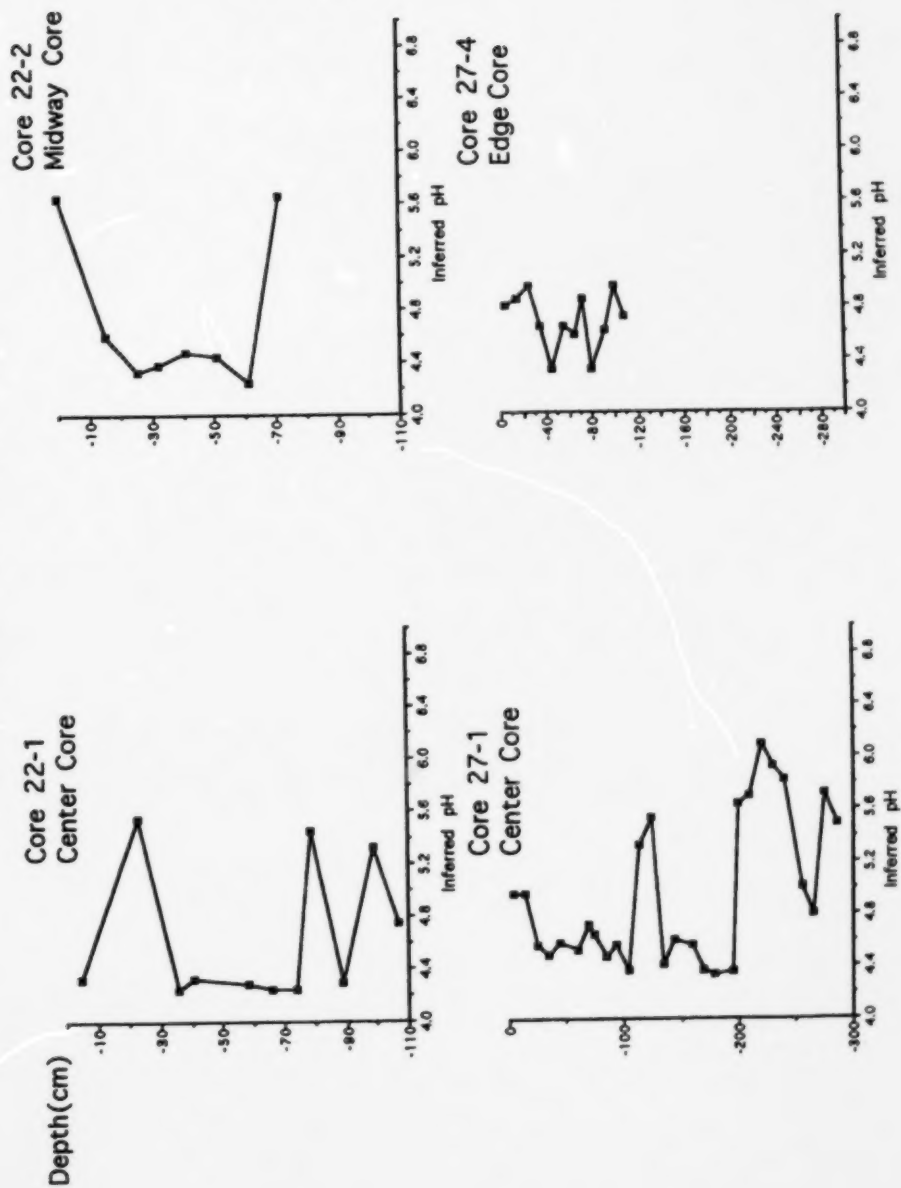


FIGURE 36
PALEOENVIRONMENTAL RECONSTRUCTIONS
Inferred pH Profiles - Bogs



Comparisons to Published Literature

In an effort to study climax successional communities and undertake pollen studies, paleoecologists have overwhelmingly directed their attention upon *Sphagnum* bogs. Few stratigraphic studies have been undertaken in fens (Zoltai and Johnson 1985, Janssen 1988, Janssens *et al.* 1992, Kubiw *et al.* 1989, Nicholson and Vitt 1990), marshes and swamps (National Wetlands Working Group 1988).

A summary of several published peat profiles for terrestrialized bogs is presented in Table 9. At each site, basal macrofossils of submerged aquatic plants or emergent marsh vegetation were identified. Common aquatic plants in Table 9 are *Chara*, *Potamogeton*, *Sparganium*, *Scirpus*, *Eleocharis*, *Myriophyllum*, *Najas*, *Nuphar*, *Nymphaeaceae*, and *Pediastrum*. Most of these sites also contain bryophyte remains such as *Sphagnum cuspidatum*, *S. subsecundum*, *S. warnstorffii*, *Drepanocladus revolvens*, *D. vernicosus*, *Scorpidium scorpioides*, *Calliergon trifarium*, and *C. giganteum*. Important early colonizers of the bogs at Elk Island National Park are the aquatic plants *Myriophyllum*, *Ceratophyllum*, *Potamogeton*, *Chara*, *Typha*, *Najas*, and *Rumex*. Bryophytes of significance are *Sphagnum warnstorffii*, *S. angustifolium*, *S. magellanicum*, *S. squarrosum*, *Drepanocladus exannulatus*, *D. aduncus*, and *Calliergon stramineum*. Peat profiles in Table 9 that contain bryophyte fossils in the basal aquatic sediments tend to develop directly into fens. Alternatively a section of pure sedge peat can occur directly after aquatic sediments before any bryophytic remains become significant

Table 9

Gypsumville Manitoba 51°46'N 98°30'W Kubry et al. 1992	La Ronge Saskatchewan 54°58'N 109°16'W Kubry et al. 1992	Porcupine Mount. Manitoba 52°31'N 101°15'W H. Nichols 1969	Gleason Michigan 45°33'N 84°42'W Miller and Putyama 1987	Point Ecuminae New Brunswick 47°04'N 69°50'W Warner et al. 1991(b)	Trevartha Ontario Warner 1989	Hook Lake Wisconsin 43°15'N 89°30'W Winkler 1988	Lake 16 Michigan 46°N 84°W Putyama and Miller 1986	Old Crow Yukon 68°N 139°30'W Ovenden 1985	TUSOCK TUNDRA
FORESTED BOG	FORESTED BOG	FORESTED BOG	OLIGOTROPHIC BOG	BOG	BOG	SPHAGNUM BOG	MAT PEAT		
Picea	Picea	Sphagnum	Sph. recurvum	Kalmia	Sphagnum	Sphagnum	Sphagnum	Picea	
Sph. fuscum	Sph. fuscum	Ericales	Chamaedaphne	Sphagnum fuscum	Carex		Carex	Rubus Chamemorus	
Polytrichum str.	Sph. magellanicum						Ericales	Sph. Sect. Cuspidata	
Poblia nutans	Polytrichum str.							Sph. Sect. Acutifolia	
FORESTED FEN	FORESTED FEN	BOG							
Sph. warnstorfii	Larix laricina	Sphagnum							
Tomenth. nitens	Picea								
Aulacom. peliastre	Campylopus stellat.								
Leodum groenland.	Tomenth. nitens								
Larix laricina	Sph. fuscum								
Picea									
richardsonii									
WET RICH FEN	FORESTED RICH FEN	POOR FEN	RICH FEN	POOR FEN	POOR FEN	EMERGENT POND	SEDGE PEAT	MARSH/SEDGE MEAD.	
Carex	Larix laricina	Carex	Scorpidium scorp.	Andromeda	Chamaedaphne	Typha	Scorpidium scorp.	Carex aquatilis	
Drep. aduncus	Carex		Calliergon trif.	Chamaedaphne	Myrica gale		Calliergon trif.	Carex rostrata	
Calliergon gigant.	Calliergon gigant.		Drep. aduncus	Carex	Mesanthus trif.		Drep. revolvens	Hippuris	
	Drep. lapponicus		Sph. cuspidatum				Mesita triquetra	Eleocharis	
			Sph. subsecundum					Ranunculus scleratus	
			Sph. magellanicum						
SHALLOW POND	MARSH	MARSH		SHALLOW POND	SHALLOW POND	FEN	EMERGENT MARSH	GYTTJA	LAKE
Chera	Typha	Carex		Potamogeton	Potamogeton	Carex	Najas flexilis		Chera
Potamogeton	Carex			Potamogeton	Potamogeton	Carex	Potamogeton		Potamogeton
Scleropus	Rumex			Eleocharis	Drep. revolvens	Diatoms	Potamogeton		Myriophyllum
Sph. warnstorfii	Larix laricina			Scirpus	Drep. revolvens	Nyphaceae	Nuphar		Nuphar
	Calliergon gigant.			Carex	Scorpidium scorp.	sponges	Ranunculus aquatilis		Ranunculus aquatilis
	Sph. warnstorfii			Glyceria canadensis	Equisetum	mooses			
POND									
Typha									
Sperganium									
Potamogeton									
Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates	Basal Radiocarbon Dates
4130	5020	6700	8000	10,900	7,000	13,000	6210		

Basal radiocarbon dates and peat stratigraphic profiles of published terrestrialized bogs in North America. Major successional stages and macrofossil species are indicated.

(Griffin 1977, Tallis 1983, Ovenden 1985, Giller and Wheeler 1988, Janssen 1988, Wilcox and Simonin 1988). The centre core at site 22 (core 22-1) at Elk Island National Park has a similar stratigraphy, about 20 cm of pure sedge peat was laid down before bryophytes became present in the sediments. Floating sedge mats which can act as the primary colonizers of open water bodies do not occur in this area, but are prevalent in England and Scotland (Tallis 1983). The occurrence of vascular plants without a significant bryophyte cover indicates a period of fluctuating water tables and drawdown conditions, two contributing factors preventing bryophyte dominance in present day marshes. Once fen vegetation becomes established, autogenic factors contribute to increased oligotrophy, acidification and peat build up, creating drier, less cation-rich surface conditions. *Sphagnum* often invades, first creating *Sphagnum* dominated poor fens and bogs (Nichols 1969, Ovenden 1985, Winkler 1988, Warner 1989). This is followed by the establishment of shrubs and trees (Nichols 1969, Ovenden 1985), creating forested bogs. Often the two processes occur together and major successional changes occur in the peat profiles (Futyma and Miller 1986, Miller and Futyma 1987, Warner *et al.* 1991b).

The peat profiles of terrestrialized bogs at Elk Island National Park and those presented in Table 9, outline a hatched arch successional series of open pond/marsh - fen - bog. Within each profile there is evidence of increasing oligotrophy, acidification, and surface desiccation. This pattern of terrestrialization is supported by Vitt and Kuhry (1992), Warner *et al.* (1991b), National Wetlands Working Group (1988) and in a model presented for boreal peatland

development by Osvald (1970) outlining general mire development in the Edmonton, Alberta area. Similar developmental patterns have been recorded in the Red Lake peatland (Janssens 1988), in south Sweden (Mörnsjö 1968), in Finland (Warén 1924), and in Colombia (Kuhry 1988).

Fens are often considered to be a seral stage leading to bog development, although many fens have been stable ecosystems for thousands of years (Zoltai and Johnson 1985, Kubiw *et al.* 1989, Nicholson and Vitt 1990). Macrofossil community changes in fens have been attributed to autogenic change, climatic factors, and hydrology. Surface patterning is considered to be a secondary event in fens (Foster and Fritz 1987, Kubiw *et al.* 1989), as well as in bogs (Foster *et al.* 1988). A stratigraphic profile from a basin fen near Jan Lake, Saskatchewan (National Wetland Working Group 1988) indicates that this wetland began as a pond that was filled in by an open fen. Subsequently the open *Carex* fen was invaded by shrubs and tamarack trees. *Sphagnum* became predominant in the macrofossil record only in the last 28 cm. Similarly the development of shrubby vegetation followed by treed vegetation did not occur in a Rocky Mountain House fen until relatively recently (after 2350 yBP) (Zoltai and Johnson 1985). Fens at Elk Island National Park have developed very slowly from shallow open water basins. Aquatic plant remains are present in the peat sediments until the last 100 cm. It appears that development took place via the gradual encroachment of floating mats over the open water bodies. Bryophytes, particularly *Drepanocladus aduncus*, *D. crassicosatus*, *Calliergon* spp., and *Meesia* spp., were important components of the mat. *Carex* and shrubs are

more recent additions to the peat forming communities. Inferred pH and water table reconstructions indicate that some of the peatlands in EINP are relatively stable, with autogenic processes occurring slowly, or arrested due to associated allogenic factors.

Swamp profiles from eastern Canada (Glooschenko and Grondin 1988), particularly a profile from Puslinch township, indicate that swamps can evolve from shallow marshes or ponds. Aquatic peats developed into sedge peats, followed by woody swamp peats. An identical series of seral stages has been described from Woody Bog, a *Populus/Salix* dominated wetland located just south of Grand Prairie, Alberta (A. Beaudoin, personal communication). Aquatic peats were replaced by sedge peats, followed by swamp peats. A similar "bog sere" was reported in Wisconsin by Frolik (Tallis 1983). He described a hydrosere consisting of aquatic plants followed by successive zones of *Typha/Scirpus/Phragmites*, *Carex* meadow, scrub (*Salix*, *Betula*), and finally *Populus tremuloides*. In the two undisturbed swamp profiles at Elk Island Park these seral sequences can also be seen. Bryophytes are rarely present in the sediments, and *Carex* dominated vegetation establishes early. This indicates that drought and drawdown play a significant role in the establishment of these communities.

Peat Chemistry Comparison to Published Literature

As compared to other reported publications on peat chemistry (Table 10), the bogs at Elk Island are high in ash (6%), sulphur and potassium, but average in calcium, magnesium, manganese, iron, and

Table 10

A	B	C	D	E	F	G	H	I	J	K	L
Welland	Description	Reference	% Ash	Ca mg/kg	Mg mg/kg	Mn mg/kg	Fe mg/kg	Na mg/kg	S mg/kg	P mg/kg	K mg/kg
1	Welland										
2	Bog	Can. Wellands W. Group, 1968	2.0-4.7	800-9300	130-2220	50-160	75-1280	60-120		230-620	170-780
3	Bog	Can. Wellands W. Group, 1968	1.8	1200	200	62	300	114		900	200
4	Boreal-Ontario	Can. Wellands W. Group, 1968	2.6	1800	210		450			380	380
5	Boreal, Ontario	Can. Wellands W. Group, 1968	1.4-4.5	1737-8760	1148-4819		435-1369		566-2478	183-620	
6	Flat Bog, Manitoba	Can. Wellands W. Group, 1968	1.4-5.2	1213-10128	829-4373		436-1581		293-3502	116-451	
7	Flat Bog, Manitoba	Can. Wellands W. Group, 1968	3-10.7	1693-20230	708-2674		428-2516		310-784	313-825	
8	Basin Bog, Manitoba	Can. Wellands W. Group, 1968	4.1	4983	442	30	1052	119	676	288	126
9	Marine Lake, Alberta	Nicholson, 1969	4.8-5.4	5500	<1260	<160	<2500		<375	600	1260-2000
10	Open Bog, Northwest Ont.	Riley and Michaud, 1969	4.7-5.3	3000-4500	1250	<160	1000-4000	500	<375	<800	<2500
11	Treed Bog, NW Ont.	Riley and Michaud, 1969	2.0-4.0	1000-5000	1100-1700					380-1620	90-2190
12	Sphagnum, Coom Rigg Moss	Chapman, 1964	3.1-4.1	700-3500				370-980		380-800	0-3000
13	Bog Plateau	Damman and Dowhan, 1961	<4	4000-16000	6000-10000	0-100	250-3000	200-600			
14	Scandinavian	Damman, 1978									
15	Bog peats above water table	Zotzel and Johnson, 1965		4126	814		724	92	831		567
16	Sphagnum	Pakarinen and Gorham, 1963	3.0-7.0	1194-3873	356-788	8-541	601-2661	19-121	540-2910	246-651	242-2103
17	Sphagnum	Pakarinen, 1967	1.2-3.1	1480-2700	300-826	175-950	200-1200	80-475			
18	Oligotrophic	Zotzel and Johnson, 1967		2724	1006	146	1352	119	802	522	1276
19											
20	Poor Fen	Can. Wellands W. Group, 1968	3.9	2400	390	120	1410	130		700	570
21	Boreal, Ontario	Can. Wellands W. Group, 1968	3.3	2400	400		1400			700	600
22	Boreal, Ontario	Can. Wellands W. Group, 1968	3.3	6420	530	30	480	142		690	650
23		Damman and Dowhan, 1961	3.5	800-1600	700-1100			340-1010		480-910	230-3170
24		Damman and Dowhan, 1961	8.7	800-300-1300				290-1010		840-960	340-3310
25	Oligotrophic	Zotzel and Johnson, 1967		3298	918	154	1992	237	890	736	1606
26	Fen	Can. Wellands W. Group, 1968	5.2-11.7	9765-21531	4810-7585		629-6018		6905-16377	348-442	
27	Fen	Can. Wellands W. Group, 1968	5-16.7	4749-18074	2817-8255		582-3640		1701-14312	435-1758	
28	Flat Bog, Manitoba	Can. Wellands W. Group, 1968	5-8.9	12012-1844	3086-4302		1472-5778		3056-4246	269-521	
29	Horizontal Fen, Alberta	Can. Wellands W. Group, 1968	8.2-15.1	9927-36602	4748-7304		631-7657		1814-22936	618-1508	
30	Horizontal Fen, Alberta	Can. Wellands W. Group, 1968	10.1-54.3	37341-18939	1445-3257		4187-14533		1841-4875	412-570	
31	Collapsed scar, Manitoba	Can. Wellands W. Group, 1968	8-17.3	10765-29099	1802-3567		7032-17014		2708-6115	405-4378	
32	Basin Fen, Saskatchewan	Can. Wellands W. Group, 1968	5.9-11.1	7786-21576	1476-2966		1154-11084		707-2696	440-1166	
33	Marine Lake, Alberta	Nicholson, 1969	12	9332	897	102	2189	162	2255	856	341
34	Poor fen	Vitt and Chase, 1990	3-3.8	3241-4086	471-1004		366-600	124-227	1034-2679	851-1307	278-640
35	Mod-rich fen	Vitt and Chase, 1990	10-12.1	14018-17421	1791-2222		890-1074	214-737	3348-8267	1263-1916	403-1597
36	Open Fen	Riley and Michaud, 1969	7.6-10.7	18000-24000	1250-2500	100-1400	6250-11250		<375	450-700	<2500
37	Macrotrophic	Zotzel and Johnson, 1967		19380	3480	708	4116	177	1920	851	1080
38	Dystrophic	Zotzel and Johnson, 1967		6999	1803	369	3432	202	1147	812	1863
39	Treed Fen	Riley and Michaud, 1969	6.8-10.6	15000-33000	<2500	100-500	3000-6000		<375	375-750	<2000
40	Rich	Richardson et al., 1978	28.6	1330	140					90	90
41	Wagner Bog, Alberta	Can. Wellands W. Group, 1968	20.7-47.7	58273-14649	3915-5333		321-8027	445-848	3458-23880		
42	Extreme rich fen	Vitt and Chase, 1990	10.6-11.5	48780-5139	3630-4518		610-1263	188-193	9364-8788	550-632	588-703
43	Macrotrophic	Zotzel and Johnson, 1967		44657	3410	786	7677	122	2757	801	844

Table 10 continued.

A	B	C	D	E	F	G	H	I	J	K	L
	Description	Reference	% Ash	Ca mg/Kg	Mg mg/Kg	Mn mg/Kg	Fe mg/Kg	Na mg/Kg	S mg/Kg	P mg/Kg	K mg/Kg
44	Wetland										
45	Marsh	Can. Wetlands W. Group, 1988	24	8060	1200	30	4510			920	3870
46	Floating marsh/ten, Ontario	Can. Wetlands W. Group, 1988	23	15020	2260	120	9080			880	4740
47	Floating marsh/ten, Ontario	Can. Wetlands W. Group, 1988	45	8780	3530	180	12300			870	8630
48	Shallow shore, Ontario	Can. Wetlands W. Group, 1988	89								
49	Spartanium marsh, Quebec	Can. Wetlands W. Group, 1988	39	5900	3090	90	9330		950	6910	
50	Meadow marsh, Ontario	Riley and Michaud, 1989	8.2-18.9	16000-22600	3800-4250	150-2250	3500-7500		<350	450-1350	500-3500
51	Meadow Marsh	Chapman, 1964	8.0-12.0	3000-10000			2000-5000	250			100
52	Phragmites, Coom Rigg Moss	Zolal and Johnson, 1987		23120	3539	464	2596	128	3046	816	1071
53	Conifer	Can. Wetlands W. Group, 1988	7.2-7.9	21600-24200	1300-4100		2200-13100			400-600	400-7800
54	Boreal, Conifer, Cochrane, Ont	Can. Wetlands W. Group, 1988	7.9	21600	1310	110	2230	100		430	430
55	Deciduous, Ontario	Can. Wetlands W. Group, 1988	9	24100	1720	101	1850	91		30	420
56	Shrub, Timmins, Ont	Can. Wetlands W. Group, 1988	9.7-15.5	5365-14045	1654-1680		1735-3616		600-764	616-809	
57	Coniferous, Manitoba	Can. Wetlands W. Group, 1988	17								
58	Picea Mariana, Quebec	Can. Wetlands W. Group, 1988	92-55								
59	Thuja occidentalis, S. Quebec	Can. Wetlands W. Group, 1988	93.9								
60	Shore swamp, Quebec	Riley and Michaud, 1989	8.2-9.3	17000-28000	4675-5525	250-1250	2126-4900		350-700	500-1300	<3285
61	Thicket Swamp										
62	Aquatic	Can. Wetlands W. Group, 1988	43.4	16631	9494		13303			434	
63	Jan Lake, Saskatchewan	Can. Wetlands W. Group, 1988	21.5-84.4	12529-25431	1478-12402		1528-31174		1918-10042	392-526	
64											
65	Published peat chemistry (% ash, calcium, magnesium, iron, sulphur, phosphorus, potassium) from North American and select European peatlands.										
66	** Indicates publications that included a depth profile of the peat chemistry.										
67											

sodium. The poor fens are mineral rich as ash, calcium, magnesium, iron, and sulphur are high (Table 7) compared to published reports (Table 10). The peat chemistry of the moderate-rich fens is comparable to published peat chemistry data (Table 10). Rich fens tend to have much higher calcium and magnesium concentrations. Reported peat chemistry of marshes indicates that they commonly have much higher ash and phosphorus concentrations than at Elk Island, but lower calcium, sodium and sulphur. Swamp peats appear to be extremely variable in ash content ranging from 7-94% (Table 10). Other reported parameters are also extremely wide ranging. In comparison, the swamps at Elk Island are low in ash and iron, but high in calcium, magnesium, sodium, sulphur, and phosphorus. Although aquatic peats have not been adequately documented chemically, the aquatic peats at Elk Island are comparable to two sites analyzed by the National Wetlands Working Group (1988, Table 10).

Peat Profiles

Physical Profiles

Ash

In an aquatic environment, mineral grains are incorporated into the sediments through wave action and lake turnover. As peat accumulates, accessibility to mineral soil in the basin is reduced (Shotyk 1986), particularly if a floating mat is formed. The floating mat provides a physical obstruction to the circulation of lake water into the peat forming community. With the accumulation of significant quantities of peat, waterflow into the peatland becomes increasingly restricted to the lagg and major flow channels (Tallis

1983) further reducing allogenic sources of ash. The ash content of mire plants reflect both allogenic and autogenic sources.

Minerotrophic mire species contain higher quantities of mineral matter derived from the water source, than corresponding bog plants (Chapman 1964, Mörnjö 1968, Pakarinen and Tolonen 1977, Karlin and Bliss 1984, Shotyk 1986). As the peatland develops ombrotrophic conditions, incoming sources are reduced to aerially deposited forms (Gorham 1957, Damman 1978, 1986), and biogenic silica produced by the plants themselves (Shotyk 1988).

Ash profiles from terrestrialized bogs and fens have high ash contents (5-49%) in the basal aquatic sediments (Gorham 1949, Chapman 1964, Mörnjö 1968, Kubiw 1987, Nicholson 1989). Once these peatlands have developed fen peat, ash contents drop to lower levels. Ash profiles from paludified fens have high mineral contents in the lowermost samples adjacent to the mineral soil boundary, after which, ash levels remain fairly constant at levels which reflect the amount of minerotrophy in the fen (Chapman 1964, Zoltai and Johnson 1985, Kubiw 1987, National Wetland Working Group 1988, Nicholson 1989, Riley and Michaud 1989). Ash profiles from paludified bogs commonly have very low ash contents that extend until basal soil sediments are reached (Chapman 1964, Sonesson 1970, Nicholson 1989). Allogenic sources of ash to paludified bogs are either insignificant or ash is leached out from the peat soils after deposition. Ash profiles of swamps and marshes tend to have consistently high ash content, with a few profiles showing a decrease towards the peat surface (National Wetland Working Group 1988, Riley and Michaud 1989) indicating a consistent mineral source.

Ash profiles from the wetlands of Elk Island National Park are typical of terrestrialized peatlands and indicate the gradual removal of the aquatic peat forming community from lake generated sources of ash. Bog site 27 has an ash profile that is slightly different from the others; ash percentages drop very quickly from greater than 50% in the aquatic peats to 4% in the fen peats. This ash profile is typical of a paludified peatland, and suggests that during development site 27 quickly became isolated from lake generated ash sources.

Bulk Density

Bulk density is affected by the species composition of the peat, its degree of consolidation and compaction, and the state of decomposition. Undecomposed *Sphagnum fuscum* peat has a bulk density of 0.02 g/cm³ (Pakarinen and Gorham 1983) which compacts to bulk densities of 0.08 g/cm³ 60 cm below the surface. Bulk density profiles generally increase with depth (Zoltai 1991). At bogs in northwestern Ontario, surface peats averaged 0.08 g/cm³, while basal peats averaged 0.21 g/cm³ (Riley and Michaud 1989). In comparison bulk density values at the surface of bog site 27 average 0.04 g/cm³, while below surface poor fen and moderate-rich fen peats ranged from 0.1-0.25 g/cm³, and aquatic peats are generally greater than 0.2 g/cm³. At bog site 22, poor fen and transitional peats had bulk densities of 0.2 g/cm³, and aquatic peats range from 0.2-1.9 g/cm³.

Bulk density profiles of a terrestrialized fen in Ontario have surface values of 0.02 g/cm³, which increased to 0.6 g/cm³ at depth. Lower aquatic gyttja sediments had bulk density values ranging from

0.4 to 1.4 g/cm³ (Warner *et al.* 1991a). At Mariana Lakes, the terrestrialized poor fens have surface bulk densities averaging 0.1 g/cm³, gradually increasing to 0.8 g/cm³ in the aquatic peats (Nicholson 1989). At Muskiki Lake, a terrestrialized fen, bulk densities in the surface peats average approx. 0.08 g/cm³. In the aquatic peats of this section, bulk density values averaged 0.4 g/cm³ (Kubiw 1987).

Bulk density profiles from the fens at Elk Island clearly demonstrate the gradual evolution of these peatlands from aquatic ecosystems with high bulk densities in the aquatic peats and low in the fen peats, and the presence of a floating mat. Fen profiles clearly show the development of floating mats. Surface peats in the fens generally have greater than 0.2 g/cm³, quickly declining in cores 2-1, 2-2, 3-1, 3-2, to less than 0.1 g/cm³. This marks the boundary between the floating mat peat and detrital peat shed from the bottom of the floating mat that accumulated on the pond bottom. Between the two types of peat is a section of detrital peat which has not been compacted. This zone of uncompacted detrital peat is not seen in the edge cores as the mat is grounded from the time of initiation.

Swamp and marsh profiles at Elk Island show the same general trend as the bogs and fens: a gradual removal of the wetland from an aquatic ecosystem, as evidenced by a decline in bulk density values from aquatic peats to surface peats. The bulk density values given here compare favorably with values given by Riley and Michaud (1989).

Chemical Profiles

Calcium

In ombrotrophic, terrestrialized peatlands calcium profiles are highest at the base and decline gradually to $< 5,000$ mg/Kg once ombrotrophic peat begins to form (Chapman 1964, Damman 1978, National Wetland Working Group 1988, Riley and Michaud 1989). Basin fens (formed under terrestrialized conditions) have similar profiles (Kubiw 1987, National Wetlands Working Group 1988, Nicholson 1989). Paludified bogs often have calcium profiles, which are calcium poor from the beginning of peat formation (Chapman 1964, Nicholson 1989). Horizontal fens that have continuous water flow through them have straight, consistently high calcium profiles (Zoltai and Johnson 1985, National Wetlands Working Group 1988, Nicholson 1989). In marshes and swamps where high decomposition rates occur and the vegetation is dominated by vascular plants that have penetrating roots, internal recycling of minerals maintains high concentrations of calcium throughout the profile (National Wetland Working Group 1988, Riley and Michaud 1989). There is also some evidence that as peat accumulates in these wetlands, minerals become increasingly limited (Chapman 1964).

In Elk Island, the calcium profiles of the central fen cores (2-1, 2-2, 3-1, 3-2) show a terrestrialized calcium profile, with highly variable calcium levels in the aquatic peats, reduced levels through the transitional aquatic/fen peats, and stabilized values of $10,000$ mg/Kg in the moderate-rich fen peats. Core 2-3 has a very stable profile, indicating that continuously low input of calcium from overland flow occurs at this site. Core 3-3 indicates that the input of

calcium to this core has increased in the past but has decreased again following the development of swamp vegetation. Bog profiles, particularly site 27, suggest a limited calcium input into these sites, that has declined steadily during development. Edge core 22-3 has a high basal calcium concentration and a more variable profile, indicating an unstable hydrologic regime. Marsh site 35 has low basal calcium concentrations (10,000 mg/Kg) that increase towards the surface as marsh peats are encountered. Falling water levels, increased decomposition, and the recycling of calcium through vascular plant roots may all contribute to this core profile. Swamp site 48 has high calcium concentrations in the aquatic sediments that rapidly decline upon the appearance of marsh and swamp peats. In this site it appears that the lake basin has contributed significantly to the calcium levels. Once marsh vegetation became established this input was lost. With the development of a shrub layer, decomposition and internal recycling rates would increase, resulting in a rise in calcium in the surface peats.

Copper

Peatlands trap copper, due to the metal's strong affinity for organic material (Shotyk 1988). Largin *et al.* (1972) clearly demonstrated that copper concentrations in a lowmoor peatland followed a path of water movement into the mire. Concave shaped copper curves are common in the peat profiles of bogs and fens (Elomaa 1987, Riley and Michaud 1989). Surface enrichment is attributed to bioaccumulation or increased industrial pollution (Shotyk 1988). Surface concentrations of 6.3-12.7 mg/Kg were found

in a paludified ombrotrophic peatland at Abisko (Sonesson 1970), and 2 mg/Kg in surface peats of a minerotrophic fen in Finland (Elomaa 1987). Basal sediments were 6.3 mg/Kg at Abisko and 8 mg/Kg in Finland. Copper profiles in bogs of northern Ontario are concave with approx. 3.8 mg/Kg at the surface and up to 56.3 mg/Kg at the base. Fens contain 1.8-5 mg/Kg at the surface and 9-24 mg/Kg at the base. Swamps and marshes tend to have straight profiles with copper contents averaging 6-9 mg/Kg (Riley and Michaud 1989).

Fen profiles at Elk Island show some tendency towards being concave with secondary surface enrichment due to either bioaccumulation or pollution. Once again the central fen cores show strong evidence of high input during the initial aquatic stages, gradually declining once floating mats formed, limiting the minerotrophic water sources. Edge cores suggest continuous low copper inputs into these sites. Bog site 27 has a profile that indicates the initial sediments were in contact with a copper source (overland flow, lake sediments), that quickly became eliminated. Bog site 22 has a variable copper profile, indicating that initial sources were low, but copper inputs have increased at least twice during the development of this peatland. Copper profiles in the swamps and marshes are similar to their calcium profiles. Site 35 has low basal values, which peak in the aquatic/marsh sediments. Site 48 has high basal copper concentrations, which decline towards the peat surface, indicating a gradual decline has occurred in the amount of copper reaching these peats.

CONCLUSIONS

Thirteen wetland vegetation groups were found to occur in Elk Island National Park. They range from saline marshes through to forested bogs. Vegetation is predominantly controlled by waterflow and cation gradients reflected in strong correlations between vegetation, pH, corrected specific conductivity, calcium, magnesium, and sodium content of the surface waters. Nutrient (nitrogen, phosphorus) content of surface waters does not play as strong a role in determining species composition of these peatlands. Marshes, swamps, and moderate-rich fens at Elk Island National Park are circumneutral in pH, with surface waters high yet variable in calcium, magnesium, and sodium. Phosphorus levels can be high in some types of swamps. Poor fens and bogs are acidic, with less mineral ion content. Nitrate, ammonium, and phosphorus levels in bogs and poor fens can be as high or higher than the marshes, moderate-rich fens, and swamps. Elevation, number of inlet and outlet streams, water level, and hummock height all contribute to the diversity and composition of species in these wetlands.

Successional patterns in the wetlands follow hydrosereal successional sequences outlined by Walker (1970). Two definite developmental pathways can be found within the park.

1) Aquatic -- Rich Fen -- Poor Fen -- Bog -- Forested Bog

2) Aquatic -- Marsh -- Swamp

When the site is conducive to bryophyte colonization, the direction of succession is towards *Sphagnum* domination. Autogenic factors drive this successional pathway by the reduction of the ionic supply reaching the wetland. Autogenic changes evident in these wetlands are the gradual elevation of the peat surface above the local water table, acidification of the environment, and oligotrophism. Peat accretion fills in the basin and leads to consolidation of the peat. Once the peat surface is no longer subjected to excessive water level fluctuations, bryophytes and shrubs can establish. Gradually hummocks are formed where the process of acidification and oligotrophism begins. Here in small microtopographic situations bryophytes lower the pH level (Glime *et al.* 1982, Gorham *et al.* 1984, Urban *et al.* 1987) due to their cation exchange properties. Organic acid production through decomposition may also be a factor (Hemond 1980, Gorham *et al.* 1984). Elevation from the surrounding watertable promotes leaching of nutrients and isolation from ionically rich waters. Amalgamation of the hummocks and channelling of water away from elevated portions of the mire expands the region affected by these two processes and ultimately leads to the formation of bogs.

Where drawdown or watertable fluctuations prevent bryophytes from establishing, the wetland develops into a swamp. Parallel trends, due to autogenic factors operate in marsh wetlands. Peat material deposited beneath the sedge community, isolates it from one ionic source, the underlying mineral soils. The layer of peat also acts as a stabilizer of hydrologic conditions, due to it's ability to swell and store water. Elevation of the marsh community from the local watertable occurs due to the accretion of peat. As a result severe

flooding is restricted. Excessive desiccation is retarded by the ability of the peat to retain water. Consequently shrub and tree growth is promoted. If through hydrological stability the site becomes conducive to bryophyte establishment, mosses invade and the swamp community may evolve into a fen. Although this successional pattern has not been documented in Elk Island, it is well known elsewhere from the literature (Nichols 1969, Tallis 1983, Winkler 1988). It is likely the high evapotranspiration rates that occur in this region is responsible for limiting this successional pathway through limiting bryophyte establishment.

Climate is an allogenic factor that has strongly affected peatland development at Elk Island National Park. The early Holocene thermal optimum delayed peat initiation until after 6720 yBP. Early wetlands were small open ponds with *Typha* marshes occurring along the shorelines. The predominance of *Typha* marshes in early basal sediments is considered to be characteristic of southern boreal peatlands and indicative of warmer climates (Kuhry *et al.* 1992). Some wetland macrofossils present in the basal sediments normally occur in more saline conditions, perhaps a sign of higher evapotranspiration rates. *Sphagnum* invasion into the bogs at Elk Island National Park is delayed compared to more northern peatlands and is considered to be climatically induced (Kuhry *et al.* 1992). Terrestrialization has been the dominant process initiating peat development in the park. Climatic limitations and topographic relief presently prevent peatlands from extending out beyond small depressional basins. The relatively late invasion of *Sphagnum* and lack of paludified peatlands in the park indicate that climate is

continuing to exert a strong influence on the development of these sites.

Climatic factors have affected peatland development at Elk Island National Park in several ways: 1) delay of peat formation immediately following deglaciation, 2) expansion of the peatlands basins between 6600 and 2910 yBP, 3) predominance of terrestrialized peatlands in the park, and 4) late invasion of *Sphagnum* into the bogs. These climatic factors have affected all peatlands uniformly over broad geographical areas and are thus not responsible for the diversity of peatlands that exist within the park.

Edaphic factors are more site specific and affect peatlands individually. They are the physical setting of a wetland such as elevation, basin morphology and hydrology. These factors influence the starting point (terrestrialization, paludification), and control the rate of development in a given peatland.

Autogenic factors occur within individual peatlands and appear to have three directions directly affecting succession. These are acidification, elevation of the peat surface, and oligotrophism.

Development of the peatlands at EINP has been studied using peat macrofossils and physical and chemical characteristics of the peat. Using this combined approach processes driving the evolution of the peatland have been elucidated. The profiles from the fens both chemically and physically indicate that minerals were abundant in the early peat forming shallow pond stage of these peatlands. Surface runoff, wave action, and lake turnover renewed the minerotrophic environment. Floating mats formed and the peat forming plant communities became increasingly more isolated from mineral sources.

Bog profiles indicate that a similar process occurred. An influx of copper, calcium and mineral matter occurred in site 22, after the infilling process began, while site 27 experienced rapid and almost complete removal from minerotrophic sources. Terrestrialization and oligotrophism processes can be seen in the chemical and physical profiles of swamps and marshes. Marsh site 35 experienced a rapid decline in ash and bulk density as development progressed from aquatic to marsh environments, with a corresponding increase in calcium and copper. Swamp site 48 experienced a gradual decline in bulk density, ash, calcium and copper from aquatic to marsh environments. Some surface enrichment is also seen, with the development of swamp vegetation that may be due to bioaccumulation from decomposition and recycling.

According to the discriminant analysis it is possible to identify peat types based upon chemical and physical characteristics. Sediments which are formed under aquatic environments have high ash and bulk density values ($>20\%$, $>0.15 \text{ g/cm}^3$), and high metal contents. Marsh and swamp peats have less than 20% ash, less than 0.15 g/cm^3 bulk density values, but contain more than $10,000 \text{ mg/Kg}$ calcium. Marshes are slightly more mineral rich and eutrophic than swamps. Fens at Elk Island have less than $10,000 \text{ mg/Kg}$ of calcium, and less than 12% ash in the peats, with bogs having under $5,000 \text{ mg/Kg}$ calcium and less than 6% ash.

This thesis describes the development of wetlands at the southern limit of the Boreal Forest, where wetlands are restricted to small topographically defined shallow basins. Here terrestrialization has been occurring since 6700 yBP. North of this site, peatland

development via terrestrialization is of less significance, as large areas of the landscape have become paludified (Sjörs 1961, Nicholson and Vitt 1990). Progression from fen to bog takes place much more quickly in these more northern peatlands as *Sphagnum* invasion occurs much more rapidly (Kuhry *et al.* 1992). Within the park boundaries a wide diversity of wetlands exist, ranging from swamps, saline marshes, fresh water marshes, moderate-rich fens, poor fens, and bogs. All of the wetlands in Elk Island began as small fresh water ponds with shore marsh vegetation along the edges. Edaphic factors not climate controls the diversity of wetlands within the park. Bog development has occurred in certain places and it's development is related to elevation, topography, basin configuration, hydrology, and stream flow. Severe watertable fluctuations prevent the establishment of bryophytes and peatland development progresses from marsh into swamp vegetation.

Beavers have affected the peatlands in EINP extensively by flooding the depressional basins. Fens, bogs, and marshes are inundated with mineral rich waters, and the peatlands regress to former successional stages. Many are completely covered creating ponds or marshes. The amount of impact which beavers have on these sites is dependant upon how high the water table rises. Fens, which have grounded central mats are most likely to be severely affected by beavers. Fens with grounded mats are: a) appealing to beavers (Rebertus 1986), b) more sensitive to flooding because of the grounded mat, c) are diverse within the park (moderate-rich and poor fens), d) and are a relatively rare (aerially) wetland type in the park. Basins containing small open water bodies surrounded by

floating mats are not likely to be impacted by beaver as dam building is limited and the floating mats merely rise with an increase in the water table. Bogs are least likely to be impacted by beavers as low water levels and dense black spruce forests are likely to discourage beavers. Bogs also tend to have few outlet streams, often none, which afford little opportunity for damming. Similarly swamps will be avoided due to the tall shrubs, shallow water tables and few streams. Effects of beavers on marshes (deep or shallow) will be the same as fens, reverting the marsh to previous successional aquatic stages.

Fire is another regular disturbance affecting wetlands within the park. It, however, has had little impact on the development of the wetlands. Only one of the six study sites contained any charcoal layers in the peat profile. All profiles at EINP indicate slow progressive development and do not show any disruptions or retrogressive changes that have occurred due to fire. Management of the wetlands at EINP should involve maintaining the exceptional diversity of wetlands within the park and should be directed towards controlling beaver populations or the negative impact of beavers in the park. A number of fens in the north east section of the park should be set aside and managed intensively to prevent beavers from colonizing and flooding these fens. Efforts should also be directed towards maintaining the few known saline springs and marshes as these are relatively rare in the park. Since fire has not had a significant impact of the development of wetlands in the park, the present practice of prescribed burning should continue.

LITERATURE CITED

- Anderson, T., R. Mathewes, and C. E. Schweger. 1987. Holocene climatic trends in Canada with special reference to the hypsithermal interval. In: Quaternary Geology of Canada. R. Fulton (ed.). Energy, Mines, and Resources, Ottawa
- Andrus, R. E. 1986. Some aspects of *Sphagnum* ecology. Canadian Journal of Botany 64: 416-426.
- Arts, G. H. P. 1990. Aquatic Bryophyta as indicators of water quality in shallow pools and lakes in the Netherlands. Annales Botanicae Fennici 27: 19-32.
- Bayley, S. E., R. S. Behr, and C. A. Kelly. 1986. Retention and release of sulphur from a freshwater wetland. Water, Air, and Soil Pollution 31: 101-114.
- Bayrock, L. A. and G. M. Hughes. 1962. Surficial geology of the Edmonton area. Alberta Research Council, Earth Sciences Preliminary Report 62-6, Edmonton, Alberta.
- Bird, R. C. 1961. Ecology of the aspen parkland of western Canada in relation to land use. Contribution No. 27, Research Station, Canada, Dept. of Agriculture, Winnipeg, Manitoba. Research Branch, Canada Dept. of Agriculture, Ottawa.
- Canada Soil Survey Committee, Subcommittee on Soil Classification. 1978. The Canadian system of soil classification. Canadian Department of Agriculture Publication 1646. Supply and Services Canada, Ottawa, Ontario.
- Campeau, S., H. R. Murkin, and R. D. Titman. The effect of fertilization on bullrush litter decomposition in a nutrient-limited wetland (unpublished).
- Chapman, S. B. 1964. The ecology of Coom Rigg Moss, Northumberland. II. The chemistry of peat profiles and the development of the bog system. Journal of Ecology 52: 315-323.

- Chee, W. and D. H. Vitt. 1989. The vegetation, surface water chemistry and peat chemistry of moderate-rich fens in central Alberta, Canada. *Wetlands* 9: 227-261.
- Clarke, A. H. 1981. The Freshwater Molluscs of Canada. National Museum of Natural Sciences, National Museums of Canada, Ottawa.
- Coker, W. B. and R. N. W. DiLabio. 1979. Initial geochemical results and exploration significance of two uraniferous peat bogs, Kasmere Lake, Manitoba. Geological Survey of Canada Paper 79-1B. Current Research. p. 199-206.
- Cowardin, L. M., V. Carter, F. C. Golet and E. T. LaRoe. 1979. Classification of wetlands and deepwater habitats of the United States. U.S. Department of the Interior, Office of Biological Services, Fish and Wildlife Service, Washington, D.C. FWS/OBS-79/31.
- Clymo, R. S. 1963. Ion exchange in *Sphagnum* and its relation to bog ecology. *Annals of Botany* N.S. 27: 309-324.
- Crown, P. 1977. Soil Survey of Elk Island National Park, A Cultural History. Parks Canada. Department of Indian and Northern Affairs.
- Crum, H. 1988. A focus on Peatlands and Peat Mosses. University of Michigan Press, Ann Arbor.
- Damman, A. W. H. 1978. Distribution and movement of elements in ombrotrophic peat bogs. *Oikos* 30: 480-495.
- Damman, A. W. H. 1986. Hydrology, development, and biogeochemistry of ombrogenous peat bogs with special reference to nutrient relocation in a western Newfoundland bog. *Canadian Journal of Botany* 64: 384-394.
- Damman, A. W. H. 1988. Regulation of nitrogen removal and retention in *Sphagnum* bogs and other peatlands. *Oikos* 51: 291-305.

- D'Elia, C. F., P. A. Steudler and N. Corwin. 1976. Determination of total nitrogen in aqueous samples using persulfate digestion. *Limnology and Oceanography* 22: 760-764.
- Ecoregions Working Group, Canada Committee on Ecological Land Classification. 1989. Climatic regions of Canada. First approximation. Sustainable Development Branch, Canadian Wildlife Service, Conservation and Protection, Environment Canada. Ecological Land Classification Series 23.
- Elomaa, P.-L. 1987. Vertical distribution of copper and manganese in the peat below vegetation types showing mire-center and mire-margin effect. *Proceedings Symposium'87. Wetlands/Peatlands. August 23-27, Edmonton, Alberta.* p. 107-115.
- Emerson, D. 1977. The surficial geology of the cooking lake moraine, east-central Alberta, Canada. M.Sc. thesis, University of Alberta, Dept. of Geology.
- Faegri, K. and J. Iversen. 1989. *Textbook of Pollen Analysis*, 4th edition. Wiley, New York.
- Forbes, J. R. and M. Hickman. 1981. Paleolimnology of two shallow lakes in central Alberta, Canada. *Internationale Revue der gesamten Hydrobiologie* 66: 863-888.
- Foster, D. R. and G. A. King. 1984. Landscape features, vegetation and developmental history of a patterned fen in south-eastern Labrador, Canada. *Journal of Ecology* 72: 115-143.
- Foster, D. R. and S. C. Fritz. 1987. Mire development, pool formation and landscape processes on patterned fens in Dalarna, central Sweden. *Journal of Ecology* 75: 409-437.
- Foster, D. R., H. E. Wright Jr., M. Thelaus and G. A. King. 1988. Bog development and landform dynamics in central Sweden and south-eastern Labrador, Canada. *Journal of Ecology* 76: 1164-1185.

- Futyma, R. P. and N. G. Miller. 1986. Stratigraphy and genesis of the Lake Sixteen peatland, northern Michigan. *Canadian Journal of Botany* 64: 3008-3019.
- Gauthier, R. 1988. Main peatland vegetation units in the Laurentides park. 8th International Peat Congress, Leningrad, USSR Vol. 1: 108-116.
- Gignac, L. D. 1987. Ecological tolerance and niche structure of *Sphagnum* along a pollution gradient near Sudbury, Ontario, Canada. *Canadian Journal of Botany* 65: 1268-1274.
- Gignac, L. D., D. H. Vitt, S. C. Zoltai and S. E. Bayley. 1991(a). Bryophyte response surfaces along climatic, chemical, and physical gradients in peatlands of western Canada. *Nova Hedwigia* 53: 27-71.
- Gignac, L. D., D. H. Vitt and S. E. Bayley. 1991(b). Bryophyte response surfaces along ecological and climatic gradients. *Vegetatio* 93: 29-45.
- Giller, K. E. and B. D. Wheeler. 1988. Acidification and succession in a flood-plain mire in the Norfolk Broadland, U. K. *Journal of Ecology* 76: 849-866.
- Glaser, P. H. 1983. Vegetation patterns in the North Black River peatland, northern Minnesota. *Canadian Journal of Botany* 61: 2085-2104.
- Glaser, P. H., G. A. Wheeler, E. Gorham and H. E. Wright Jr. 1981. The patterned mires of Red Lake peatland, northern Minnesota: vegetation, water chemistry, and landforms. *Journal of Ecology* 69: 575-599.
- Glaser, P. H., J. A. Janssens and D. I. Siegel. 1990. The response of vegetation to chemical and hydrological gradients in the Lost River peatland, northern Minnesota. *Journal of Ecology* 78: 1021-1048.

- Gleason, C. F. and J. A. Coope. 1966. Some observations on the distribution of metals in swamps in eastern Canada. Proceedings, Symposium on Geochemical Prospecting, Ottawa. Geological Survey of Canada, Paper 66-54. Dept. of Energy, Mines and Resources.
- Glime, J. M., R. G. Wetzel and B. J. Kennedy. 1982. The effects of bryophytes on succession from alkaline marsh to *Sphagnum* bog. American Midland Naturalist 108: 209-223.
- Glooschenko, V. and P. Grondin 1988. Wetlands of Eastern Temperate Canada. Chapter 6 In: Wetlands of Canada. Ecological Land Classification. Series No. 24. C. D. A. Rubec (Coordinator). Sustainable development Branch, Environment Canada, Ottawa, Ontario, and Polyscience Publication Inc., Montreal, Quebec. p. 199-248.
- Gorham, E. 1949. Some chemical aspects of a peat profile. Journal of Ecology 37: 24-27.
- Gorham, E. 1956. The ionic composition of some bog and fen waters in the English Lake District. Journal of Ecology 44: 142-152.
- Gorham, E. 1957. The development of peat lands. Quarterly Review of Biology 32: 145-165.
- Gorham, E., S. J. Eisenreich, J. Ford and M. V. Santelmann. 1984. The chemistry of bog waters. Chapter 15. In: Chemical Processes in Lakes. W. Strumm (ed.). Wiley, New York. p. 339-363.
- Gorham, E., J. A. Janssens, G. A. Wheeler, and P. H. Glaser. 1987. The natural and anthropogenic acidification of peatlands. In: Effects of Atmospheric Pollutants on Forests, Wetlands, and Agricultural Ecosystems. T. C. Hutchinson and K. M. Meema (eds.), Springer-Verlag, New York. p. 493-512.
- Greening, H. S. and J. Gerritsen. 1987. Changes in macrophyte community structure following drought in the Okefenokee Swamp, Georgia, USA. Aquatic Botany 28: 113-128.

- Griffin, K. O. 1977. Paleoecological aspects of the Red Lake peatland, northern Minnesota. *Canadian Journal of Botany* 55: 172-192.
- Hale, M. E. 1979. *How to Know the Lichens*. Second edition. Wm. C. Brown Company Publishers, Dubuque, Iowa.
- Heinselman, M. L. 1963. Forest site, bog processes and peatland types in the Glacial Lake Agassiz region. *Ecological Monographs* 33: 327-374.
- Hemond, H. F. 1980. Biogeochemistry of Thoreau's Bog Concord, Massachusetts. *Ecological Monographs* 50: 507-526.
- Hemond, H. F. 1983. The nitrogen budget of Thoreau's Bog. *Ecology* 64: 99-109.
- Hickman, M. 1987. Paleolimnology of a large shallow lake; Cooking Lake, Alberta, Canada. *Archiv fur Hydrobiologie* 111: 121-136.
- Hickman, M. and D. M. Klarer. 1981. Paleolimnology of Lake Isle, Alberta, Canada (Including sediments, chemistry, pigments and diatom stratigraphy). *Archiv fur Hydrobiologie* 91: 490-508.
- Hill, M. O. 1979. TWINSpan- a fortran program for arranging multivariate data in an ordered two-way table by classification of the individuals and attributes. *Ecology and Systematics*, Cornell University, Ithaca, NY.
- Horton, D. G., D. H. Vitt and N. G. Slack. 1979. Habitats of circumboreal-subarctic *Sphagna*: I. A quantitative analysis and review of species in the Caribou Mtn., northern Alberta. *Canadian Journal of Botany* 57: 2283-2317.
- Husband, B. C. and M. Hickman. 1985. Growth and biomass allocation of *Ruppia occidentalis* in three lakes, differing in salinity. *Canadian Journal of Botany* 63: 2004-2014.
- Ireland, R. R. 1982. Moss Flora of the Maritime Provinces. National Museums of Canada, Ottawa. Publications in Botany No. 13.

- Jackson, S. T. and D. F. Charles. 1988. Aquatic macrophytes in Adirondack (New York) lakes: Patterns of species composition in relation to environment. *Canadian Journal of Botany* 66: 1449-1460.
- Janssen, C. R. 1988. Development of peatland vegetation types and patterns in the Myrtle Lake peatland: An example of a continental boreal peatland near the present-day prairie border in the upper midwest of the U.S.A. *Proceedings of the 8th International Peat Congress, Leningrad, USSR*. Vol. 1: 156-162.
- Janssens, J. A. 1983. Past and extant distribution of *Drepanocladus* in North America, with notes on the differentiation of fossil fragments. *Journal of the Hattori Botanical Laboratory* 54: 251-298.
- Janssens, J. A. 1988. Fossil bryophytes and paleoenvironmental reconstruction of peatlands. In: *Methods in Bryology. Proceedings of the Bryological Methods Workshop, Mainz Germany*. Aug. 17-23. 1987. J. M. Glime (ed.). The Hattori Botanical Laboratory, Nichinan, Japan. p 299-306.
- Janssens, J. A. and P. H. Glaser. 1986. The bryophyte flora and major peat-forming mosses at Red Lake peatland, Minnesota. *Canadian Journal of Botany* 64: 427-442.
- Janssens, J. A., B. C. S. Hansen, P. H. Glaser and C. W. Barnosky. 1989. Development of a raised-bog complex in northern Minnesota. In: *Patterned Peatlands of Northern Minnesota*. H. E. Wright Jr., B. Coffin, and N. Aasing (eds.), (unpublished).
- Jeglum, J. K. 1971. Plant indicators of pH and water level in peatlands at Candle Lake, Saskatchewan. *Canadian Journal of Botany* 49: 1661-1676.
- Jeglum, J. K. 1972. Boreal forest wetlands near Candle Lake, Central Saskatchewan. *The Musk-Ox* 11: 41-58.
- Kangas, P. C. 1990. Long-term development of forested wetlands. Chapter 3. In: *Ecosystems of the World: Forested Wetlands*. D. W. Goodall (ed). Elsevier, New York, p. 25-51.

- Karlin, E. F. and L. C. Bliss. 1984. Variation in substrate chemistry along microtopographical and water-chemistry gradients in peatlands. *Canadian Journal of Botany* 62: 142-153.
- Keddy, P. A. and A. A. Reznicek. 1986. Great lakes vegetation dynamics: The role of fluctuating water levels and buried seeds. *Journal of Great Lakes Research* 12: 25-36.
- Klinger, F. L., S. A. Elias, V. M. Behan-Pelletier and N. E. Williams. 1990. The bog climax-hypothesis: Fossil arthropod and stratigraphic evidence in peat sections from southeast Alaska. *Holarctic Ecology* 13: 72-80.
- Kratz, T. K. and C. B. DeWitt. 1986. Internal factors controlling peatland-lake ecosystem development. *Ecology* 67: 100-107.
- Kubiw, H. 1987. The development and chemistry of Muskiki and Marguerite Lake Peatlands. M.Sc. thesis, University of Alberta, Dept. of Botany.
- Kubiw, H., M. Hickman and D. H. Vitt. 1989. The developmental history of peatlands at Muskiki and Marguerite lakes, Alberta. *Canadian Journal of Botany* 67: 3534-3544.
- Kuhry, P. 1988. A paleobotanical and palynological study of Holocene peat from the El Bosque mire, located in a volcanic area of the Cordillera Central of Columbia. *Review of Palaeobotany and Palynology* 55: 19-72.
- Kuhry, P., B. J. Nicholson, L. D. Gignac, D. H. Vitt and S. E. Bayley. 1992. Development of *Sphagnum* dominated peatlands in Boreal continental Canada. *Canadian Journal of Botany* (in press).
- Lang, A. H. 1974. Guide to the Geology of Elk Island National Park, The Origin of its Hills and other Scenery. The Geological Survey of Canada, Dept. of Energy, Mines and Resources, Ottawa. Miscellaneous Report 22.

- Largin, I. F., S. E. Priemskaya, A. N. Sventikhovskaya, and S. N. Tyuremnov. 1972. Microelement content and distribution in peat deposits. Proceedings of the 4th International Peat Congress, Otaniemi, Finland, June 25-30. p. 77-85.
- Lembrechts, J. F. M. and O. L. J. Vanderborght. 1985. Mineral content of *Sphagnum* mosses in Belgian bog ecosystems. Journal of Environmental Quality 14: 217-223.
- Lewis, F. J., E. S. Dowding and E. H. Moss. 1928. The vegetation of Alberta. II. The swamp, moor, and bog forest vegetation of central Alberta. Journal of Ecology 16: 19-70.
- Lichti-Federovich, S. 1970. The pollen stratigraphy of a dated section of Late Pleistocene lake sediment from central Alberta. Canadian Journal of Earth Sciences 7: 938-945.
- Mallik, A. U. and R. W. Wein. 1986. Response of a *Typha* marsh community to draining, flooding, and seasonal burning. Canadian Journal of Botany 64: 2136-2143.
- Malmer, N. 1962. Studies on mire vegetation in the Archaean area of southwestern Gotaland (South Sweden). 1. Vegetation and habitat conditions on the Akhult mire. Opera Botanica 7: 1 - 322.
- Malmer, N. 1985. Remarks to the classification of mires and mire vegetation. Scandinavian arguments. Aquilo Ser. Bot. 21: 9-17.
- Malmer, N. 1986. Vegetational gradients in relation to environmental conditions in northwestern European mires. Canadian Journal of Botany 64: 375-383.
- Malmer, N. and H. Sjörs. 1955. Some determinations of elementary constituents in mire plants and peat. Botaniska Notiser 108: 46-80.
- Millar, J. B. 1976. Wetland classification in western Canada. A guide to marshes and shallow open water wetlands in the grasslands and the parklands of the prairie provinces. Canadian Wildlife Service. Report Series Number 37.

- Miller, N. G. and R. P. Futyma. 1987. Paleohydrological implications of Holocene peatland development in northern Michigan. *Quaternary Research* 27: 297-311.
- Mitsch, W. J. and J. G. Gosselink. 1986. *Wetlands*. Van Nostrand Reinhold, New York.
- Mörnsjö, T. 1968. Stratigraphical and chemical studies on two peatlands in Scania, south Sweden. *Botaniska Notiser* 121: 343-360.
- Moss, E. H. 1953. Marsh and bog vegetation in northwestern Alberta. *Canadian Journal of Botany* 31: 448-470.
- Moss, E. H. 1983. *Flora of Alberta, A manual of flowering plants, conifers, ferns and fern allies found growing without cultivation in the province of Alberta, Canada*. Revised by J. G. Packer. University of Alberta, Canada. University of Toronto Press, Toronto.
- National Wetlands Working Group. 1988. *Wetlands of Canada. Ecological Land Classification Series No. 24*. C. D. A. Rubec (Coordinator). Sustainable Development Branch, Environment Canada, Ottawa, Ont., and Polyscience Publications Inc., Montreal, Quebec, 425 p.
- Nichols, H. 1969. The late Quaternary history of vegetation and climate at Porcupine Mountain and Clearwater Bog, Manitoba. *Arctic and Alpine Research* 1: 155-167.
- Nicholson, B. J. 1987. Peatland paleoecology and peat chemistry at Mariana Lakes, Alberta. M.Sc. thesis, University of Alberta, Dept. of Botany.
- Nicholson, B. J. 1989. Peat chemistry of a continental mire complex in western Canada. *Canadian Journal of Botany* 67: 763-775.
- Nicholson, B. J. and D. H. Vitt. 1990. The paleoecology of a peatland complex in continental western Canada. *Canadian Journal of Botany* 68: 121-138.

- Osvald, H. 1970. Vegetation and stratigraphy of peatlands in North America. *Nova Acta Regiae Societatis Scientiarum Upsaliensis* Ser. V: C, Vol. 1.
- Ovenden, L. 1985. Hydroseral histories of the Old Crow peatland, northern Yukon. Ph.D. thesis, University of Toronto, Department of Botany.
- Ovenden, L. and G. R. Brassard. 1989. Wetland vegetation near Old Crow, northern Yukon. *Canadian Journal of Botany* 67: 954-960.
- Pakarinen, P. and E. Gorham. 1983. Mineral element composition of *Sphagnum fuscum* peats collected from Minnesota, Manitoba and Ontario. Proceedings of the International Symposium on Peat Utilization. C. H. Fuchsman and S. A. Spigarelli (eds.). Bemidji State University, Bemidji, Minnesota; October 10-13, 1983. p. 417-429.
- Pakarinen, P. and K. Tolonen. 1977. Nutrient contents of *Sphagnum* mosses in relation to bog water chemistry in northern Finland. *Lindbergia* 4: 27-33.
- Parks Canada. 1986. Elk Island National Park: Resource description and analysis. Vol 1. Resource Conservation, Elk Island National Park.
- Rebertus, A. J. 1986. Bogs as beaver habitat in north-central Minnesota. *American Midland Naturalist* 116: 240-245.
- Renato, G. 1990. Vegetation patterns and nutrient status of two mixed mires in the southern Alps. *Journal of Vegetation Science* 1: 663-668.
- Richardson, C. J., D. L. Tilton, J. A. Kadlec, J. P. M. Chamie and W. A. Wentz. 1978. Nutrient dynamics of a northern wetland ecosystem. In: *Freshwater Wetlands*. R. E. Good, D. F. Whigham, R. L. Simpson (eds.). Academic Press, San Francisco. p. 217-241.

- Riley, J. L. and L. Michaud. 1989. Peat and peatland resources of northwestern Ontario. Ontario Geological Survey Miscellaneous Paper 144. Ontario Ministry of Northern Development and Mines.
- Rochèfort, L. 1987. Biological effects of wet acid deposition on peatland bryophytes. M.Sc. thesis, University of Alberta, Dept. of Botany.
- Rowe, J. S. 1972. Forest Regions of Canada. Department of the Environment, Canadian Forestry Service, Publication No. 1300. Information Canada, Ottawa.
- Scace and Associates Ltd. 1976. Elk Island National Park, A Cultural History. Parks Canada, Department of Indian and Northern Affairs.
- Schuster, R. M. 1953. Boreal Hepaticae. A manual of the liverworts of Minnesota and adjacent regions. American Midland Naturalist 49: 257-684.
- Schwintzer, C. R. 1981. Vegetation and nutrient status of northern Michigan bogs and conifer swamps with a comparison to fens. Canadian Journal of Botany 59: 842-853.
- Shay, J. M. and C. T. Shay. 1986. Prairie marshes in western Canada, with specific reference to the ecology of five emergent macrophytes. Canadian Journal of Botany 64: 443-454.
- Shotyk, W. 1986. An overview of the inorganic chemistry of peats. Proceedings: Advances in Peatland Engineering. National Research Council Canada, Carleton University, 25-36, August. p. 249-258.
- Shotyk, W. 1988. Review of the inorganic geochemistry of peats and peatland waters. Earth-Science Reviews 25: 95-176.
- Siegel, D. I. 1983. Ground water and the evolution of patterned mires, Glacial Lake Agassiz Peatlands, northern Minnesota. Journal of Ecology 71: 913-921.

- Siegel D. I. 1988. The recharge-discharge function of wetlands near Juneau, Alaska. Part II. Geochemical Investigations. *Groundwater* 26: 580-586.
- Siegel, D. I. and P. H. Glaser. 1987. Groundwater flow in a bog-fen complex, Lost River Peatland, northern Minnesota. *Journal of Ecology* 75: 743-754.
- Sims, R. A., D. W. Cowell and G. M. Wickware. 1982. Classification of fens near southern James Bay, Ontario, using vegetational physiognomy. *Canadian Journal of Botany* 60: 2608-2623.
- Sjörs, H. 1950. Regional studies in north Swedish mire vegetation. *Botaniska Notiser* 2: 173-222.
- Sjörs, H. 1952. On the relationship between vegetation and electrolytes in North Swedish mire waters. *Oikos* 2 (1950): 241-258.
- Sjörs, H. 1961. Forest and peatland at Hawley Lake, northern Ontario. *Bulletin of the Natural Museum of Canada* 171:1-31.
- Slack, N. G., D. H. Vitt and D. G. Horton. 1980. Vegetation gradients of minerotrophically rich fens in western Alberta. *Canadian Journal of Botany* 58: 330-350.
- Sonesson, M. 1970. Studies on the mire vegetation in the Torneträsk area, northern Sweden. IV. Some habitat conditions of the poor mires. *Botaniska Notiser* 123: 67-111.
- Starr, M. R. and J. Laine. 1988. Interactions among peat properties, nutrient contents and water chemistry at Lakkasuo Mire, central Finland. *Proceedings of the 8th International Peat Congress, Leningrad, USSR, Vol 1: 140-147.*
- Stewart, R. E. and H. A. Kantrud. 1972. Vegetation of prairie potholes, North Dakota, in relation to quality of water and other environmental factors. *Hydrology of prairie potholes in North Dakota. Geological Survey Professional paper 585-D. U.S. Dept. of the Interior.*

- Strong, W. L. and K. R. Leggat. 1981. Ecoregions of Alberta. Alberta Energy and Natural Resources, Resource Evaluation and Planning Division.
- Tallis, J. H.. 1983. Changes in wetland communities. In: Ecosystems of the World. Mires: Swamp, Bog, Fen, and Moor. Vol. 4A. A. J. P. Gore (ed.). Elsevier Scientific Publishing Co., New York. p. 311-349.
- Taylor, T. M. C. 1983. The Sedge Family. British Columbia Provincial Museum Handbook #43, Victoria, B. C.
- Thorndike, R. M. 1978. Correlational Procedures for Research. Gardner Press, Inc., New York.
- Urban, N. R. and S. E. Bayley. 1988. The relative importance of denitrification and nitrate assimilation in midcontinental bogs. *Limnology and Oceanography* 33: 1611-1617.
- Urban, N. R., S. J. Eisenreich and E. Gorham. 1987. Proton cycling in bogs: Geographic variation in northeastern North America. In: Effects of Atmospheric Pollutants on Forests, Wetlands and Agricultural Ecosystems. T. C. Hutchinson and K. Meema (eds.) Springer-Verlag Berlin Heidelberg, NATO AST Series, Vol. G16.
- Urban, N. R. and S. J. Eisenreich. 1988. Nitrogen cycling in a forested Minnesota bog. *Canadian Journal of Botany* 66: 435-449.
- Vance, R. E. 1979. Late Holocene paleoecology of the aspen parkland region of western Canada. M. A. thesis, University of Alberta, Dept. of Anthropology.
- Vance, R. E., D. Emerson, and T. Habgood. 1983. A mid-Holocene record of vegetative change in central Alberta. *Canadian Journal of Earth Sciences* 20: 364-376.
- Verhoeven, J. T. A., E. Maltby and M. B. Schmitz. 1990. Nitrogen and phosphorus mineralization in fens and bogs. *Journal of Ecology* 78: 713-726.

- Vitt, D. H. and R. E. Andrus. 1977. The genus *Sphagnum* in Alberta. *Canadian Journal of Botany* 55: 331-357.
- Vitt, D. H. and S. E. Bayley. 1984. The vegetation and water chemistry of four basin mires in northwestern Ontario. *Canadian Journal of Botany* 62: 1485-1500.
- Vitt, D. H. and W. L. Chee. 1990. Relationships of vegetation to surface water chemistry and peat chemistry in fens of Alberta, Canada. *Vegetatio* 89: 87-106.
- Vitt, D. H., and P. Kuhry. 1992. Changes in moss-dominated wetland ecosystems. In: *Bryophytes and Lichens in a Changing Environment*. J. W. Bates and A. M. Farmer (eds.). Oxford University Press, Oxford. p. 178-210.
- Vitt, D. H. and N. G. Slack. 1975. An analysis of vegetation of *Sphagnum*-dominated kettle-hole bogs in relation to environmental gradients. *Canadian Journal of Botany* 53: 332-359.
- Vitt, D. H. and N. G. Slack. 1984. Niche diversification of *Sphagnum* relative to environmental factors in northern Minnesota peatlands. *Canadian Journal of Botany* 62: 1409-1430.
- Vitt, D. H., P. Achuff, and R. E. Andrus. 1975. The vegetation and chemical properties of the patterned fens in the Swan Hills, north-central Alberta. *Canadian Journal of Botany* 53: 2776-2795.
- Vitt, D. H., D. G. Horton, N. G. Slack and N. Malmer. 1990. *Sphagnum*-dominated peatlands of the hyperoceanic British Columbia coast: Patterns in surface water chemistry and vegetation. *Canadian Journal of Forest Research* 20: 696-711.
- Vitt, D. H., S. E. Bayley and T.-L. Jin. 1992. Seasonal variation in water chemistry over a bog-rich fen gradient in continental western Canada. *Canadian Journal of Fisheries and Aquatic Sciences* (submitted).

- Walker, B. H. and R. T. Coupland. 1970. Herbaceous wetland vegetation in the aspen grove and grassland regions of Saskatchewan. *Canadian Journal of Botany* 48: 1861-1878.
- Walker, D. 1970. Direction and rate in some British post-glacial hydrosere. In: *Studies in the vegetational history of the British Isles*. P. Walker and R. G. West (ed.). Cambridge at the University Press.
- Warén, H. 1924. Untersuchungen über die botanische entwicklung der moore mit berücksichtigung der chemischen zusammensetzung des torfes. *Finska Mosskulturforeningen Vetenskapliga Skrifter* 5.
- Warner, B. G. 1989. Geological and paleoecological aspects of some *Sphagnum* bogs in Ontario. In: *Wetlands: Inertia or momentum?* M. J. Bardecki and N. Patterson (ed.). Federation of Ontario Naturalists, Don Mills, Ontario. p. 329-338.
- Warner, B. G., H. J. Kubiw, and P. F. Karrow. 1991(a). Origin of a postglacial kettle-fill sequence near Georgetown, Ontario. *Canadian Journal of Earth Sciences* 28: 1965-1973.
- Warner, B. G., K. Tolonen, and M. Tolonen. 1991(b). A postglacial history of vegetation and bog formation at Point Escuminac, New Brunswick. *Canadian Journal of Earth Sciences* 28: 1572-1582.
- Wassen, M. J., A. Barendregt, M. C. Bootsma, and P. P. Schot. 1989. Groundwater chemistry and vegetation gradients from rich fen to poor fen in the Naardermeer (the Netherlands). *Vegetatio* 79: 117-132.
- Weider, R. K. 1985. Peat and water chemistry at Big Run Bog, a peatland in the Appalachian mountains of West Virginia, USA. *Biogeochemistry* 1: 277-302.
- Welling, C. H., R. L. Pederson and A. G. van der Valk. 1988a. Temporal patterns in recruitment from the seed bank during drawdowns in a prairie wetland. *Journal of Applied Ecology* 25: 999-1007.

- Welling, C. H., R. L. Pederson and A. G. van der Valk. 1988b. Recruitment from the seed bank and the development of zonation of emergent vegetation during drawdown in a prairie wetland. *Journal of Ecology* 76: 483-496.
- Wetzel, R. G. 1983. *Limnology*. 2nd Edition. Saunders College Publishing, Toronto.
- Wilcox, D. A. and J. E. Meeker. 1991. Disturbance effects on aquatic vegetation in regulated and unregulated lakes in northern Minnesota. *Canadian Journal of Botany* 69: 1542-1551.
- Wilcox, D. A. and H. A. Simonin. 1988. The stratigraphy and development of a floating peatland, Pinhook Bog, Indiana. *Wetlands* 8: 75-90.
- Winkler, M. G. 1988. Effect of climate on development of two *Sphagnum* bogs in south-central Wisconsin. *Ecology* 69: 1032-1043.
- Yefimov, V. N. and Z. S. Yefimova. 1973. Chemical composition of bog water in the northwestern part of European USSR. *Soviet Soil Science* 641-650. Translated from *Pochvovedeniye* 1973: 27-36.
- Zoltai, S. C. 1975. Southern limit of coniferous trees on the Canadian Prairies. Environment Canada Forestry Service, Northern Forest Research Center, Information Report NOR-X-128, Edmonton, Alberta.
- Zoltai, S. C. 1988. Wetland Environments and Classification. Chapter 1 In: *Wetlands of Canada. Ecological Land Classification*. Series No. 24. C. D. A. Rubec (Coordinator). Sustainable Development Branch, Environment Canada, Ottawa, Ontario, and Polyscience Publication Inc., Montreal, Quebec. p. 1-23.
- Zoltai, S. C. 1991. Estimating the age of peat samples from their weight: A study from west-central Canada. *The Holocene* 1: 68-73.

- Zoltai, S. C. and J. D. Johnson. 1985. Development of a treed bog island in a minerotrophic fen. *Canadian Journal of Botany* 63: 1076-1085.
- Zoltai, S. C. and J. D. Johnson. 1987. Relationships between nutrients and vegetation in peatlands of the prairie provinces. *Proceedings of the Symposium'87 Wetlands/Peatlands*, Edmonton, August 23-28. p. 535-543.
- Zoltai, S. C. and D. H. Vitt. 1990. Holocene climatic change and the distribution of peatlands in western interior Canada. *Quaternary Research* 33: 231-240.
- Zoltai, S. C., F. C. Pollet, J. K. Jeglum and G. O. Adams. 1975. Developing a wetland classifications for Canada. In: *Proceedings of the fourth North American Forest Soils Conference*. B. Bernier and C. H. Winget (ed.). Laval University, Quebec.

Appendix I

1	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20	21	22	23	24	25	26	27	28	29	30	31	32	33	34	35	36	37	38	39	40	41	42	43	44	45	46	47	48	49	50	51	52	53	54	55	56	57	58	59	60	61	62	63	64	65	66	67	68	69	70	71	72	73	74	75	76	77	78	79	80	81	82	83	84	85	86	87	88	89	90	91	92	93	94	95	96	97	98	99	100	101	102	103	104	105	106	107	108	109	110	111	112	113	114	115	116	117	118	119	120	121	122	123	124	125	126	127	128	129	130	131	132	133	134	135	136	137	138	139	140	141	142	143	144	145	146	147	148	149	150	151	152	153	154	155	156	157	158	159	160	161	162	163	164	165	166	167	168	169	170	171	172	173	174	175	176	177	178	179	180	181	182	183	184	185	186	187	188	189	190	191	192	193	194	195	196	197	198	199	200	201	202	203	204	205	206	207	208	209	210	211	212	213	214	215	216	217	218	219	220	221	222	223	224	225	226	227	228	229	230	231	232	233	234	235	236	237	238	239	240	241	242	243	244	245	246	247	248	249	250	251	252	253	254	255	256	257	258	259	260	261	262	263	264	265	266	267	268	269	270	271	272	273	274	275	276	277	278	279	280	281	282	283	284	285	286	287	288	289	290	291	292	293	294	295	296	297	298	299	300	301	302	303	304	305	306	307	308	309	310	311	312	313	314	315	316	317	318	319	320	321	322	323	324	325	326	327	328	329	330	331	332	333	334	335	336	337	338	339	340	341	342	343	344	345	346	347	348	349	350	351	352	353	354	355	356	357	358	359	360	361	362	363	364	365	366	367	368	369	370	371	372	373	374	375	376	377	378	379	380	381	382	383	384	385	386	387	388	389	390	391	392	393	394	395	396	397	398	399	400	401	402	403	404	405	406	407	408	409	410	411	412	413	414	415	416	417	418	419	420	421	422	423	424	425	426	427	428	429	430	431	432	433	434	435	436	437	438	439	440	441	442	443	444	445	446	447	448	449	450	451	452	453	454	455	456	457	458	459	460	461	462	463	464	465	466	467	468	469	470	471	472	473	474	475	476	477	478	479	480	481	482	483	484	485	486	487	488	489	490	491	492	493	494	495	496	497	498	499	500	501	502	503	504	505	506	507	508	509	510	511	512	513	514	515	516	517	518	519	520	521	522	523	524	525	526	527	528	529	530	531	532	533	534	535	536	537	538	539	540	541	542	543	544	545	546	547	548	549	550	551	552	553	554	555	556	557	558	559	560	561	562	563	564	565	566	567	568	569	570	571	572	573	574	575	576	577	578	579	580	581	582	583	584	585	586	587	588	589	590	591	592	593	594	595	596	597	598	599	600	601	602	603	604	605	606	607	608	609	610	611	612	613	614	615	616	617	618	619	620	621	622	623	624	625	626	627	628	629	630	631	632	633	634	635	636	637	638	639	640	641	642	643	644	645	646	647	648	649	650	651	652	653	654	655	656	657	658	659	660	661	662	663	664	665	666	667	668	669	670	671	672	673	674	675	676	677	678	679	680	681	682	683	684	685	686	687	688	689	690	691	692	693	694	695	696	697	698	699	700	701	702	703	704	705	706	707	708	709	710	711	712	713	714	715	716	717	718	719	720	721	722	723	724	725	726	727	728	729	730	731	732	733	734	735	736	737	738	739	740	741	742	743	744	745	746	747	748	749	750	751	752	753	754	755	756	757	758	759	760	761	762	763	764	765	766	767	768	769	770	771	772	773	774	775	776	777	778	779	780	781	782	783	784	785	786	787	788	789	790	791	792	793	794	795	796	797	798	799	800	801	802	803	804	805	806	807	808	809	810	811	812	813	814	815	816	817	818	819	820	821	822	823	824	825	826	827	828	829	830	831	832	833	834	835	836	837	838	839	840	841	842	843	844	845	846	847	848	849	850	851	852	853	854	855	856	857	858	859	860	861	862	863	864	865	866	867	868	869	870	871	872	873	874	875	876	877	878	879	880	881	882	883	884	885	886	887	888	889	890	891	892	893	894	895	896	897	898	899	900	901	902	903	904	905	906	907	908	909	910	911	912	913	914	915	916	917	918	919	920	921	922	923	924	925	926	927	928	929	930	931	932	933	934	935	936	937	938	939	940	941	942	943	944	945	946	947	948	949	950	951	952	953	954	955	956	957	958	959	960	961	962	963	964	965	966	967	968	969	970	971	972	973	974	975	976	977	978	979	980	981	982	983	984	985	986	987	988	989	990	991	992	993	994	995	996	997	998	999	1000
Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var	Drop/Var																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																																				

Appendix I Site 2-2 Macrofossil Data Continued

	W	X	Y	Z	AA	AB	AC	AD	AE	AF	AG	AH	AI	AJ	AK	AL	AM	AN	AO	AP	AQ	AR
	Collar 24	Artisan 24	Drill 24	Post	Wood	Upright	Charcoal 24	Charcoal 24	Char	Zircon 24	Zircon 24	Roma 24	TPHA	Patron 24	Ceraphy	Algal 24	Revered 24	Nat 24	Salt 24	Salt 24	Salt 24	Moyn 24
1																						
2																						
3																						
4	3.6	3.6				0.5																
5																						
6																						
7																						
8																						
9																						
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23																						
24																						
25																						
26																						
27	0.5																					
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Appendix I Site 2-2 Macrofossil Data Continued

	AB	AT	AU	AV	AW	AX	AY	AZ	BA	BB	BC	BD	BE	BF	BG	BH
	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed	Grass seed
1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
2	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
3	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
4	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
5	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
6	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
7	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
8	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
9	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
10	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
11	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
12	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
13	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
14	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
15	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
16	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
17	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
18	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
19	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
20	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
21	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
22	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
23	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
24	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
25	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
26	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
27	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
28	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
29	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
30	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
31	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
32	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
33	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
34	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
35	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
36	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
37	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
38	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
39	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
40	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
41	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
42	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
43	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
44	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1
45	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1	1

Appendix I Site 2-3 Macrofossil Data

1	Depth	A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S	T	U	V
2	6-5	10	0.1	25	3	63																	
3	18-Oct																						
4	20-25	0.1	10	13.5	63.8	2																	
5	20-25			17.5	12.8	65.4	1.5							0.5									
6	40-45			25	10	64.2	2												2.5	1.5	1		
7	45-50			12.5	10	55																	17.5
8	55-60			25	5	31																	
9	65-70			22.5	5	44	8																
10	75-80			15	10	25.5	4																
11	80-85			7.5	8	65	1.5																
12	85-100			25	4	49.4																	
13	110-115	0.5	8.5	15	64	1																	
14	120-125	1	17.5	3.5	64.3																		
15	130-135	1	32.3	6	41	3.5																	
16	145-150	0.5	35	1	25																		
17	160-165	0.1	63.9		7.5																		
18	180-185	0.5	17.5	5	8																		
19	170-175	0.1	40		7.5																		
20	180-185	0.5	18	1.5	11.5																		
21	190-195	8	54.5	3	4.5																		
22	200-205	8	87.8																				
23	210-215	2	87.3	2.5	10																		
24	220-225	1	80		7.5																		
25	230-235	1	72.9	1	15																		
26	240-245	3	61.8	1.7																			
27	250-255	1	64	22.5																			
28	260-265	1	89.8																				
29	270-275	1	83.4		7.5																		
30	280-285	1	70.5		15																		
31	290-295	0.1	73.5		15	0.1																	
32	300-305	1	73.5		15	0.1																	
33	310-315	0.5	88.1		1																		
34	320-325	10	75.4		2																		
35	330-335	5	80		5																		
36	340-345	2	73.1		15																		
37	350-355	7.5	68.5		15																		
38	360-365	1	86.4		36.5																		
39	370-375	17.5	43.7		22.5																		
40	380-385	2	72.5		22.5																		
41	390-395	20	64.2		10																		
42	400-405	30	57.8		5																		
43	410-415	34.3	5		35																		
44	420-425	10	62		5																		
45	430-435	5	89.4		3																		
46	440-445	20	63.8		5																		
47	450-455	20	74.7		5																		
48	460-465	25	71.9																				
49	470-475	10	84.1		10																		
50	480-485	5	80		7.5																		

Appendix I

[illegible]

Appendix I

AB	AT	AU	AV	AW	AZ	AY	AZ	BA	BB	BC	BD	BE	BF	BG	BH	BI
Blum scale	Wing length	Birds per	Birds nest	Carni nest	Carni nest	Carni per	Rat gray	Carni nest	Other Sea	Overcasts	Carni in	Chadron	Cadality	Chadron	Dred mal	Shallow
1	1	1	6	1	1					0.1					1.80	3.5
2	2	4	6	3				11							2.48	3.6
3	4	1	3.8	6	1					0.1				1	2.24	3.2
4	4	1	3	4				2							2.07	1.7
5	8	1	3	11										1	2.23	4.8
6	6	1	3	12				1							2.1	4
7	1	2	8	7		2								3	0.5	2.55
8	6	1	10	23			1	6						2	2.17	6
9	6	1	10	23					1						2.11	5.2
10	19	1	23	18	3									1	2.18	5.2
11	23	2	21	22										2	2.25	4.2
12	3	8	1	8	14	1		2						3	2.86	4.4
13	1	4	1	8	11				3					1	3.12	4
14	1	18	6	17	7			2						11	2.99	4.2
15	18	6	1	22	7			1	6					6	3.24	3.8
16	3	2	10	3										8	3.11	6
17	7	1	18	6					2					12	2.93	4.7
18	6	1	8	7			1							1	2.88	4.5
19	2	19	2	9		3		2		0.1				22		
20	1	1	6	6				1						12		
21	1	2	8	7						0.5				6		
22	2	3	8	6					1	5				11		
23	3	5	5							0.1				22		
24	1	8	14	3		2			2	0.1				20		
25	1	1	8	6						0.5				4.3		
26	1	10	3						6	0.1				10		
27	2	6	5					3						79		
28	2	8	6						1	0.3				2		
29	2	9	5					2	1					19		
30	2	7								1				35		
31	1	4	6	1		1		1		0.1				18		
32	1	1	4											1		
33	1	1	5			1				0.1				68		
34	3	3				2			2	0.5				40		
35	3	3							2	0.5				1		
36	3	3						1	1	0.1				2		
37	3	3												15		
38	3	3								1				33		
39	2	12				3		3						67		
40	1	2	2					6	1	1				62		
41	2	2												83		
42	2	7				1		4						6		
43	2	7							2	0.5				68		
44	15								2	0.5				48		
45	6					5		1	2	0.5				147		
46	1	4								0.5				71		
47	1	1	5			1								0.1		
48	1	1	4			2								0.1		
49	1	1	5											40		
50	1	1	5											0.1		
51	1	1	5											135		
52	16					3		1		0.1				0.1		
53	14													35		
54	14									0.5				114		
55	4					3								1		
56	4									0.1				53		
57	4					6			2	0.1				0.1		
58	2					2		2						151		
59	2								1	2				165		
60	4									0.5				88		
61	4					1		2						0.1		
62	4													6.19		
63	4													6.77		
64	4													3.5		

Appendix I Site 3-3 Macrofossil Data

	A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S	T	U	V
	Depth	Mineral	Orbitulus	Herb	Herb root	Myrica	Ceratoph	Polans	Drep. Sp.	Drep. Crest	Ad. Spring	Spring	Spring	Spring	Spring	W. Myrica	Pice	Wood	Ugracae	Decid. hrt	Late need	Others
1	18-20		48	2	10											19	7.5	0.5	10.5	7.7	2.5	0.5
2	30-35		55	2.5	20											9	2.7	0.1	7.7	2.5	2	0.5
3	40-45		20	3.5	3.5											3.5	10		47.5	8		
4	50-55		48	17.5	7.5														8.5			
5	60-65		53.5	7.5	7.5														22.5			
6	70-75		38.5	1.5	1.5														45			
7	80-85		88.5	5	18														7.5			
8	90-95		51.5	11	25.5														1.5	7.5		
9	100-105		66.5	2	23														2			
10	110-115	3.5	62		22.5	1		1	1	2	1								6			
11	120-125	1	64		19.5	1		2.5	1	1									9			3
12	130-135	7	71.5		8.5	1	2	1.5	1										6.5			
13	140-145	1	77.5		6			1	1										12.5			
14	150-155	2	67.5		7.5			1	1										20			1
15	160-165	2	71					1	1										10			2
16	170-175	5	73		10			1	1										2.5			1
17	175-180	50	18		3														8			

Appendix I Site 3-3 Macrofossil Data Continued

W	X	Y	Z	AA	AB	AC	AD	AE	AF	AG	AH	AI	AJ	AK	AL	AM	AN	AO	AP	AQ	AR
Cherted	Strata	Type	Myoga	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata	Strata
1		1																			
2																					
3																					
4																					
5																					
6																					
7																					
8																					
9																					
10																					
11																					
12																					
13																					
14																					
15																					
16																					
17																					
18																					
19																					

Appendix I
Site 3-3 Macrofossil Data Continued

	AB	AT	AU	AV	AW	AX	AY	AZ	BA	BB	BC	BD
	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard	Martha's Vineyard
1	1	1	1	1	1	1	1	1	1	1	1	1
2	1	1	1	1	1	1	1	1	1	1	1	1
3	1	1	1	1	1	1	1	1	1	1	1	1
4	1	1	1	1	1	1	1	1	1	1	1	1
5	1	1	1	1	1	1	1	1	1	1	1	1
6	1	1	1	1	1	1	1	1	1	1	1	1
7	1	1	1	1	1	1	1	1	1	1	1	1
8	1	1	1	1	1	1	1	1	1	1	1	1
9	1	1	1	1	1	1	1	1	1	1	1	1
10	1	1	1	1	1	1	1	1	1	1	1	1
11	1	1	1	1	1	1	1	1	1	1	1	1
12	1	1	1	1	1	1	1	1	1	1	1	1
13	1	1	1	1	1	1	1	1	1	1	1	1
14	1	1	1	1	1	1	1	1	1	1	1	1
15	1	1	1	1	1	1	1	1	1	1	1	1
16	1	1	1	1	1	1	1	1	1	1	1	1
17	1	1	1	1	1	1	1	1	1	1	1	1
18	1	1	1	1	1	1	1	1	1	1	1	1
19	1	1	1	1	1	1	1	1	1	1	1	1
20	1	1	1	1	1	1	1	1	1	1	1	1
21	1	1	1	1	1	1	1	1	1	1	1	1
22	1	1	1	1	1	1	1	1	1	1	1	1
23	1	1	1	1	1	1	1	1	1	1	1	1
24	1	1	1	1	1	1	1	1	1	1	1	1
25	1	1	1	1	1	1	1	1	1	1	1	1
26	1	1	1	1	1	1	1	1	1	1	1	1
27	1	1	1	1	1	1	1	1	1	1	1	1
28	1	1	1	1	1	1	1	1	1	1	1	1
29	1	1	1	1	1	1	1	1	1	1	1	1
30	1	1	1	1	1	1	1	1	1	1	1	1
31	1	1	1	1	1	1	1	1	1	1	1	1
32	1	1	1	1	1	1	1	1	1	1	1	1
33	1	1	1	1	1	1	1	1	1	1	1	1
34	1	1	1	1	1	1	1	1	1	1	1	1
35	1	1	1	1	1	1	1	1	1	1	1	1
36	1	1	1	1	1	1	1	1	1	1	1	1
37	1	1	1	1	1	1	1	1	1	1	1	1
38	1	1	1	1	1	1	1	1	1	1	1	1
39	1	1	1	1	1	1	1	1	1	1	1	1
40	1	1	1	1	1	1	1	1	1	1	1	1
41	1	1	1	1	1	1	1	1	1	1	1	1
42	1	1	1	1	1	1	1	1	1	1	1	1
43	1	1	1	1	1	1	1	1	1	1	1	1
44	1	1	1	1	1	1	1	1	1	1	1	1
45	1	1	1	1	1	1	1	1	1	1	1	1
46	1	1	1	1	1	1	1	1	1	1	1	1
47	1	1	1	1	1	1	1	1	1	1	1	1
48	1	1	1	1	1	1	1	1	1	1	1	1
49	1	1	1	1	1	1	1	1	1	1	1	1
50	1	1	1	1	1	1	1	1	1	1	1	1
51	1	1	1	1	1	1	1	1	1	1	1	1
52	1	1	1	1	1	1	1	1	1	1	1	1
53	1	1	1	1	1	1	1	1	1	1	1	1
54	1	1	1	1	1	1	1	1	1	1	1	1
55	1	1	1	1	1	1	1	1	1	1	1	1
56	1	1	1	1	1	1	1	1	1	1	1	1
57	1	1	1	1	1	1	1	1	1	1	1	1
58	1	1	1	1	1	1	1	1	1	1	1	1
59	1	1	1	1	1	1	1	1	1	1	1	1
60	1	1	1	1	1	1	1	1	1	1	1	1
61	1	1	1	1	1	1	1	1	1	1	1	1
62	1	1	1	1	1	1	1	1	1	1	1	1
63	1	1	1	1	1	1	1	1	1	1	1	1
64	1	1	1	1	1	1	1	1	1	1	1	1
65	1	1	1	1	1	1	1	1	1	1	1	1
66	1	1	1	1	1	1	1	1	1	1	1	1
67	1	1	1	1	1	1	1	1	1	1	1	1
68	1	1	1	1	1	1	1	1	1	1	1	1
69	1	1	1	1	1	1	1	1	1	1	1	1
70	1	1	1	1	1	1	1	1	1	1	1	1
71	1	1	1	1	1	1	1	1	1	1	1	1
72	1	1	1	1	1	1	1	1	1	1	1	1
73	1	1	1	1	1	1	1	1	1	1	1	1
74	1	1	1	1	1	1	1	1	1	1	1	1
75	1	1	1	1	1	1	1	1	1	1	1	1
76	1	1	1	1	1	1	1	1	1	1	1	1
77	1	1	1	1	1	1	1	1	1	1	1	1
78	1	1	1	1	1	1	1	1	1	1	1	1
79	1	1	1	1	1	1	1	1	1	1	1	1
80	1	1	1	1	1	1	1	1	1	1	1	1
81	1	1	1	1	1	1	1	1	1	1	1	1
82	1	1	1	1	1	1	1	1	1	1	1	1
83	1	1	1	1	1	1	1	1	1	1	1	1
84	1	1	1	1	1	1	1	1	1	1	1	1
85	1	1	1	1	1	1	1	1	1	1	1	1
86	1	1	1	1	1	1	1	1	1	1	1	1
87	1	1	1	1	1	1	1	1	1	1	1	1
88	1	1	1	1	1	1	1	1	1	1	1	1
89	1	1	1	1	1	1	1	1	1	1	1	1
90	1	1	1	1	1	1	1	1	1	1	1	1
91	1	1	1	1	1	1	1	1	1	1	1	1
92	1	1	1	1	1	1	1	1	1	1	1	1
93	1	1	1	1	1	1	1	1	1	1	1	1
94	1	1	1	1	1	1	1	1	1	1	1	1
95	1	1	1	1	1	1	1	1	1	1	1	1
96	1	1	1	1	1	1	1	1	1	1	1	1
97	1	1	1	1	1	1	1	1	1	1	1	1
98	1	1	1	1	1	1	1	1	1	1	1	1
99	1	1	1	1	1	1	1	1	1	1	1	1
100	1	1	1	1	1	1	1	1	1	1	1	1

Appendix I Site 27-4 Macrofossil Data

	A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S	T	U	V
1	Depth	Mineral	Diatoms	Heterococci/Herb rest	Algae	Forams	Sp. warts	Sp. sub	Sp. series	Dead h	Oryzopsis	Pteridium	Callery	Set	Sp. h	Sp. ang	Sp. m	Sp. l	Sp. n	Sp. o	Sp. p	Sp. q
2	0-5		31.5														17.4	75.6		10	5.5	12.5
3	10-15		5														1.5	75.5				
4	20-25		5														45.4	45.5				
5	30-35		5														65.7	7.3				
6	40-45		3.5														52.8	17.6				
7	50-55		6														40	28				
8	60-65		10														6.2	55.3				
9	67-72		22.5														5.8	30.8				
10	75-80		30														2.5	12.5				
11	87-92		38.5																			
12	95-100		26																			
13	104-108		30																			

Appendix I
Site 27-4 Macrofossil Data Continued

	W	X	Y	Z	AA	AB	AC	AD	AE	AF	AG	AH	AI	AJ	AK	AL
	Unpolluted	Pods recd	Charcoal	Shells and Salt br	Flat Gray	Orange sand	Eriophorum	Carve seed	Carve seed	Carve post	Beak br	Unknown	Cladocore	Ostracoda	Submerged	Dry wt.
1																2
2	12.5	1	1.5						1							1.52
3																3.3
4																1.48
5																4
6																1.48
7																3.3
8																1.48
9																4
10																1.48
11																3.3
12																1.48
13																4
14																1.48
15																3.3
16																1.48
17																4
18																1.48
19																3.3
20																1.48
21																4
22																1.48
23																3.3
24																1.48
25																4
26																1.48
27																3.3
28																1.48
29																4
30																1.48
31																3.3
32																1.48
33																4
34																1.48
35																3.3
36																1.48
37																4
38																1.48
39																3.3
40																1.48
41																4
42																1.48
43																3.3
44																1.48
45																4
46																1.48
47																3.3
48																1.48
49																4
50																1.48
51																3.3
52																1.48
53																4
54																1.48
55																3.3
56																1.48
57																4
58																1.48
59																3.3
60																1.48
61																4
62																1.48
63																3.3
64																1.48
65																4
66																1.48
67																3.3
68																1.48
69																4
70																1.48
71																3.3
72																1.48
73																4
74																1.48
75																3.3
76																1.48
77																4
78																1.48
79																3.3
80																1.48
81																4
82																1.48
83																3.3
84																1.48
85																4
86																1.48
87																3.3
88																1.48
89																4
90																1.48
91																3.3
92																1.48
93																4
94																1.48
95																3.3
96																1.48
97																4
98																1.48
99																3.3
100																1.48

Appendix I Site 35-2 and Site 35-3 Macrofossil Data

	A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S	T	U	V
1	Site 35-2																					
2	Depth																					
3	Soil																					
4	0-5																					
5	10-15																					
6	20-25																					
7	30-35																					
8	40-45																					
9	50-55																					
10	60-65																					
11	70-75																					
12	80-85																					
13	90-95																					
14	100-105																					
15	110-115																					
16	120-125																					
17	130-135																					
18	140-145																					
19	150-155																					
20	Site 35-3																					
21	Depth																					
22	Soil																					
23	0-5																					
24	10-15																					
25	20-25																					
26	30-35																					
27	40-45																					

Appendix I

[illegible]

Appendix I

Appendix I
Site 48-3 Macrofossil Data Continued

	W	X	Y	Z	AA	AB	AC
1	Calcareous Shells		Prorocentrum	Cribroretella	Ammonia	Charcoal	Subsammol
2							3
3							4
4							4
5							4.2
6							4.2
7							4.2
8							3.7
9							3.1
10							3.7
11							3.7

APPENDIX II

Weighted Means for pH and Height Above Water Table of 1331 Relèves in Western Canada

Species	Mean pH Corrected for Skewness	Mean Height Above Water Table
<i>Sphagnum russowii</i>	4.24	2.8
<i>Sphagnum riparium</i>	4.26	0.5
<i>Sphagnum magellanicum</i>	4.33	2.1
<i>Sphagnum capillifolium</i>	4.37	3.9
<i>Polytrichum strictum</i>	4.58	2.4
<i>Sphagnum angustifolium</i>	4.62	1.7
<i>Pohlia nutans</i>	4.7	1.3
<i>Pleurozium schreberi</i>	4.73	2.8
<i>Pohlia sphagnicola</i>	4.8	2.5
<i>Drepanocladus exannulatus</i>	4.92	0.4
<i>Sphagnum fuscum</i>	4.97	3
<i>Sphagnum teres</i>	4.98	1.5
<i>Dicranum undulatum</i>	5.03	2.3
<i>Sphagnum subsecundum</i>	5.25	1
<i>Sphagnum squarrosum</i>	5.32	1.2
<i>Calliergon richardsonii</i>	5.81	0.7
<i>Calliergon stramineum</i>	5.69	0.7
<i>Aulacomnium palustre</i>	5.7	1.9
<i>Calliergonella cuspidata</i>	5.92	0.5
<i>Meesia longiseta</i>	5.96	- -
<i>Drepanocladus lapponicus</i>	6.05	0.8
<i>Drepanocladus aduncus</i>	6.06	1
<i>Sphagnum warnstorffii</i>	6.11	1.7
<i>Calliergon giganteum</i>	6.2	1.1
<i>Tomenthypnum nitens</i>	6.2	1.5
<i>Bryum pseudotriquetrum</i>	6.22	1
<i>Calliergon trifarium</i>	6.39	0
<i>Drepanocladus vernicosus</i>	6.4	1.1
<i>Scorpidium scoripoides</i>	6.62	-0.1
<i>Meesia triquetra</i>	6.66	0.6
<i>Campylium stellatum</i>	6.67	1.1
<i>Drepanocladus revolvens</i>	6.76	0.8
<i>Brachythecium mildeanum</i>	6.79	1.5
<i>Drepanocladus polycarpus</i>	6.86	1.9

Appendix III

160

Appendix III Peat Chemistry - Elk Island National Park Continued

Site	Depth (cm)	Peat	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S
			PEAT CHEM	H ₂ O	Ca	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg
83	185-190		187	8877.6	161.7	1732.7	8.4131	8.3485	8.3735	5.4541	11.204	374.17	11.204	315.44	315.44	315.44	315.44	315.44	315.44
84	195-200		188	8728.8	177.2	164.36	8.2661	10.889	10.385	2.8289	2215.6	180.89	180.89	180.89	180.89	180.89	180.89	180.89	180.89
85	205-210		189	8123.7	1813.9	164.89	781.12	10.897	9.9566	10.408	4.1797	7914.9	194.77	260.06	13.319	194.77	260.06	13.319	194.77
86	215-220		190	8116.6	1833.2	147.38	884.22	6.7536	9.8988	11.782	4.96	9891.1	231.28	193.44	12.318	789.48	3711.9	2	2
87	225-230		191	8543.0	1878.8	145.9	910.58	6.884	7.4428	12.887	4.8293	11791.8	289.81	198.43	12.903	865	3870.4	2	2
88	235-240		192	10402.6	1728.1	154.6	900.76	4513	10.939	7.184	11.06	6.8219	14212.9	318.08	208.34	14.193	1088.9	6061.7	2
89	245-250		193	10419.2	2127.1	157.97	1088.1	6673.7	11.767	13.309	14.857	3.8482	17892.1	482.48	174.69	18.468	2388.4	5266.2	2
90	255-260		194	11805.8	2034.3	172.59	838.49	6711.4	10.762	13.276	14.858	6.4278	19004.6	496.21	165.81	18.421	2259.6	7015.2	2
91	265-270		195	11742.9	2017.2	240.92	857.23	8448.9	15.152	10.808	14.533	4.3748	20385.9	504.35	138.98	16.594	1814.7	8872.4	2
92	275-280		196	12176.8	2034.6	185.05	958.77	6758.1	17.461	10.881	13.648	4.3744	20218.9	531.07	138.91	16.528	2004.3	8480.6	2
93	285-290		197	12176.8	2034.6	185.05	958.77	6758.1	17.461	10.881	13.648	4.3744	20218.9	531.07	138.91	16.528	2004.3	8480.6	2
94	295-300		198	12176.8	2034.6	185.05	958.77	6758.1	17.461	10.881	13.648	4.3744	20218.9	531.07	138.91	16.528	2004.3	8480.6	2
95	305-310		199	13002.4	2099.1	140.38	1233.4	6484.9	17.604	14.474	16.134	4.8088	20180.9	574.19	112.92	17.321	3064.2	8645.7	2
96	315-320		200	13780.3	2099.1	140.38	1233.4	6484.9	17.604	14.474	16.134	4.8088	20180.9	574.19	112.92	17.321	3064.2	8645.7	2
97	325-330		201	17116.3	2556.9	188.38	1649.6	8910.6	16.682	14.765	17.878	7.8008	18482.4	483.42	101.36	18.137	885.1	9850.6	2
98	335-340		202	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
99	345-350		203	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
100	355-360		204	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
101	365-370		205	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
102	375-380		206	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
103	385-390		207	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
104	395-400		208	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
105	405-410		209	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
106	415-420		210	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
107	425-430		211	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
108	435-440		212	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
109	445-450		213	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
110	455-460		214	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
111	465-470		215	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
112	475-480		216	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
113	485-490		217	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
114	495-500		218	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
115	505-510		219	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
116	515-520		220	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
117	525-530		221	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
118	535-540		222	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
119	545-550		223	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
120	555-560		224	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
121	565-570		225	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
122	575-580		226	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
123	585-590		227	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
124	595-600		228	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
125	605-610		229	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
126	615-620		230	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
127	625-630		231	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
128	635-640		232	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
129	645-650		233	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
130	655-660		234	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
131	665-670		235	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
132	675-680		236	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
133	685-690		237	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
134	695-700		238	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
135	705-710		239	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
136	715-720		240	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
137	725-730		241	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
138	735-740		242	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
139	745-750		243	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.42	11199.6	288.69	181.17	14.848	888.37	4287.3	2
140	755-760		244	10321.9	2115.3	184.88	1574.3	6875.6	9.9289	14.765	13.638	5.4							

Appendix III Peat Chemistry - Elk Island National Park Continued

A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S
DATE	DEPTH (cm)	PEAT CHEMICALS	Ca mg/kg	Mg mg/kg	Na mg/kg	K mg/kg	Al mg/kg	Ti mg/kg	Fe mg/kg	As mg/kg	Cu mg/kg	Zn mg/kg	Mn mg/kg	Co mg/kg	Pb mg/kg	Cr mg/kg	B mg/kg	Feat Type
1993	84-82	E409	8438	2167.8	266.44	1880	9389.8	23.247	11.107	11.488	6.9038	4700.5	46.897	14.718	8.3071	433.82	1588.1	6
1993	100-104	E470	12015.5	1925	152.48	871.71	8741.6	18.008	2.7823	6.4015	4.0816	3726.7	82.497	10.437	8.1248	824.15	2344.1	6
1993	100-108	E471	9188.7	2203.3	248.2	1060.1	8541.7	23.076	6.867	11.591	5.864	6720.8	82.497	48.756	13.153	444.16	1891.2	6
1993	100-145	E472	8435.2	2004.8	238.85	1335.5	8411.3	24.333	6.0055	6.6479	7.1771	6339.8	106.4	33.204	14.825	483.04	1737.2	2
1993	100-169	E473	8435.9	2081.7	235.81	1810.7	8123.8	24.246	6.9718	10.746	7.1771	6079.3	116.4	83.68	11.727	489.68	2149.2	2
1993	100-169	E474	8908.1	2138.9	231.27	1872.6	8284.3	25.092	12.538	11.73	6.1071	6820.9	124.9	86.71	14.836	823.1	2655.4	2
1993	100-169	E475	6653	1725.3	214.92	1444.9	8493.1	21.565	8.2537	8.8224	4.8538	8397.9	102.57	63.875	12.317	388.73	2407.7	2
1993	200-208	E741	6018.3	1999.6	231.84	2714.5	8439.1	30.887	8.4183	7.3459	4.1803	6107.2	101.1	47.893	10.776	398.77	2128.7	2
1993	220-225	E476	10575.8	2754.1	328.87	2316	10281.1	30.887	11.988	18.498	7.5118	8432.8	176.55	69.85	18.893	510.64	4054.4	2
1993	241-244	E477	8580.3	2893.4	355.13	2892.3	11780.3	13.195	16.295	12.398	6.0495	8693.1	167.11	85.428	18.893	432.48	4888.1	1
1993	261-264	E478	8588.5	2453.9	354.33	2899.1	11083.7	13.195	17.953	17.953	7.1771	12880.5	244.7	62.898	10.385	465.77	7887.2	1
1993	280-285	E479	9311	2821.9	355.76	2535.6	11114.3	38.563	12.826	16.95	6.8284	11088.8	192.04	65.487	10.532	465.77	7887.2	1
1993	300-305	E480	13888.2	3140.7	359.86	2491.7	11114.3	24.974	12.826	16.95	6.8284	11088.8	192.04	65.487	10.532	465.77	7887.2	1
1993	320-325	E481	10514	2818.3	309.86	2395.1	8884.4	17.25	18.825	18.825	7.404	10519.4	214.08	98.417	10.138	478.08	8905.8	1
1993	337-342	E482	12053.1	4459.3	399.55	2732.4	11823.9	39.972	13.647	15.358	7.404	10519.4	214.08	98.417	10.138	478.08	8905.8	1
1993	355-359	E743	18488.3	3105	317.48	2404.6	8423.9	39.972	13.647	15.358	7.404	10519.4	214.08	98.417	10.138	478.08	8905.8	1
1993	377-382	E744	8780.5	2108.2	302.82	2395.5	8779.5	31.638	13.109	14.054	6.9335	12857.3	280.85	59.733	15.335	418.2	4822.7	1
1993	395-400	E745	8280.1	1711.8	229.91	1920.7	7514.1	36.741	14.103	11.63	6.1555	8491.1	143.1	38.107	12.373	331.4	4822.7	1
1993	417-422	E483	10546.6	3130.8	413.14	2824.8	10858	30.888	17.742	18.752	7.0904	8488.2	112.03	34.916	0.512	248.06	3394.3	1
1993	437-442	E484	10822.6	2710.8	354.91	2481.5	10844.9	38.27	13.853	15.097	6.4764	7881.3	181.89	85.039	22.886	467.24	6753.3	1
1993	457-462	E485	9364.6	2836.9	371.72	2738	10548.5	39.487	16.949	15.977	6.0574	8788.3	148.01	58.245	18.225	494.6	6781.5	1
1993	467-472	E486	7670.6	2886.7	487.84			25.228	18.95		8.11	9001.8	134.7	85.285	21.136	387.27	4676.7	1
1993	487-492	E487																
1993	505	E746	4189.9	1025.8	173.58	1213.8	1204.1	14.828	7.3044	1.28	3.8588	971.46	83.517	39.005	4.4674	1289.5	1345.3	6
1993	520-525	E747	8248.3	11120	123.72	1168.6	837.36	14.46	15.749	4.3883	3.2373	1027.7	40.882	37.101	8.6431	1419.6	1593.1	6
1993	536-540	E487	5900.6	1251.5	77.108	491.81	1587.1	10.813	2.8986	3.3219	0.8591	1457.4	25.618	20.43	1.8548	931.18	1899.3	2
1993	550-555	E488	8038.7	1791.3	140.18	1202.6	4824.4	15.848	6.82	8.8199	3.1026	4062.6	207.7	92.142	10.717	912.23	4555.6	2
1993	575-580	E748	10847.6	2048.3	137.72	1302.6	5828.6	20.782	18.027	14.543	5.2052	9007.9	316.38	99.538	17.207	804.99	6691.4	2
1993	595-600	E749	11376.3	2083	140.74	1362.1	6088.6	24.331	17.037	13.591	5.8135	10808.4	384.82	97.84	186.47	840.44	7442.7	2
1993	615-620	E750	12897.5	2169.1	154.06	1847.8	8472.3	12.327	18.831	19.042	4.2138	13405.6	432.27	93.822	19.361	885.32	9318.5	2
1993	635-640	E751	12085	2243.6	154.89	1893.9	7183.7	12.327	18.798	18.835	4.3916	13998.4	438.66	73.67	31.08	853.75	8468	2
1993	655-660	E752	10640	2898.6	224.85	2882.5	10195.5	10.811	27.735	20.459	5.8423	14255.2	379.05	69.248	23.015	878.17	7154.9	2
1993	675-680	E753	11631.6	2488.4	224.74	1974	8519.3	18.486	21.411	20.472	5.8208	18974.9	381.95	77.589	22.976	700.1	8109.9	2
1993	695-700	E754	13493.5	2822.9	239.62	2161.9	9227.7	31.827	25.444	23.878	4.335	20910.8	489.32	78.288	23.983	832.41	1845.1	2
1993	715-720	E755	17868.2	315.7	278.36	2214.3	8461.3	31.827	25.444	23.878	4.335	20910.8	489.32	78.288	23.983	832.41	1845.1	2
1993	735-740	E756	33168.6	3419.2	278.6	1842.7	9237.9	21.42	24.612	25.08	6.8129	24898	611.1	102.32	26.95	908.94	18002.6	2
1993	755-760	E757	47870	3737.5	215.06	1855.4	8466.7	21.42	24.612	25.08	6.8129	24898	611.1	102.32	26.95	908.94	18002.6	2
1993	775-780	E758	280.44	2103.9	210.94	2228.9	11584.6	16.441	24.945	18.486	6.0084	17085.2	440.48	69.991	21.488	588.88	14300.3	2
1993	795-800	E759	21411.3	3851.7	280.44	2103.9	10848.2	25.825	26.86	24.058	6.1555	18828.6	481.8	78.842	24.724	788.75	18005.3	2
1993	815-820	E760	18091.7	3288.3	310.84	2228.9	11584.6	16.441	24.945	18.486	6.0084	17085.2	440.48	69.991	21.488	588.88	14300.3	2
1993	835-840	E761	37558.4	3934.6	282.24	1338.5	9578.7	34.472	24.181	24.736	6.8222	21687.2	693.48	78.908	24.913	845.48	18865.9	2
1993	845-850	E762	99331.8	4580	315.7	1112.2	11023.9	35.782	26.041	26.041	8.043	18672	932.4	75.843	25.316	899.94	17477.3	2
1993	865-870	E763	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	885-890	E764	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	905-910	E765	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	915-920	E766	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	925-930	E767	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	935-940	E768	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	945-950	E769	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	955-960	E770	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	965-970	E771	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	975-980	E772	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	985-990	E773	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2
1993	995-1000	E774	81284.8	4771.3	288.05	1514.2	10080.1	35.088	23.017	25.088	4.18	13116.4	1044.5	76.148	22.728	1147.4	18802.4	2

Appendix III

163

Appendix III

A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S
DATE	DEPTH (m)	PEAT CHEMICAL	Ca mg/kg	Mg mg/kg	Na mg/kg	K mg/kg	Al mg/kg	Ti mg/kg	Pb mg/kg	As mg/kg	Cu mg/kg	Fe mg/kg	Mn mg/kg	Zn mg/kg	Hg mg/kg	P mg/kg	S mg/kg	Peat Type
182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.785	6.4801	16391	181.89	80.442	21.366	682.09	918.9	1
182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.785	6.4801	16391	181.89	80.442	21.366	682.09	918.9	1
182-8-168	182.8	E153	17385.6	1409.6	200.78	1750.9	7748.1	17.438	18.215	11.118	6.822	9802.9	180.55	76.864	16.888	810.18	12945.8	1
201-8-204-4	201.8	E149	17385.6	1409.6	200.78	1750.9	7748.1	17.438	18.215	11.118	6.822	9802.9	180.55	76.864	16.888	810.18	12945.8	1
182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.785	6.4801	16391	181.89	80.442	21.366	682.09	918.9	1
182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.785	6.4801	16391	181.89	80.442	21.366	682.09	918.9	1
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182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.785	6.4801	16391	181.89	80.442	21.366	682.09	918.9	1
182-8-167-8	182.8	E048	12272.8	2341.3	346.22	3458.8	13384	16.283	14.592	16.78								

Appendix III

165

Appendix III Peat Chemistry - Elk Island National Park Continued

A	B	C	D	E	F	G	H	I	J	K	L	M	N	O	P	Q	R	S	
SITE	DEPTH (cm)	PEAT CHEMICALS	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg	mg/kg
307 SITE 36.3	0-5	E781	14828.9	2420.3	373.38	2018.3	8776.1	8.1882	15.073	10.865	10.305	7092	128.11	143.74	<1.08	1288.4	4998.1	3	
308			18288.9	2527.7	199.1	979.98	8575	8.6444	10.865	11.552	10.305	7092	132.41	42.821	<0.863	971.66	6891.3	1	
309			18287.8	2591	320.91	1800.9	13791.5	31.298	15.997	15.952	17.326	8259.2	27.845	197.8	<0.8328	924.05	2431.2	1	
310	21-26	E783	8639.6	1869.6	231.59	1706	18588.1	10.304	15.759	14.226	8.7866	502.5	23.148	82.953	<1.04	316.84	453.99	1	
311	20-33	E784	29911.9	2417.6	201.35	132.29	1215.3	16.517	<2.18	3.8448	12.992	2225.2	4.3748	82.828	<1.07	828.76	12478.1	1	
312	40-45	E785	2885.6	2384.7	300.45	2338.7	12719.4	9.9979	16.575	13.738	2.4456	8888.8	43.591	37.83	<1.03	312.63	334.84	1	
313	48-50	E786																	
314			43581.2	2966.8	3139.6	680.72	2633.9	34.928	<2.28	6.7596	6.3889	3394	278.82	58.145	10.793	1884.8	8913.4	3	
315	0-5	E532	40482.3	3021.6	2519.8	612.85	2977.8	35.348	<2.28	6.6194	6.3878	4146.8	178.87	64.144	6.3187	1108.4	11775.3	3	
316	10-20	E533	39027.8	3019.9	3271.9	685.45	3338.6	39.303	<2.28	6.2045	6.7523	3885.6	76.534	25.479	7.7847	1487.8	13455.4	5	
317	20-28	E534	36882.7	2468	3308.7	304	1289.7	22.180	<2.28	3.4379	6.402	3478.9	61.927	16.157	5.1095	1169.2	14761.4	3	
318	30-38	E535	37202.8	2863.5	3316.8	491.51	2241.3	33.887	<2.28	4.2608	4.1953	3327.3	87.83	41.874	7.828	1102.4	14762.3	3	
319	40-48	E536	32382.7	2518.2	2588.8	348.22	1811.4	28.41	<2.28	3.9404	5.0188	3322.1	82.214	23.424	6.8737	907.62	12660.6	3	
320	60-65	E537	31681.6	2570.9	2727.3	680.84	4815.5	49.951	<2.28	8.6177	6.4141	4986.1	51.053	23.454	11.922	891.37	16486.7	3	
321	70-78	E538	31688.7	231.3	2605.9	482.81	3207.3	40.907	<2.28	8.0411	8.8451	5048.5	53.53	25.35	10.897	787.06	16873.7	3	
322	90-95	E539	32783.8	2380.3	2351.5	776.81	4895	43.187	<2.28	9.0883	9.7688	4325.9	75.248	37.027	11.816	716.79	15004.8	3	
323	110-115	E540	68955.7	3080.5	1887.1	1008.5	6346	63.159	<2.28	12.843	8.8455	4325.9	189.66	38.937	12.131	215.03	15809.8	3	
324	130-135	E541	68972.6	4411.4	1684.3	1055.7	6333.7	67.577	<2.28	15.794	10.415	7309.9	432.76	51.888	14.1	277.88	14847.5	1	
325	150-155	E542	68903.8	6885.1	1512.2	1568.6	7951.8	69.162	3.5717	17.434	13.402	9808.8	860.36	89.749	154.8	728.95	13373.1	1	
326	170-175	E543	68983.6	6499.1	1541.9	1483.3	7953.8	43.444	<2.27	17.082	12.804	9050.2	602.76	57.135	164.8	739.97	13175.1	1	
327	190-195	E544	69180.8	8852.7	1628.6	2288.8	10285.2	27.377	6.3787	23.13	14.028	11218.7	831.02	62.495	19.03	739.97	13175.1	1	
328	210-215	E545	69248.8	7760.7	1328.6	1651.4	7848.4	30.385	4.0579	21.785	13.918	8198.5	850.11	38.7	14.738	418.07	10439.1	1	
329	234-239	E546																1	
330	240-245	E547																1	
331			26521.8	2881.1	1888.8	1070.3	2418.6	26.371	8.9756	4.3027	8.0303	2328.4	165.76	87.11	<1.16	1601.5	6845.8	1	
332	0-5	E787	37080.8	2901.8	2459.1	317.84	1822.2	28.053	21.87	9.7476	8.2404	2089.3	154.45	24.982	8.4772	1209.7	10900	1	
333	20-25	E788	40514.8	3113	2957.4	241.08	1788.8	25.223	<2.40	3.3119	5.9192	1501.1	66.537	34.788	<1.18	1186.4	12283.2	2	
334	30-35	E789	40129.9	2743.7	2978.6	188.03	1548.2	31.516	<2.39	3.3722	6.1525	3313.9	42.547	14.518	<1.17	829.85	15535	2	
335	40-45	E790	30248	2208.3	2887.4	251.08	1247.9	34.147	<1.84	4.6477	7.4782	4265.2	31.65	8.2128	<0.8985	903.92	14070	1	
336	60-65	E791	30211.1	2128.1	2983.1	171.85	1578.7	28.989	<2.09	4.4471	7.4782	4502.3	25.388	11.778	<1.02	916.35	15455.8	1	
337	70-75	E792	31443.1	2100.2	2560.8	392.83	3825.6	44.41	2.7489	6.4261	8.0613	5388.5	33.753	12.226	21.005	873.48	14515.9	1	
338	90-95	E793	32100.7	2077.8	2709	514.37	4953.1	44.48	22.457	12.392	11.738	7454.2	51.897	10.854	15.459	888.32	18173.9	1	
339	110-115	E794	41132.6	2300.4	3115.5	311.22	2381.7	35.618	<2.28	6.5042	11.368	8774.7	69.056	29.254	11.989	740.7	18118	1	
340	130-135	E795	60886.8	2048.8	2287.3	614.08	4824.3	38.291	6.3813	9.1655	13.201	7989.8	137.13	32.995	11.747	873.94	26000.2	2	
341	150-155	E796	64826.1	2540.3	2337.2	728.45	7226.7	50.149	2.8459	10.868	14.644	12123.9	865.86	108.71	191.82	19.183	798.45	14528.2	2
342	170-175	E797	69913.1	6479.9	1514.3	1483.2	7551.6	51.273	11.757	19.128	14.644	12123.9	865.86	108.71	191.82	19.183	798.45	14528.2	2
343	185-190	E798	69901.7	6083.7	1576	2170.7	7251.2	25.048	14.081	24.491	14.221	15493.2	1088.4	60.078	18.399	825.32	14559.9	2	
344	205-210	E799	151125.3	6341.5	622.74	1516.3	7329.3	43.308	41.418	30.184	12.411	9127.7	985.97	28.718	23.579	349.8	8161.9	1	
345	210-215	E800	68225	5887.8	892.81	1537.6	6248.1	19.349	6.7641	14.434	10.17	8264.4	927.48	27.047	14.215	331.53	6156.5	1	
346																			
347																			
348	0-5	E554	89147.1	7507.8	1213.1	3687.5	14182.1	33.285	18.282	22.842	12.292	11331.7	1013.8	65.668	21.862	534.79	5896.3	4	
349	10-20	E555	39286.7	2207.7	2950.8	253.95	1759.8	27.512	<2.27	<1.65	8.8621	1981.9	51.343	80.776	6.2865	1224.1	12542.5	4	
350	30-35	E556	34760.9	2019.6	2883.8	223.12	2149.4	39.225	<2.27	<1.88	13.226	3124.2	59.896	88.893	7.5035	899.03	15849.5	4	
351	40-45	E557	36772.3	2023.2	2883.7	223.44	2185.7	29.205	<2.27	2.8745	18.137	4593.4	65.945	130.86	9.491	883.13	15411.5	3	
352	60-65	E558	38879.3	1929.8	2379.4	257.34	2904.8	40.181	<2.27	4.8928	30.785	4204.9	81.825	253.58	10.253	1242.7	1713.1	3	
353	80-85	E559	63787.8	3324.3	2378.6	1180.5	6695.4	44.819	12.395	4.8928	30.785	4204.9	288.84	63.248	15.813	772.17	16689.9	3	
354	100-105	E560	69101.4	5295.6	848.12	1870.9	1289.3	81.953	18.344	16.115	17.78	13233.3	490.05	95.957	24.488	721.6	2372.5	1	
355	120-125	E561	115769.8	7084.8	954.2	2861.3	16269.3	128.42	28.744	27.006	12.081	18185.5	862.78	65.505	18.181	492.86	1171.5	1	
356	140-145	E562																	
357	160-165	E563																	
358	180-185	E564																	
359	200-205	E565																	

350 Peat Type = catagories 1-8 identified from peat macrofossils. See description analysis section.
357/58 = A sample which was redone